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Indoor particulate matter during HOMEChem: Concentrations, size distributions, and exposures

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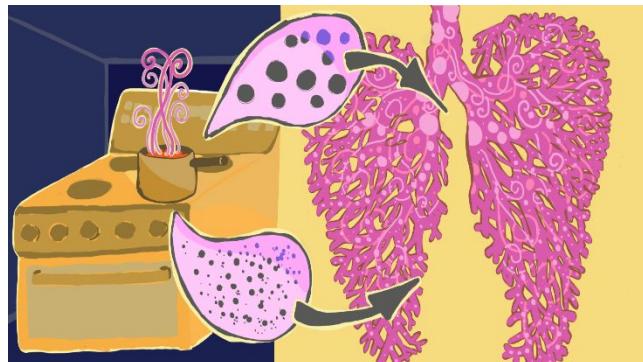
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40 Abstract

41 It is important to improve our understanding of exposure to particulate matter (PM) in residences
42 because of associated health risks. The HOMEChem campaign was conducted to investigate
43 indoor chemistry in a manufactured test house during prescribed everyday activities, such as
44 cooking, cleaning, and opening doors and windows. This paper focuses on measured size
45 distributions of PM (0.001-20 μm), along with estimated exposures and respiratory-tract
46 deposition. Number concentrations were highest for sub-10 nm particles during cooking using a
47 propane-fuelled stovetop. During some cooking activities, calculated PM_{2.5} mass concentrations
48 (assuming a density of 1 g cm^{-3}) exceeded 250 $\mu\text{g m}^{-3}$ and exposure during the post-cooking decay
49 phase exceeded that of the cooking period itself. The modeled PM respiratory deposition for an
50 adult residing in the test house kitchen for 12 hours varied from 7 μg on a day with no indoor
51 activities, to 68 μg during a simulated day (including breakfast, lunch, and dinner preparation
52 interspersed by cleaning activities), and rose to 149 μg and during a simulated Thanksgiving day.

53 Abstract Art



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57

58 1. Introduction

59 Outdoor air quality in much of the developed world has improved substantially over the past half
60 century, corresponding to expected improvements in pollution-associated health risks.^{1,2} On the
61 other hand, indoor air quality (IAQ) is neither well-regulated nor well understood. This knowledge
62 gap is critical because surveys indicate that people spend ~90% of their time indoors³ and most of
63 that time is spent in one's residence. Pollutants of outdoor origin are an important determinant of
64 IAQ.^{4,5} However, with decreasing ambient pollutant levels, the role of pollutants of indoor origin
65 is becoming a relatively more important influence for personal exposure. Beyond intrusion of
66 pollutants from outdoors, indoor emission sources can degrade IAQ. Activities that contribute to
67 indoor air pollution include cooking,^{6,7} smoking,^{8–10} and cleaning,^{11,12} in addition to emissions
68 from indoor constituents like building materials,^{13,14} personal care products,¹⁵ consumer
69 electronics,¹⁶ and even human occupants themselves.¹⁷ The physical and chemical phenomena that
70 control pollutant transport and transformations outdoors differ from those which are important
71 indoors, due to factors such as shorter air residence times indoors, altered abundances of oxidative
72 species, and higher surface-to-volume ratios in indoor environments compared to urban and
73 regional atmospheres.¹⁸

74 Particulate matter (PM) is a key pollutant from health and environmental perspectives both indoors
75 and outdoors. Exposure to PM is associated with various adverse health outcomes, such as
76 cardiovascular disease, chronic obstructive pulmonary disease, and asthma.^{19–21} It is one of the
77 leading global causes of mortality and ill-health.²² A study quantifying the relationship between
78 global mortality and ambient PM_{2.5} estimated that ~2 million premature deaths can be avoided by
79 reducing PM_{2.5} to 10 µg m⁻³ globally.²³

80 Bekö et al. measured ultrafine particle exposure of sixty nonsmoking residents of Copenhagen and
81 found that their homes and other built environments accounted, respectively, for 50% and ~40%
82 of their daily personal exposure.²⁴ A recently developed framework attributed $42 \pm 24\%$ and $28 \pm$
83 26% of the total exposure (including different microenvironments and outdoors) to PM_{2.5} of
84 outdoor and indoor origin in residences, respectively.²⁵ Among all indoor microenvironments
85 considered, residences contributed the most to exposure and associated mortality burden. The
86 considerable uncertainty associated with these estimates highlights our limited understanding of
87 PM in residences.

88 Many studies have reported on the characteristics of PM in residences. Sources considered in such
89 studies include cooking,²⁶ other combustion sources such as candles and incense,^{27,28} activities
90 involving hot surfaces such as irons and hair dryers,^{29,30} and penetration from outdoors.⁴ Studies
91 with multiple activities and sources in a house-like controlled environment are sparse.

92 The House Observations of Microbial and Environmental Chemistry (HOMEChem) campaign
93 was conducted in a test house in June 2018 to investigate how everyday activities impact the
94 emissions, chemical transformations, and removal of trace gases and particles in a residential
95 environment.³¹ Performing prescribed activities such as cooking and cleaning in the test house
96 enabled the simulation of residential conditions that were more controlled than observational field
97 campaigns but less controlled than chamber studies. This work focuses on insights from PM size
98 distributions (0.001-20 μm) measured using a range of particle sizing instruments. Overall size-
99 segregated PM number and calculated mass concentrations and their variation throughout different
100 indoor activities are discussed. Sources of PM in different size modes and estimated exposure and
101 lung deposition are also reported.

102 **2. Materials and Methods**

103 *2.1 Test House*

104 The HOMEChem study was conducted in a three-bedroom, two-bathroom manufactured test
105 house (111 m^2 floor area and $\sim 250 \text{ m}^3$ volume), located at the University of Texas at Austin. The
106 house layout and instrument locations are presented in Fig. S1 of the supporting information (SI).
107 The test house has been described in detail elsewhere;³¹ main characteristics are summarized here.
108 The outdoor air supply system was kept continuously on to maintain a positive pressure while
109 providing an air exchange rate of $0.5 \pm 0.1 \text{ h}^{-1}$. The air conditioning system turned on and off
110 intermittently to maintain the target temperature ($\sim 25^\circ\text{C}$, under thermostatic control). The fan in
111 the air handling unit operated continuously, without a filter, at a flow rate of $2000 \text{ m}^3 \text{ h}^{-1}$ to
112 maintain a high rate of internal mixing (equivalent to 8 house volumes h^{-1}). All external doors and
113 windows were kept shut, except when specific experiments required otherwise. All internal doors,
114 except those of the two bathrooms, were kept open throughout. More details about the test house
115 location and its ventilation system are available in the SI.

116 *2.2 Experimental design*

117 The HOMEChem experimental design is comprehensively detailed in Farmer et al.³¹ Two
118 categories of experiments were performed: (1) Sequential, in which the same type of activity was
119 performed repeatedly throughout the day at regular intervals, sometimes interspersed by a venting
120 period, with opened doors and windows; (2) Layered, in which different types of activities were
121 performed throughout the day—without opening the house—to allow emissions to interact.
122 Sequential experiments focused on quantifying the impacts of specific activities; layered
123 experiments simulated a “day in the life” of a residential environment.

124 Three types of sequential experiments, Sequential Stir-Fry, Sequential Cleaning, and Sequential
125 Ventilation, were performed. In Sequential Stir-Fry experiments, a meal consisting of a vegetable
126 stir-fry and white rice was cooked. This process was repeated four times (three on the propane
127 stove and one on an electric hot plate) on each of four days. Similarly, during each day devoted to
128 Sequential Cleaning, the test house was mopped multiple times using a variety of cleaners. The
129 Sequential Ventilation experiment consisted of repeatedly opening and closing all external doors
130 and windows throughout the day to assess its impact on the time-dependent relationship between
131 indoor and outdoor PM levels.

132 Two types of Layered experiments were conducted: Layered Day and Thanksgiving Day. In the
133 four Layered Day experiments, volunteers entered the house in the morning, cooked breakfast
134 (pan-fried sausage, fried eggs, fried tomato, toast, and coffee), mopped the floors with a pine-
135 scented cleaner, prepared lunch (same as in Sequential Stir-Fry), prepared coffee and toast, cooked
136 dinner (either lasagna or chili), mopped the house with bleach, started the dishwasher, and left the
137 house. During the two Thanksgiving Day experiments, four volunteers entered the house in the
138 morning, prepared breakfast (same as in Layered Days), then prepared a typical Thanksgiving
139 meal, including oven-roasted turkey, bread stuffing/dressing, brussels sprouts, and sweet potato
140 casserole, in addition to pies, cranberry sauce, and gravy. Then, 12-15 guests entered the house,
141 dined, performed cleaning activities, started the dishwasher, and exited the house. Example
142 activity schedules followed during these experiments are presented in Tables S1-S5.

143 Two more experiments (naked stove and naked hot plate) were performed to characterize PM
144 generated during the operation of the stove and hot plate, without a pan or pot, at the highest power
145 level for ~20 minutes. These two experiments were not part of either Layered Day or Sequential
146 Day experiments. Test house PM levels were at background levels before the start of experiments.

147 The Institutional Review Board's (IRB) human research review approval requirement was waived
148 for the HOMEChem study because no identifying information was collected from house
149 occupants.

150 *2.3 Instrumentation, sampling scheme, and data analysis*

151 An A11 Nano Condensation Nucleus Counter system (A11-nCNC, Airmodus Oy, Helsinki,
152 Finland) was deployed to sample particles in the 1-4 nm activation size range.³²⁻³⁴ Two Scanning
153 Mobility Particle Sizers (SMPS, TSI Inc., Shoreview, MN) measured number distributions in the
154 4-532 nm electrical mobility size range. Aerodynamic size distributions in the diameter range 0.5-
155 20 μm were measured using Aerodynamic Particle Sizers (APS, TSI 3321). Optical measurements
156 made by an Ultra-High Sensitivity Aerosol Spectrometer (UHSAS, Droplet Measurement
157 Technology, Longmont, CO) were used to compensate for data loss in the 105-533 nm size range
158 from one of the SMPS units on certain days. The A11-nCNC and both SMPSs were set to capture
159 one size distribution every five minutes, whereas the APS and UHSAS recorded a size distribution
160 every minute and every second, respectively. To minimize particle losses and to ensure that the
161 smaller particles were captured, all particle sizers except the UHSAS were placed inside the test
162 house (Figs. S1b and S1c). More details about the instruments are available in the SI.

163 Depending on PM composition, airborne particle density typically varies within the range 0.8-2.5
164 g cm^{-3} .³⁵⁻³⁸ For this study, mass concentrations were calculated from the number size distribution
165 data assuming spherical particles with a density of 1 g cm^{-3} similar to previous studies on indoor
166 PM.^{39,40} The SI presents a discussion on time-resolved PM_1 density estimated using composition
167 data obtained from an aerosol mass spectrometer, density assumptions, and associated
168 uncertainties (Figs. S2 and S3, Tables S6 and S7). Briefly, using a constant density of 1 g cm^{-3}

169 underestimated PM₁ mass concentrations during cooking and post-cooking decay phase by less
170 than 25% compared to PM₁ mass concentrations calculated using composition-dependent density.

171 For quantitative analyses, calculated PM mass concentrations were segregated into six size
172 groupings: PM_{0.1}, PM_{0.5}, PM₁, PM_{2.5}, PM₁₀, and PM₂₀, whereas the number concentrations were
173 divided into two size bins: 4-100 nm (commonly referred to as ultrafine particles, or UFP) and 0.1-
174 20 μm . For Layered Days and Thanksgiving, average concentrations over 12 hours (8 AM - 8 PM
175 local time) are reported. The average mass and number concentrations corresponding to different
176 meals were calculated over the period between the moments in which the heat source is first turned
177 on and finally turned off.

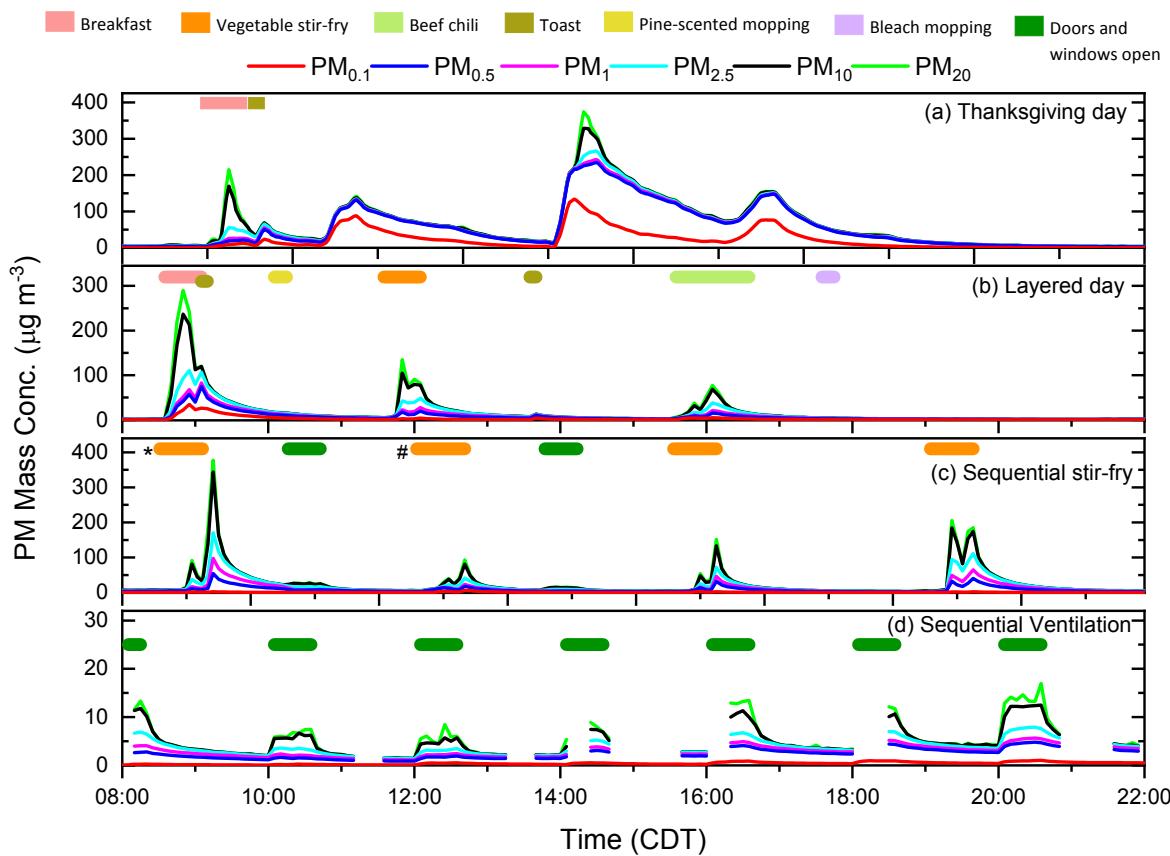
178 Integrated PM_{2.5} exposures for an occupant residing continuously in the test house were calculated
179 over 12 hours for Layered Days and for Thanksgiving. For each cooking activity, exposures during
180 the cooking and post-cooking decay period, are reported. PM mass deposited in the respiratory
181 tract (0.004-20 μm) was estimated using a lung deposition model (Eq. S1-S4) of the International
182 Commission on Radiological Protection^{41,42} for a particle density of 1 g cm^{-3} , utilizing an age-
183 weighted average volumetric inhalation rate for 20-60 year old adults of 11 L min^{-1} .⁴³ Because
184 measurements were performed at a single location near the kitchen, the reported exposures and
185 respiratory deposition estimates are specific to that location.

186 **3. Results and Discussion**

187 *3.1 Overview of particulate matter concentrations and trends*

188 Fig. 1 presents the temporal variation of size-segregated PM mass concentrations during four
189 experimental days. Data corresponding to the other experimental replicates are presented in Fig.
190 S4-S6. Cooking activities were the single largest source of indoor PM on a mass basis per activity.

191 During cooking, PM_{2.5} concentrations as high as 250 $\mu\text{g m}^{-3}$ were recorded (Fig. 1a). The highest
 192 PM concentrations observed were similar to those in the world's most polluted cities;^{44,45} however,
 193 peak levels were short-lived. Average PM levels and geometric mean diameters (both number-
 194 and volume-based) are shown in Table 1.



195 **Figure 1.** Temporal variation of size segregated mass concentrations of particulate matter (PM_{0.1}, PM_{0.5},
 196 PM₁, PM_{2.5}, PM₁₀, and PM₂₀) during (a) Thanksgiving Day (TG-2), (b) Layered Day (LD-2), (c) Sequential
 197 Stir-fry experiment (SF-2), and (d) Sequential Ventilation experiments. Relevant activities performed
 198 during the experiments are marked for all experiment types except the Thanksgiving Day experiment. A
 199 detailed activity log for the Thanksgiving Day experiment is presented in Fig. S11. The symbol * indicates
 200 vegetable stir-fry cooked on the hotplate; the symbol # denotes vegetable stir-fry cooked in a cast-iron pan.
 201 The gaps in the PM_{0.5}, PM₁, PM_{2.5}, PM₁₀, and PM₂₀ time-profiles for the Sequential Ventilation experiment
 202 correspond to periods when the UHSAS instrument was sampling outdoor air.

204
 205 In terms of number concentration, UFP dominated indoor PM number concentration during
 206 cooking (Table 1 and Fig. S7-S9) and throughout the Sequential Ventilation experiment (Fig. S10).

207 Indoor PM_{2.5} levels were comparable to prior literature reports, although we note that observed
208 concentrations are the product of not only emission rates but also ventilation conditions and house
209 volume. For example, Wan et al.⁴⁶ reported an average PM_{2.5} concentration of ~160 µg m⁻³ for 30
210 cooking episodes in the kitchens of 12 Hong Kong homes. Long et al.⁴⁷ reported an average PM_{2.5}
211 concentration of 37 ± 31 µg m⁻³ for stir-frying compared to 30 ± 10 µg m⁻³ in this study (Table 1).

212 Comparable PM levels (Table 1) and trends (Fig. 1a and S5) were observed during both
213 Thanksgiving days. These represented the highest UFP levels during HOMEChem, in terms of
214 both mass (as high as 100 µg m⁻³) and number (as high as 2.7×10⁶ cm⁻³, Fig S8). During this
215 experiment, the PM_{2.5} level remained above 50 µg m⁻³ for more than five hours. Multiple cooking
216 activities, using both the oven and stove, were performed in parallel throughout the day (Fig. S5
217 and S11). These PM levels reflect the cumulative influence of emissions from different sources,
218 but most prominently from cooking.

219 Average PM_{2.5} mass concentrations (over 12-h periods) were ~60 µg m⁻³ (Table 1) for both
220 Thanksgiving days, about four times higher than during Layered Days (~15 µg m⁻³). While both
221 types of day-long experiments showed a preponderance of UFP on a number basis (97% for
222 Thanksgiving and 99% for Layered Days), the mass fraction of UFP among total PM was nearly
223 three times higher for Thanksgiving (34%) compared to Layered Days (11%). A larger geometric
224 mean particle size was observed for the Thanksgiving (~18 nm) compared to Layered Days (10 ±
225 2 nm, average ± standard deviation) (Fig. S12a). In terms of particle volume (Fig. S12b), the
226 geometric mean size for Thanksgiving (~180 nm) was much smaller than that for Layered Days
227 (800 ± 200 nm) consistent with the observation that PM_{0.5} constituted a higher fraction of total PM
228 mass during Thanksgiving (86%) compared to the Layered Days (50 ± 10%).

229

230

Table 1. Average size-segregated PM number[#] (<100 nm and >100 nm) and mass (PM_{0.1}, PM_{0.5}, PM₁, PM_{2.5}, PM₁₀, and PM₂₀) concentrations

	Cooking duration (min)	Number Conc. (cm ⁻³) (average ± SD)		Mass Conc. (µg m ⁻³) (average ± SD)					Geometric mean diameter (nm) (average ± SD)		
		N _{<100 nm}	N _{>100 nm}	PM _{0.1}	PM _{0.5}	PM ₁	PM _{2.5}	PM ₁₀	PM ₂₀	By number	By volume
<i>12-hour averages</i>											
Layered Day	NA	(2.2 ± 0.9) × 10 ⁵	(2 ± 1) × 10 ³	2.0 ± 0.5	9 ± 2	11 ± 2	14 ± 2	18 ± 2	19 ± 2	10 ± 2	(8 ± 2) × 10 ²
Thanksgiving Day	NA	7.7 × 10 ⁵ 8.0 × 10 ⁵	2.4 × 10 ⁴ 2.3 × 10 ⁴	23.4 22.0	59.0 56.4	61.1 57.9	63.1 60.9	66.3 64.8	67.7 66.5	18 17	1.7 × 10 ² 1.9 × 10 ²
<i>Meal-wise (averaged over the cooking duration)</i>											
Breakfast	34 ± 10	(1.2 ± 0.5) × 10 ⁶	(10 ± 8) × 10 ³	11 ± 6	30 ± 20	35 ± 20	48 ± 30	80 ± 50	100 ± 50	11 ± 5	(1.8 ± 0.7) × 10 ³
Stir-fry (gas stove)	37 ± 10	(9 ± 5) × 10 ⁵	(3 ± 1) × 10 ³	3 ± 1	13 ± 5	17 ± 7	30 ± 10	44 ± 20	50 ± 20	8 ± 3	(1.4 ± 0.7) × 10 ³
Stir-fry (hot plate)	52 ± 6	(1.5 ± 0.6) × 10 ⁵	(1.9 ± 0.4) × 10 ³	1.8 ± 0.4	10 ± 3	15 ± 7	25 ± 10	40 ± 30	46.3 ± 30	16 ± 5	(1.9 ± 0.5) × 10 ³
Chili	68 ± 3	(8 ± 2) × 10 ⁵	(2 ± 1) × 10 ³	3 ± 1	11 ± 6	15 ± 8	20 ± 10	30 ± 20	40 ± 20	7 ± 1	(1.2 ± 0.1) × 10 ³
Lasagna	75	2.7 × 10 ⁵	3.3 × 10 ³	3.0	11.5	12.0	12.9	13.3	13.4	13	200
Toast	*	(10 ± 2) × 10 ⁴	(8 ± 7) × 10 ³	2.1 ± 0.4	9 ± 2	11 ± 2	12 ± 2	12 ± 2	12 ± 2	16 ± 3	(1.9 ± 0.6) × 10 ²
<i>Sequential Ventilation experiment</i>											
All external doors and windows open	^	(1.5 ± 0.2) × 10 ⁴	(9 ± 3) × 10 ²	0.5 ± 0.3	3 ± 1	4 ± 1	6 ± 1	9 ± 2	10 ± 3	21 ± 8	(1.4 ± 0.3) × 10 ³
<i>Unoccupied background</i>											
NA		(2 ± 1) × 10 ³	(2.3 ± 0.4) × 10 ²	0.13 ± 0.04	1.2 ± 0.3	1.5 ± 0.3	2.3 ± 0.3	2.5 ± 0.4	2.5 ± 0.4	35 ± 10	(6.1 ± 0.9) × 10 ²

#Size-segregated number concentrations of sub-100 number particles are presented in Table S8.

*Since toast preparation took only 2-3 minutes, the scan including the toast preparation and the following scan has been used to calculate average values.

^External doors and windows were opened for 28 ± 6 minutes

232 Comparing average PM levels over the cooking duration by individual meal type, the highest PM
233 mass concentrations—at all PM size ranges—were observed during breakfast, which might be
234 attributed to the cumulative effect of cooking multiple items (sausage, tomato, and egg) in an open
235 pan with oil. Average PM_{2.5} levels for breakfast were about twice those for stir-fry and chili and
236 nearly 4× those for lasagna and toast. However, PM levels for many stir-fry events were
237 comparable or exceeded that for some breakfasts. Lasagna and toast exhibited the highest PM_{2.5}
238 mass fraction (~97% and ~98%, respectively), compared to <65% for other meals. Stir-fry,
239 breakfast, and chili preparation required ingredients to be cooked in hot oil and stirred during the
240 process, which generates coarse particles⁴⁷ and, therefore, lowers PM_{2.5}/PM₂₀ ratios. The formation
241 of UFP observed during cooking is discussed in §3.2.

242 Temporal PM mass concentration (Fig. 1d) and size distribution (Fig. S10) during the Sequential
243 Ventilation experiment showed that indoor PM levels peaked within five minutes after opening
244 the house and remained nearly constant at concentrations of 6–17 µg m⁻³, only ~6% lower than
245 outdoor levels (Fig. S13). As compared with cooking events, UFP constituted smaller fractions of
246 the total PM mass (5 ± 2%) and number (94 ± 2%). After doors and windows were closed, it took
247 44 ± 5 minutes for PM concentrations to return to stable background levels.

248 Mopping did not produce distinct changes in mass concentrations of PM₁ and smaller mass
249 fractions during Layered Days (Fig 1b and S4) and Sequential Cleaning experiments (Fig. S14 and
250 S15). Observed increases in concentrations of larger size fractions during mopping (Fig. S14) are
251 probably attributable to particle release through human movements, such as dust resuspension,
252 shedding of skin cells, and emission of clothing fibers.^{48,49} The formation and growth of sub-30
253 nm particles observed after some mopping events are discussed in §3.3.

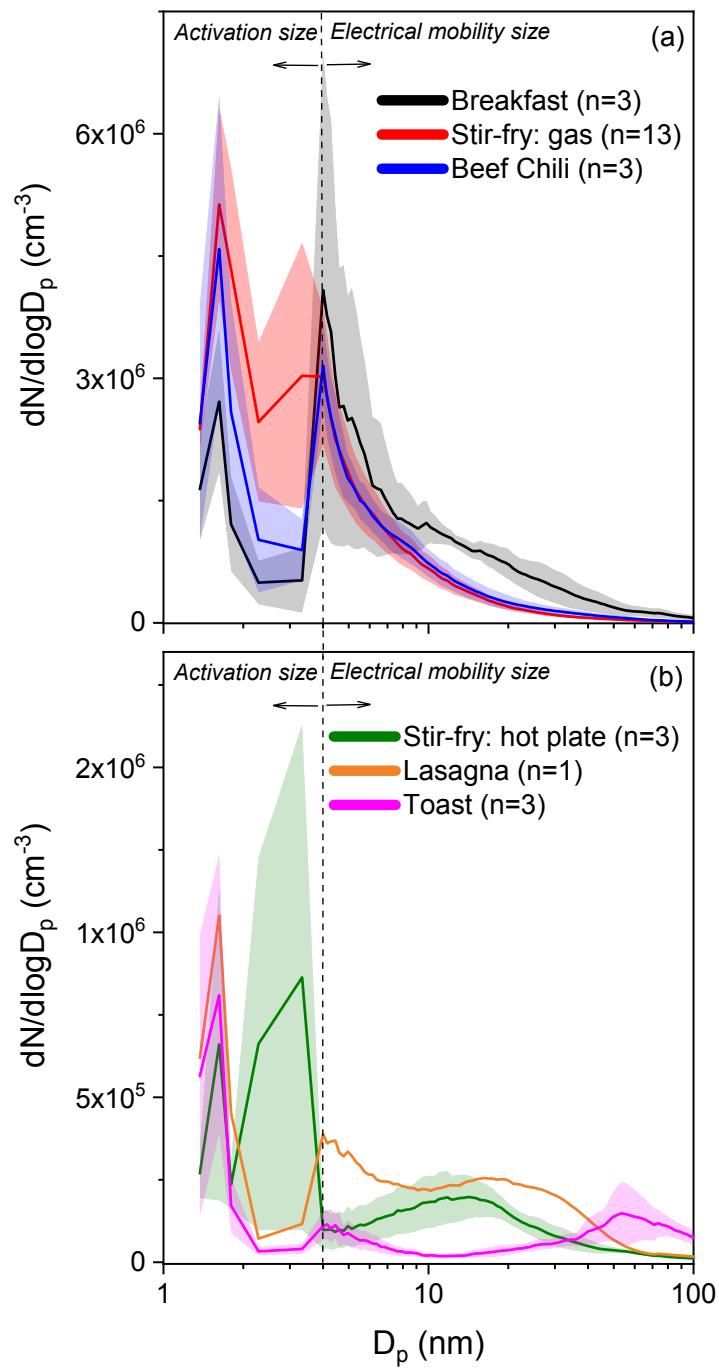
254 *3.2 Particle formation during cooking*

255 For all meals, UFP dominated number concentrations with number geometric mean diameters <20
256 nm, while the volume geometric mean diameter varied from ~200 nm for toast and lasagna to ~2
257 µm for stir-fry (Table 1). Multi-modal size distributions were observed for all meals (Fig. 2) with
258 1-2 modes in size range measured by diethylene glycol (DEG) activation (<4 nm). Peaks at 1.6 nm
259 may be the result of clusters of gaseous species activating and growing with DEG as observed in
260 a study of atmospheric aerosol nucleation.⁵⁰ A study of traffic-related emissions also reported
261 multi-modal size distributions in the DEG activation size range.⁵¹

262 In terms of PM > 4 nm, the number mode from toast emissions was 50 ± 10 nm, which is
263 comparable to the 30-50 nm range reported for toasters in previous studies.⁵²⁻⁵⁴ High
264 concentrations of sub-10 nm particles were observed for meals cooked on the gas stove (breakfast,
265 stir-fry, and chili, Fig. 2a) with number modes in the range 4-11 nm. Even though the oven was
266 also propane-fuelled, sub-10 nm particle concentrations were 5-7× lower for the oven-baked
267 lasagna (Fig. 2b) compared to the meals cooked on the stove. PM emissions taking place in the
268 oven might be subjected to enhanced condensational growth, coagulation, and wall losses,
269 especially for sub-10 nm particles.

270 During cooking, particles can originate from both the heat source and the food, leading to some
271 distinguishing source-specific characteristics. The combustion of gaseous fuels for cooking is
272 known to generate high numbers of sub-10 nm particles⁶ and electric appliances, such as a toaster,
273 a sandwich maker, and a hot plate (with no food) have also been shown to release large numbers
274 of <10-nm particles.^{30,55}

275



276

277 **Figure 2.** Ultrafine particle number size distributions, averaged over the cooking duration, for different
 278 types of meals, plotted in two panels with different y -axis scaling to facilitate visualization: (a) breakfast,
 279 vegetable stir-fry cooked on gas, and beef chili and (b) vegetable stir-fry cooked on the hot plate, lasagna,
 280 and toast. The shaded region represents standard error. n = number of replicates.
 281

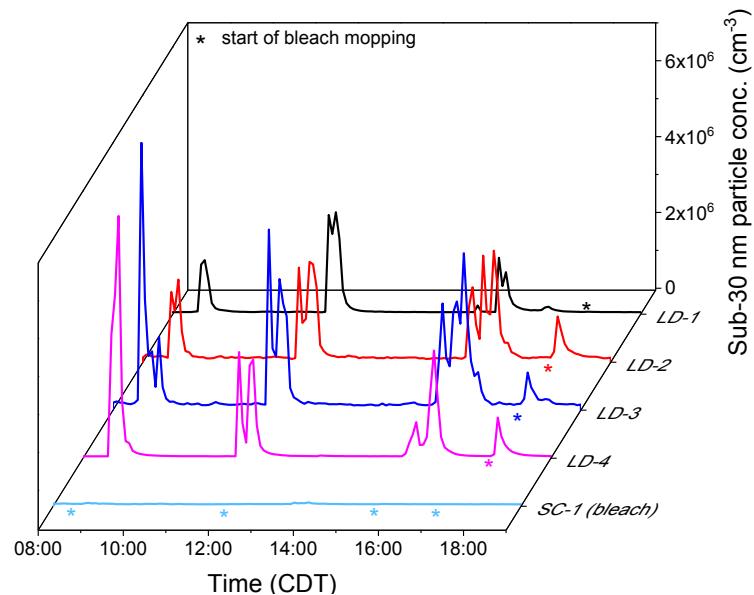
282 Naked hot plate and naked stove experiments were performed to characterize PM generated during
283 the operation of just the heat source. Indoor particle number concentrations were $>1000\times$ higher
284 than background levels during the operation of the stove and hot plate (with no food or pans).
285 While most particles were <10 nm for both the stove (99.2%) and hot plate (96.2%), the
286 concentration of sub-10 nm particles from operating the stove ($7.6 \times 10^6 \text{ cm}^{-3}$) was almost 7×
287 higher than from operating the hot plate ($1.1 \times 10^6 \text{ cm}^{-3}$, Fig. S16a). Wallace et al.⁶ also reported
288 higher emissions of sub-10 nm particles from a gas burner compared to an electric stovetop.
289 Additionally, we observed a second aerosol mode at ~ 21 nm for the hot plate (Fig. S16a). A similar
290 observation made by Wallace et al.⁶ was attributed to a thin film formation or deposition of dust
291 on the stovetop surface between uses, and this explanation was further strengthened in another
292 study⁵⁵ where the continuous use of a hot plate eventually led to zero UFP emissions, but the same
293 hot plate could again generate UFP when heated after a few days of no use. Residues of detergents,
294 skin oils, and organics accumulated on a cooking pot surface can also generate UFP at rates varying
295 with factors such as pan type, temperature, and duration of disuse.^{55,56}

296 Sub-10 nm particle concentrations for stir-fry cooked on the stove ($5.5 \pm 0.4 \times 10^7 \text{ cm}^{-3}$) were $>8\times$
297 higher than for those cooked on the hot plate ($6.3 \pm 0.6 \times 10^6 \text{ cm}^{-3}$, Fig. S16b). Because particles
298 originating from the heat source were too small to affect PM mass, most of the mass-based PM
299 emissions observed during cooking can be attributed to food ingredients, especially oils.³¹ A
300 previous study⁷ reported a much lower sub-100 nm particle concentration ($2.7 \times 10^4 \text{ cm}^{-3}$) for
301 vegetable stir-fry because their lower size cut-off was 10 nm, and therefore, a big fraction of
302 particles emitted from the gas stove was not measured. The same study also reported a much lower
303 particle diameter mode (118 nm) because their upper size cut-off was 950 nm. In this study, we
304 observed a mode beyond 1 μm (Fig. S16).

305 *3.3 Ultrafine particle formation during mopping*

306 Few studies have investigated the effects of chlorine-based cleaners on particle-phase chemistry
307 indoors. Wong et al.¹² recorded increases in particle-phase chlorine after each bleach cleaning as
308 measured by an aerosol mass spectrometer but could not attribute the increase in particle-phase
309 chlorine to a chemical or physical process. Various studies have investigated chemical pathways
310 taken by chlorinated species originating from bleach,^{12,57,58} but these investigations focused on
311 gas-phase chemistry or on uptake of gas species on PM surfaces.⁵⁹ Wang et al.⁶⁰ reported particle
312 formation from terpenes and bleach emissions upon illumination via radical chemistry initiated by
313 the photolysis of HOCl and Cl₂.

314 During HOMEChem, the living room and kitchen floors were mopped with a bleach solution for
315 10 minutes on the evenings of Layered Days. The living room had large west-facing windows and
316 the kitchen had small east-facing windows. Particle formation and growth (up to ~30 nm) were
317 observed within five minutes of the start of mopping (Fig. 3). Peak number concentrations were
318 comparable to those associated with cooking; however, the resulting increase in PM mass
319 concentration was <10 ng m⁻³. The time-series concentration of the smallest particles, a proxy
320 indicator for nucleation, peaked within ten minutes of the start of mopping (Fig S17). One size
321 distribution was recorded every five minutes and, therefore, the transition from particle formation
322 via nucleation to particle growth via condensation might not have been fully captured.



323

324 **Figure 3.** Sub-30 nm particle concentration in the test house during the four Layered Day experiments
325 (coded as LD-1, LD-2, LD-3, and LD-4) and one Sequential Cleaning experiment with bleach (coded as
326 SC-1). During LD-1, a lasagna was cooked for dinner, whereas chili was prepared for all other Layered
327 Day experiments.

328

329 Interestingly, no new particle formation was observed in association with any of the bleach
330 mopping events during the two Sequential Cleaning experiments (Fig. 3 and Fig. S15a-b) or during
331 the Layered Day in which bleach mopping was preceded by lasagna instead of chili cooking. We
332 hypothesize that residual gas-phase emissions associated with chili cooking and bleach cleaning
333 participated in the physicochemical processes leading to new particle formation.

334 *3.4 PM exposure and uptake in the human respiratory system*

335 The United States Environmental Protection Agency (USEPA) ambient air quality standard for
336 PM_{2.5} is 35 $\mu\text{g m}^{-3}$ for 24 hours, corresponding to a cumulative daily exposure of 840 $\mu\text{g m}^{-3} \text{ h}$.
337 PM_{2.5} exposure levels over 12-h periods during the Thanksgiving events ($\sim 740 \mu\text{g m}^{-3} \text{ h}$) were
338 close to this health-based exposure limit. Exposures for 12-h periods on Layered Days (170 ± 20
339 $\mu\text{g m}^{-3} \text{ h}$) were about a quarter of those during the Thanksgiving experiments, well below the

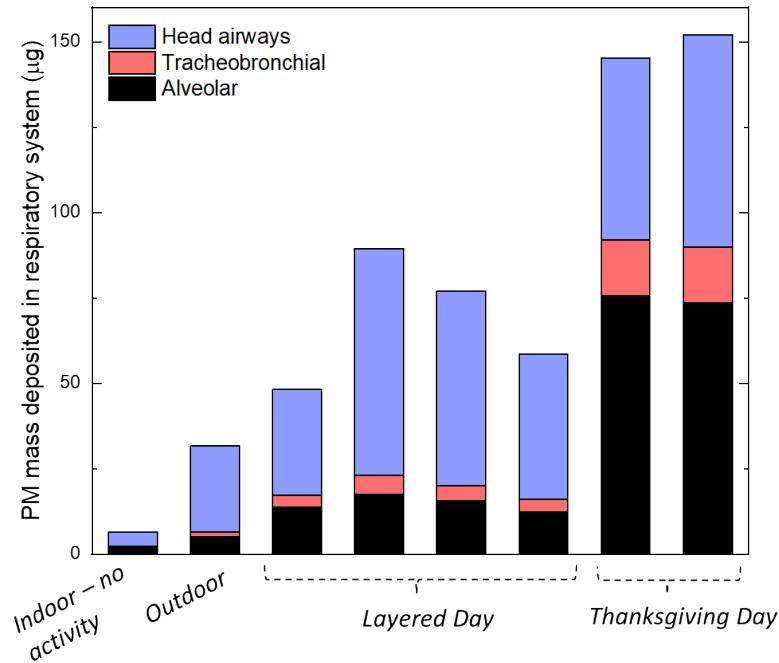
340 USEPA threshold. Table 1 focussed on PM concentrations averaged over the cooking duration. To
341 quantify the importance of the post-cooking decay period, Table 2 presents the exposures during
342 cooking and post-cooking decay periods. These post-cooking periods contributed 36-77% of total
343 exposure for each type of cooking activity. Exposures during breakfast, stir-fry, and chili cooking
344 were comparable, but average exposure during the post-cooking decay was $>2\times$ higher for
345 breakfast compared to chili.

346 **Table 2.** Average exposures for $\text{PM}_{2.5}$ ($\mu\text{g m}^{-3} \text{ h}$) over the duration of cooking and post-cooking decay period
347 for different meals cooked during the Layered Day (breakfast, stir-fry, toast, beef chili, and lasagna) and
348 Sequential Day experiments (stir-fry).

Meal	During cooking	During post-cooking decay period
Breakfast	25 ± 20	40 ± 20
Stir-fry	20 ± 10	30 ± 20
Beef chili	30 ± 10	15 ± 10
Lasagna	16	10
Toast	3 ± 1	10 ± 4

349

350 Fig. 4 presents the PM respiratory deposition (0.004-20 μm) calculated for an adult residing in the
351 test house from 8 AM to 8 PM on a Layered Day and on Thanksgiving Day. Two comparison
352 cases, Indoor-no activity and Outdoor, are also included. The Indoor-no activity case represents
353 the lower bound of indoor PM respiratory deposition during HOMEChem; all windows and doors
354 were closed, and no activities were performed. The Outdoor bar represents the respiratory
355 deposition of an adult spending the same duration outside the test house. Lacking direct outdoor
356 measurements, indoor data recorded when doors and windows were open (Sequential Ventilation,
357 Table 1) were used as a proxy for outdoor data (Fig. S13).



358

359 **Figure 4.** Estimated PM mass deposited in different parts of the respiratory system (head airways
360 tracheobronchial, and alveolar) of an adult residing in the test house kitchen during four Layered Day
361 experiments and two Thanksgivings. Indoor-no activity represents the potential exposure of an adult
362 residing in the test house kitchen when no activities were performed and Outdoor represents spending the
363 same amount of time outside the test house. For each bar, the duration of exposure is 12 h, spanning 8 AM
364 to 8 PM.

365

366 The respiratory deposition during Indoor-no activity (7 µg) was lower than that corresponding to
367 Outdoors (32 µg), as expected. In the absence of indoor sources, the house envelope provides
368 partial protection from outdoor PM. Respiratory deposition masses were highest during
369 Thanksgiving (145 µg and 152 µg) followed by Layered Day (70 ± 20 µg), 5× and ~2× higher,
370 respectively, than that for Outdoors. These results suggest that in the US and in other countries
371 that generally meet outdoor air quality standards, indoor exposure levels may dominate the overall
372 exposure for groups of people who spend most of their time in residences, especially if cooking
373 activities are frequent.

374 The PM fraction deposited in the head airways dominated (~75%) total mass deposited for all
375 experiments except for the two Thanksgiving days, when the fraction deposited in the alveolar

376 region (~50%) was higher than in the head airways (~39%). During Thanksgiving, UFP were
377 prominent contributors to PM mass, and particle deposition efficiency in the alveolar region is
378 particularly high for this size range.

379 Respiratory depositions corresponding to the combined cooking and post-cooking decay period
380 for different meals (Table S9) follow the same trends as exposures (Table 2). For all meal types,
381 <10% of the total respiratory deposition occurred in the tracheobronchial region. The head airways
382 fraction represented ~70% for stir-fry, chili, and breakfast and ~50% for lasagna and toast. The
383 highest fraction of alveolar deposition (~50% of total) was obtained for toast and lasagna.

384 *3.5 Implications*

385 The test house was operated continuously at fixed ventilation and recirculation rates (without filter)
386 to characterize indoor air pollutants in a simplified setting relative to a real house. Therefore, the
387 test house is not representative of a typical residence with regards to ventilation. However, the
388 measured PM concentrations and calculated exposures and resulting respiratory deposition during
389 simulated daily activities highlight the importance of seeking a deeper understanding of indoor air
390 quality in residences. In the absence of indoor sources, PM exposure and respiratory deposition in
391 indoor environments are expected to be lower than outdoors, but cooking (and potentially other)
392 activities can lead to high indoor concentrations of PM and, therefore, potentially higher indoor
393 exposure levels compared to spending the same amount of time outdoors. PM exposures during
394 the post-cooking decay phase were higher than those during the cooking phase for some meals.
395 Exposure to cooking emissions is likely also important for house occupants away from the kitchen
396 area. High concentrations of UFP were observed during all cooking activities and propane
397 combustion emitted mostly sub-10 nm particles. We lack an understanding of the health effects of
398 exposure to high number concentrations of UFP, especially sub-10 nm particles from stove and

399 hot plate, which rarely accumulate to a high enough mass concentration to be of concern based on
400 current air quality standards and guidelines. Apart from the apparent source of PM, i.e. cooking,
401 new particle formation from the interplay between the emissions from cooking and bleach
402 demonstrates that there might be other, non-obvious sources of UFP in residential environments.

403

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409 experimental design.

410

411 **Supporting Information (SI)**

412 • Descriptions of instrumentation and settings with figures; discussion, figures, and tables on
413 particulate matter density assumption and associated uncertainty; equations used for lung
414 deposition modeling; additional figures showing measurements during cooking, cleaning, and
415 ventilation experiments.

416 **References**

- 417 (1) Cohen, A. J.; Brauer, M.; Burnett, R.; Anderson, H. R.; Frostad, J.; Estep, K.; Balakrishnan, K.;
418 Brunekreef, B.; Dandona, L.; Dandona, R.; Feigin, V.; Freedman, G.; Hubbell, B.; Jobling, A.; Kan, H.; Knibbs, L.; Liu, Y.; Martin, R.; Morawska, L.; Pope, C. A.; Shin, H.; Straif, K.; Shaddick, G.; Thomas, M.; van Dingenen, R.; van Donkelaar, A.; Vos, T.; Murray, C. J. L.; Forouzanfar, M. H. Estimates and 25-Year Trends of the Global Burden of Disease Attributable to Ambient Air Pollution: An Analysis of Data from the Global Burden of Diseases Study 2015. *The Lancet* **2017**, 389 (10082), 1907–1918. [https://doi.org/10.1016/S0140-6736\(17\)30505-6](https://doi.org/10.1016/S0140-6736(17)30505-6).
- 419 (2) Zhang, Y.; West, J. J.; Mathur, R.; Xing, J.; Hogrefe, C.; Roselle, S. J.; Bash, J. O.; Pleim, J. E.; Gan, C.-M.; Wong, D. C. Long-Term Trends in the Ambient PM_{2.5}- and O₃-Related Mortality Burdens in

- 426 the United States under Emission Reductions from 1990 to 2010. *Atmospheric Chemistry and*
427 *Physics* **2018**, *18* (20), 15003–15016. <https://doi.org/10.5194/acp-18-15003-2018>.
- 428 (3) Klepeis, N. E.; Nelson, W. C.; Ott, W. R.; Robinson, J. P.; Tsang, A. M.; Switzer, P.; Behar, J. V.;
429 Hern, S. C.; Engelmann, W. H. The National Human Activity Pattern Survey (NHAPS): A Resource
430 for Assessing Exposure to Environmental Pollutants. *Journal of Exposure Analysis and*
431 *Environmental Epidemiology* **2001**, *11* (3), 231–252. <https://doi.org/10.1038/sj.jea.7500165>.
- 432 (4) Weisel, C. P.; Zhang, J.; Turpin, B. J.; Morandi, M. T.; Colome, S.; Stock, T. H.; Spektor, D. M.; Korn,
433 L.; Winer, A.; Alimokhtari, S.; Kwon, J.; Mohan, K.; Harrington, R.; Giovanetti, R.; Cui, W.; Afshar,
434 M.; Maberti, S.; Shendell, D. Relationship of Indoor, Outdoor and Personal Air (RIOPA) Study:
435 Study Design, Methods and Quality Assurance/Control Results. *Journal of Exposure Analysis and*
436 *Environmental Epidemiology* **2005**, *15* (2), 123–137. <https://doi.org/10.1038/sj.jea.7500379>.
- 437 (5) Avery, A. M.; Waring, M. S.; DeCarlo, P. F. Seasonal Variation in Aerosol Composition and
438 Concentration upon Transport from the Outdoor to Indoor Environment. *Environmental Science:*
439 *Processes & Impacts* **2019**, *21* (3), 528–547.
- 440 (6) Wallace, L.; Wang, F.; Howard-Reed, C.; Persily, A. Contribution of Gas and Electric Stoves to
441 Residential Ultrafine Particle Concentrations between 2 and 64 Nm: Size Distributions and
442 Emission and Coagulation Rates. *Environmental Science & Technology* **2008**, *42* (23), 8641–8647.
- 443 (7) Wallace, L. Indoor Sources of Ultrafine and Accumulation Mode Particles: Size Distributions, Size-
444 Resolved Concentrations, and Source Strengths. *Aerosol Science and Technology* **2006**, *40* (5),
445 348–360. <https://doi.org/10.1080/02786820600612250>.
- 446 (8) Xu, M.; Nematollahi, M.; Sextro, R. G.; Gadgil, A. J.; Nazaroff, W. W. Deposition of Tobacco Smoke
447 Particles in a Low Ventilation Room. *Aerosol Science and Technology* **1994**, *20* (2), 194–206.
448 <https://doi.org/10.1080/02786829408959676>.
- 449 (9) Brauer, M.; Hirtle, R.; Lang, B.; Ott, W. Assessment of Indoor Fine Aerosol Contributions from
450 Environmental Tobacco Smoke and Cooking with a Portable Nephelometer. *Journal of Exposure*
451 *Analysis and Environmental Epidemiology* **2000**, *10* (2), 136–144.
452 <https://doi.org/10.1038/sj.jea.7500076>.
- 453 (10) DeCarlo, P. F.; Avery, A. M.; Waring, M. S. Thirdhand Smoke Uptake to Aerosol Particles in the
454 Indoor Environment. *Science Advances* **2018**, *4* (5), eaap8368.
455 <https://doi.org/10.1126/sciadv.aap8368>.
- 456 (11) Singer, B. C.; Coleman, B. K.; Destaillats, H.; Hodgson, A. T.; Lunden, M. M.; Weschler, C. J.;
457 Nazaroff, W. W. Indoor Secondary Pollutants from Cleaning Product and Air Freshener Use in the
458 Presence of Ozone. *Atmospheric Environment* **2006**, *40* (35), 6696–6710.
459 <https://doi.org/10.1016/j.atmosenv.2006.06.005>.
- 460 (12) Wong, J. P. S.; Carslaw, N.; Zhao, R.; Zhou, S.; Abbatt, J. P. D. Observations and Impacts of Bleach
461 Washing on Indoor Chlorine Chemistry. *Indoor Air* **2017**, *27* (6), 1082–1090.
462 <https://doi.org/10.1111/ina.12402>.
- 463 (13) Uhde, E.; Salthammer, T. Impact of Reaction Products from Building Materials and Furnishings on
464 Indoor Air Quality—A Review of Recent Advances in Indoor Chemistry. *Atmospheric Environment*
465 **2007**, *41* (15), 3111–3128. <https://doi.org/10.1016/j.atmosenv.2006.05.082>.
- 466 (14) Yao, M.; Zhao, B. SOA in Newly Decorated Residential Buildings. *Building and Environment* **2017**,
467 *111*, 132–139. <https://doi.org/10.1016/j.buildenv.2016.10.022>.
- 468 (15) Steinemann, A. Fragranced Consumer Products: Exposures and Effects from Emissions. *Air Qual*
469 *Atmos Health* **2016**, *9* (8), 861–866. <https://doi.org/10.1007/s11869-016-0442-z>.
- 470 (16) Morawska, L.; Xiu, M.; He, C.; Buonanno, G.; McGarry, P.; Maumy, B.; Stabile, L.; Thai, P. K.
471 Particle Emissions from Laser Printers: Have They Decreased? *Environ. Sci. Technol. Lett.* **2019**, *6*
472 (5), 300–305. <https://doi.org/10.1021/acs.estlett.9b00176>.

- 473 (17) Avery, A. M.; Waring, M. S.; DeCarlo, P. F. Human Occupant Contribution to Secondary Aerosol
474 Mass in the Indoor Environment. *Environmental Science: Processes & Impacts* **2019**, *21* (8), 1301–
475 1312.
- 476 (18) Abbatt, J. P. D.; Wang, C. The Atmospheric Chemistry of Indoor Environments. *Environ. Sci.:
477 Processes Impacts* **2019**, DOI: 10.1039/C9EM00386J. <https://doi.org/10.1039/C9EM00386J>.
- 478 (19) Pope, C. A.; Dockery, D. W.; Schwartz, J. Review of Epidemiological Evidence of Health Effects of
479 Particulate Air Pollution. *Inhalation Toxicology* **1995**, *7* (1), 1–18.
480 <https://doi.org/10.3109/08958379509014267>.
- 481 (20) Li, N.; Hao, M.; Phalen, R. F.; Hinds, W. C.; Nel, A. E. Particulate Air Pollutants and Asthma: A
482 Paradigm for the Role of Oxidative Stress in PM-Induced Adverse Health Effects. *Clinical
483 Immunology* **2003**, *109* (3), 250–265. <https://doi.org/10.1016/j.clim.2003.08.006>.
- 484 (21) Patel, S.; Leavey, A.; Sheshadri, A.; Kumar, P.; Kandikappa, S.; Tarsi, J.; Mukhopadhyay, K.;
485 Johnson, P.; Balakrishnan, K.; Schechtman, K. B.; Castro, M.; Yadama, G.; Biswas, P. Associations
486 between Household Air Pollution and Reduced Lung Function in Women and Children in Rural
487 Southern India. *Journal of Applied Toxicology* **2018**, *38* (11), 1405–1415.
488 <https://doi.org/10.1002/jat.3659>.
- 489 (22) Stanaway, J. D.; Afshin, A.; Gakidou, E.; Lim, S. S.; Abate, D.; Abate, K. H.; Abbafati, C.; Abbasi, N.;
490 Abbastabar, H.; Abd-Allah, F.; Abdela, J.; Abdelalim, A.; Abdollahpour, I.; Abdulkader, R. S.;
491 Abebe, M.; Abebe, Z.; Abera, S. F.; Abil, O. Z.; Abraha, H. N.; Abrham, A. R.; Abu-Raddad, L. J.;
492 Abu-Rmeileh, N. M.; Accrombessi, M. M. K.; Acharya, D.; Acharya, P.; Adamu, A. A.; Adane, A. A.;
493 Adebayo, O. M.; Adedoyin, R. A.; Adekanmbi, V.; Ademi, Z.; Adetokunboh, O. O.; Adib, M. G.;
494 Admasie, A.; Adsuar, J. C.; Afanvi, K. A.; Afarideh, M.; Agarwal, G.; Aggarwal, A.; Aghayan, S. A.;
495 Agrawal, A.; Agrawal, S.; Ahmadi, A.; Ahmadi, M.; Ahmadieh, H.; Ahmed, M. B.; Aichour, A. N.;
496 Aichour, I.; Aichour, M. T. E.; Akbari, M. E.; Akinyemiju, T.; Akseer, N.; Al-Aly, Z.; Al-Eyadhy, A.; Al-
497 Mekhlafi, H. M.; Alahdab, F.; Alam, K.; Alam, S.; Alam, T.; Alashi, A.; Alavian, S. M.; Alene, K. A.;
498 Ali, K.; Ali, S. M.; Alijanzadeh, M.; Alizadeh-Navaei, R.; Aljunid, S. M.; Alkerwi, A.; Alla, F.; Alsharif,
499 U.; Altirkawi, K.; Alvis-Guzman, N.; Amare, A. T.; Ammar, W.; Anber, N. H.; Anderson, J. A.; Andrei,
500 C. L.; Androudi, S.; Animut, M. D.; Anjomshoa, M.; Ansha, M. G.; Antó, J. M.; Antonio, C. A. T.;
501 Anwari, P.; Appiah, L. T.; Appiah, S. C. Y.; Arabloo, J.; Aremu, O.; Ärnlöv, J.; Artaman, A.; Aryal, K.
502 K.; Asayesh, H.; Ataro, Z.; Ausloos, M.; Avokpaho, E. F. G. A.; Awasthi, A.; Quintanilla, B. P. A.;
503 Ayer, R.; Ayuk, T. B.; Azzopardi, P. S.; Babazadeh, A.; Badali, H.; Badawi, A.; Balakrishnan, K.; Bali,
504 A. G.; Ball, K.; Ballew, S. H.; Banach, M.; Banoub, J. A. M.; Barac, A.; Barker-Collo, S. L.;
505 Bärnighausen, T. W.; Barrero, L. H.; Basu, S.; Baune, B. T.; Bazargan-Hejazi, S.; Bedi, N.; Beghi, E.;
506 Behzadifar, M.; Behzadifar, M.; Béjot, Y.; Bekele, B. B.; Bekru, E. T.; Belay, E.; Belay, Y. A.; Bell, M.
507 L.; Bello, A. K.; Bennett, D. A.; Bensenor, I. M.; Bergeron, G.; Berhane, A.; Bernabe, E.; Bernstein,
508 R. S.; Beuran, M.; Beyranvand, T.; Bhala, N.; Bhalla, A.; Bhattarai, S.; Bhutta, Z. A.; Biadgo, B.;
509 Bijani, A.; Bikbov, B.; Bilano, V.; Billilign, N.; Sayeed, M. S. B.; Bisanzio, D.; Biswas, T.; Bjørge, T.;
510 Blacker, B. F.; Bleyer, A.; Borschmann, R.; Bou-Orm, I. R.; Boufous, S.; Bourne, R.; Brady, O. J.;
511 Brauer, M.; Brazinova, A.; Breitborde, N. J. K.; Brenner, H.; Briko, A. N.; Britton, G.; Brugha, T.;
512 Buchbinder, R.; Burnett, R. T.; Busse, R.; Butt, Z. A.; Cahill, L. E.; Cahuana-Hurtado, L.; Campos-
513 Nonato, I. R.; Cárdenas, R.; Carreras, G.; Carrero, J. J.; Carvalho, F.; Castañeda-Orjuela, C. A.;
514 Rivas, J. C.; Castro, F.; Catalá-López, F.; Causey, K.; Cercy, K. M.; Cerin, E.; Chaiah, Y.; Chang, H.-Y.;
515 Chang, J.-C.; Chang, K.-L.; Charlson, F. J.; Chattopadhyay, A.; Chattu, V. K.; Chee, M. L.; Cheng, C.-
516 Y.; Chew, A.; Chiang, P. P.-C.; Chimed-Ochir, O.; Chin, K. L.; Chitheer, A.; Choi, J.-Y. J.; Chowdhury,
517 R.; Christensen, H.; Christopher, D. J.; Chung, S.-C.; Cicuttini, F. M.; Cirillo, M.; Cohen, A. J.;
518 Collado-Mateo, D.; Cooper, C.; Cooper, O. R.; Coresh, J.; Cornaby, L.; Cortesi, P. A.; Cortinovis, M.;
519 Costa, M.; Cousin, E.; Criqui, M. H.; Cromwell, E. A.; Cundiff, D. K.; Daba, A. K.; Dachew, B. A.;
520 Dadi, A. F.; Damasceno, A. A. M.; Dandona, L.; Dandona, R.; Darby, S. C.; Dargan, P. I.; Daryani, A.;

521 Gupta, R. D.; Neves, J. D.; Dasa, T. T.; Dash, A. P.; Davitoiu, D. V.; Davletov, K.; Cruz-Góngora, Ia V.
522 D.; Hoz, F. P. D. L.; Leo, D. D.; Neve, J.-W. D.; Degenhardt, L.; Deiparine, S.; Dellavalle, R. P.;
523 Demoz, G. T.; Denova-Gutiérrez, E.; Deribe, K.; Dervenis, N.; Deshpande, A.; Jarlais, D. C. D.;
524 Dessie, G. A.; Deveber, G. A.; Dey, S.; Dharmaratne, S. D.; Dhimal, M.; Dinberu, M. T.; Ding, E. L.;
525 Diro, H. D.; Djalalinia, S.; Do, H. P.; Dokova, K.; Doku, D. T.; Doyle, K. E.; Driscoll, T. R.; Dubey, M.;
526 Dubljanin, E.; Duken, E. E.; Duncan, B. B.; Duraes, A. R.; Ebert, N.; Ebrahimi, H.; Ebrahimpour, S.;
527 Edvardsson, D.; Effiong, A.; Eggen, A. E.; Bcheraoui, C. E.; El-Khatib, Z.; Elyazar, I. R.; Enayati, A.;
528 Endries, A. Y.; Er, B.; Erskine, H. E.; Eskandarieh, S.; Esteghamati, A.; Estep, K.; Fakhim, H.;
529 Faramarzi, M.; Fareed, M.; Farid, T. A.; Farinha, sá C. S. E.; Farioli, A.; Faro, A.; Farvid, M. S.;
530 Farzaei, M. H.; Fatima, B.; Fay, K. A.; Fazaeli, A. A.; Feigin, V. L.; Feigl, A. B.; Fereshtehnejad, S.-M.;
531 Fernandes, E.; Fernandes, J. C.; Ferrara, G.; Ferrari, A. J.; Ferreira, M. L.; Filip, I.; Finger, J. D.;
532 Fischer, F.; Foigt, N. A.; Foreman, K. J.; Fukumoto, T.; Fullman, N.; Fürst, T.; Furtado, J. M.; Futran,
533 N. D.; Gall, S.; Gallus, S.; Gamkrelidze, A.; Ganji, M.; Garcia-Basteiro, A. L.; Gardner, W. M.; Gebre,
534 A. K.; Gebremedhin, A. T.; Gebremichael, T. G.; Gelano, T. F.; Geleijnse, J. M.; Geramo, Y. C. D.;
535 Gething, P. W.; Gezae, K. E.; Ghadimi, R.; Ghadiri, K.; Falavarjani, K. G.; Ghasemi-Kasman, M.;
536 Ghimire, M.; Ghosh, R.; Ghoshal, A. G.; Giampaoli, S.; Gill, P. S.; Gill, T. K.; Gillum, R. F.; Ginawi, I.
537 A.; Giussani, G.; Gnedovskaya, E. V.; Godwin, W. W.; Goli, S.; Gómez-Dantés, H.; Gona, P. N.;
538 Gopalani, S. V.; Goulart, A. C.; Grada, A.; Grams, M. E.; Grossi, G.; Gugnani, H. C.; Guo, Y.; Gupta,
539 R.; Gupta, R.; Gupta, T.; Gutiérrez, R. A.; Gutiérrez-Torres, D. S.; Haagsma, J. A.; Habtewold, T. D.;
540 Hachinski, V.; Hafezi-Nejad, N.; Hagos, T. B.; Hailegiyorgis, T. T.; Hailu, G. B.; Haj-Mirzaian, A.; Haj-
541 Mirzaian, A.; Hamadeh, R. R.; Hamidi, S.; Handal, A. J.; Hankey, G. J.; Hao, Y.; Harb, H. L.;
542 Harikrishnan, S.; Haro, J. M.; Hassankhani, H.; Hassen, H. Y.; Havmoeller, R.; Hawley, C. N.; Hay, S.
543 I.; Hedayatizadeh-Omrani, A.; Heibati, B.; Heidari, B.; Heidari, M.; Hendrie, D.; Henok, A.; Heredia-
544 Pi, I.; Herteliu, C.; Heydarpour, F.; Heydarpour, S.; Hibstu, D. T.; Higazi, T. B.; Hilawe, E. H.; Hoek,
545 H. W.; Hoffman, H. J.; Hole, M. K.; Rad, E. H.; Hoogar, P.; Hosgood, H. D.; Hosseini, S. M.;
546 Hosseinzadeh, M.; Hostiuc, M.; Hostiuc, S.; Hoy, D. G.; Hsairi, M.; Hsiao, T.; Hu, G.; Hu, H.; Huang,
547 J. J.; Hussen, M. A.; Huynh, C. K.; Ibburg, K. M.; Ikeda, N.; Ilesanmi, O. S.; Iqbal, U.; Irvani, S. S. N.;
548 Irvine, C. M. S.; Islam, S. M. S.; Islami, F.; Jackson, M. D.; Jacobsen, K. H.; Jahangiry, L.; Jahanmehr,
549 N.; Jain, S. K.; Jakovljevic, M.; James, S. L.; Jassal, S. K.; Jayatilleke, A. U.; Jeemon, P.; Jha, R. P.;
550 Jha, V.; Ji, J. S.; Jonas, J. B.; Jonnagaddala, J.; Shushtari, Z. J.; Joshi, A.; Jozwiak, J. J.; Jürisson, M.;
551 Kabir, Z.; Kahsay, A.; Kalani, R.; Kanchan, T.; Kant, S.; Kar, C.; Karami, M.; Matin, B. K.; Karch, A.;
552 Karema, C.; Karimi, N.; Karimi, S. M.; Kasaeian, A.; Kassa, D. H.; Kassa, G. M.; Kassa, T. D.;
553 Kassebaum, N. J.; Katikireddi, S. V.; Kaul, A.; Kawakami, N.; Kazemi, Z.; Karyani, A. K.; Kefale, A. T.;
554 Keiyoro, P. N.; Kemp, G. R.; Kengne, A. P.; Keren, A.; Kesavachandran, C. N.; Khader, Y. S.; Khafaei,
555 B.; Khafaie, M. A.; Khajavi, A.; Khalid, N.; Khalil, I. A.; Khan, G.; Khan, M. S.; Khan, M. A.; Khang, Y.-
556 H.; Khater, M. M.; Khazaei, M.; Khazaie, H.; Khoja, A. T.; Khosravi, A.; Khosravi, M. H.; Kiadaliri, A.
557 A.; Kiarithio, D. N.; Kim, C.-I.; Kim, D.; Kim, Y.-E.; Kim, Y. J.; Kimokoti, R. W.; Kinfu, Y.; Kisa, A.;
558 Kissimova-Skarbek, K.; Kivimäki, M.; Knibbs, L. D.; Knudsen, A. K. S.; Kochhar, S.; Kokubo, Y.;
559 Kolola, T.; Kopec, J. A.; Kosen, S.; Koul, P. A.; Koyanagi, A.; Kravchenko, M. A.; Krishan, K.; Krohn,
560 K. J.; Kromhout, H.; Defo, B. K.; Bicer, B. K.; Kumar, G. A.; Kumar, M.; Kuzin, I.; Kyu, H. H.; Lachat,
561 C.; Lad, D. P.; Lad, S. D.; Lafranconi, A.; Laloo, R.; Lallukka, T.; Lami, F. H.; Lang, J. J.; Lansingh, V.
562 C.; Larson, S. L.; Latifi, A.; Lazarus, J. V.; Lee, P. H.; Leigh, J.; Leili, M.; Leshargie, C. T.; Leung, J.;
563 Levi, M.; Lewycka, S.; Li, S.; Li, Y.; Liang, J.; Liang, X.; Liao, Y.; Liben, M. L.; Lim, L.-L.; Linn, S.; Liu, S.;
564 Lodha, R.; Logroscino, G.; Lopez, A. D.; Lorkowski, S.; Lotufo, P. A.; Lozano, R.; Lucas, T. C. D.;
565 Lunevicius, R.; Ma, S.; Macarayan, E. R. K.; Machado, I. E.; Madotto, F.; Mai, H. T.; Majdan, M.;
566 Majdzadeh, R.; Majeed, A.; Malekzadeh, R.; Malta, D. C.; Mamun, A. A.; Manda, A.-L.; Manguerra,
567 H.; Mansournia, M. A.; Mantovani, L. G.; Maravilla, J. C.; Marcenes, W.; Marks, A.; Martin, R. V.;
568 Martins, S. C. O.; Martins-Melo, F. R.; März, W.; Marzan, M. B.; Massenburg, B. B.; Mathur, M. R.;

569 Mathur, P.; Matsushita, K.; Maulik, P. K.; Mazidi, M.; McAlinden, C.; McGrath, J. J.; McKee, M.;
570 Mehrotra, R.; Mehta, K. M.; Mehta, V.; Meier, T.; Mekonnen, F. A.; Melaku, Y. A.; Melese, A.;
571 Melku, M.; Memiah, P. T. N.; Memish, Z. A.; Mendoza, W.; Mengistu, D. T.; Mensah, G. A.;
572 Mensink, G. B. M.; Mereta, S. T.; Meretoja, A.; Meretoja, T. J.; Mestrovic, T.; Mezgebe, H. B.;
573 Miazgowski, B.; Miazgowski, T.; Millear, A. I.; Miller, T. R.; Miller-Petrie, M. K.; Mini, G. K.;
574 Mirarefin, M.; Mirica, A.; Mirrakhimov, E. M.; Misganaw, A. T.; Mitiku, H.; Moazen, B.; Mohajer,
575 B.; Mohammad, K. A.; Mohammadi, M.; Mohammadifard, N.; Mohammadnia-Afrouzi, M.;
576 Mohammed, S.; Mohebi, F.; Mokdad, A. H.; Molokhia, M.; Momeniha, F.; Monasta, L.; Moodley,
577 Y.; Moradi, G.; Moradi-Lakeh, M.; Moradinazar, M.; Moraga, P.; Morawska, L.; Morgado-Da-
578 Costa, J.; Morrison, S. D.; Moschos, M. M.; Mouodi, S.; Mousavi, S. M.; Mozaffarian, D.; Mruts, K.
579 B.; Muche, A. A.; Muchie, K. F.; Mueller, U. O.; Muhammed, O. S.; Mukhopadhyay, S.; Muller, K.;
580 Musa, K. I.; Mustafa, G.; Nabhan, A. F.; Naghavi, M.; Naheed, A.; Nahvijou, A.; Naik, G.; Naik, N.;
581 Najafi, F.; Nangia, V.; Nansseu, J. R.; Nascimento, B. R.; Neal, B.; Neamati, N.; Negoi, I.; Negoi, R.
582 I.; Neupane, S.; Newton, C. R. J.; Ngunjiri, J. W.; Nguyen, A. Q.; Nguyen, G.; Nguyen, H. T.; Nguyen,
583 H. L. T.; Nguyen, H. T.; Nguyen, M.; Nguyen, N. B.; Nichols, E.; Nie, J.; Ningrum, D. N. A.; Nirayo, Y.
584 L.; Nishi, N.; Nixon, M. R.; Nojomi, M.; Nomura, S.; Norheim, O. F.; Noroozi, M.; Norrving, B.;
585 Noubiap, J. J.; Nouri, H. R.; Shiadeh, M. N.; Nowroozi, M. R.; Nsoesie, E. O.; Nyasulu, P. S.;
586 Obermeyer, C. M.; Odell, C. M.; Ofori-Asenso, R.; Ogbo, F. A.; Oh, I.-H.; Oladimeji, O.; Olagunju, A.
587 T.; Olagunju, T. O.; Olivares, P. R.; Olsen, H. E.; Olusanya, B. O.; Olusanya, J. O.; Ong, K. L.; Ong, S.
588 K.; Oren, E.; Orpana, H. M.; Ortiz, A.; Ota, E.; Ostavnov, S. S.; Øverland, S.; Owolabi, M. O.; A, M.
589 P.; Pacella, R.; Pakhare, A. P.; Pakpour, A. H.; Pana, A.; Panda-Jonas, S.; Park, E.-K.; Parry, C. D. H.;
590 Parsian, H.; Patel, S.; Pati, S.; Patil, S. T.; Patle, A.; Patton, G. C.; Paudel, D.; Paulson, K. R.;
591 Ballesteros, W. C. P.; Pearce, N.; Pereira, A.; Pereira, D. M.; Perico, N.; Pesudovs, K.; Petzold, M.;
592 Pham, H. Q.; Phillips, M. R.; Pillay, J. D.; Piradov, M. A.; Pirsahib, M.; Pisched, T.; Pishgar, F.;
593 Plana-Ripoll, O.; Plass, D.; Polinder, S.; Polkinghorne, K. R.; Postma, M. J.; Poulton, R.; Pourshams,
594 A.; Poustchi, H.; Prabhakaran, D.; Prakash, S.; Prasad, N.; Purcell, C. A.; Purwar, M. B.; Qorbani,
595 M.; Radfar, A.; Rafay, A.; Rafiei, A.; Rahim, F.; Rahimi, Z.; Rahimi-Movagh, A.; Rahimi-Movagh, V.;
596 Rahman, M.; Rahman, M. H.; Rahman, M. A.; Rai, R. K.; Rajati, F.; Rajsic, S.; Raju, S. B.; Ram,
597 U.; Ranabhat, C. L.; Ranjan, P.; Rath, G. K.; Rawaf, D. L.; Rawaf, S.; Reddy, K. S.; Rehm, C. D.; Rehm,
598 J.; Reiner, R. C.; Reitsma, M. B.; Remuzzi, G.; Renzaho, A. M. N.; Resnikoff, S.; Reynales-
599 Shigematsu, L. M.; Rezaei, S.; Ribeiro, A. L. P.; Rivera, J. A.; Roba, K. T.; Rodríguez-Ramírez, S.;
600 Roever, L.; Román, Y.; Ronfani, L.; Roshandel, G.; Rostami, A.; Roth, G. A.; Rothenbacher, D.; Roy,
601 A.; Rubagotti, E.; Rushton, L.; Sabanayagam, C.; Sachdev, P. S.; Saddik, B.; Sadeghi, E.;
602 Moghaddam, S. S.; Safari, H.; Safari, Y.; Safari-Faramani, R.; Safdarian, M.; Safi, S.; Safiri, S.; Sagar,
603 R.; Sahebkar, A.; Sahraian, M. A.; Sajadi, H. S.; Salam, N.; Salamat, P.; Saleem, Z.; Salimi, Y.;
604 Salimzadeh, H.; Salomon, J. A.; Salvi, D. D.; Salz, I.; Samy, A. M.; Sanabria, J.; Sanchez-Niño, M. D.;
605 Sánchez-Pimienta, T. G.; Sanders, T.; Sang, Y.; Santomauro, D. F.; Santos, I. S.; Santos, J. V.;
606 Milicevic, M. M. S.; Jose, B. P. S.; Sardana, M.; Sarker, A. R.; Sarmiento-Suárez, R.; Sarrafzadegan,
607 N.; Sartorius, B.; Sarvi, S.; Sathian, B.; Satpathy, M.; Sawant, A. R.; Sawhney, M.; Saylan, M.;
608 Sayyah, M.; Schaeffner, E.; Schmidt, M. I.; Schneider, I. J. C.; Schöttker, B.; Schutte, A. E.;
609 Schwebel, D. C.; Schwendicke, F.; Scott, J. G.; Seedat, S.; Sekerija, M.; Sepanlou, S. G.; Serre, M. L.;
610 Serván-Mori, E.; Seyedmousavi, S.; Shabaninejad, H.; Shaddick, G.; Shafeeasabet, A.; Shahbazi, M.;
611 Shaheen, A. A.; Shaikh, M. A.; Levy, T. S.; Shams-Beyranvand, M.; Shamsi, M.; Sharifi, H.; Sharifi,
612 K.; Sharif, M.; Sharif-Alhoseini, M.; Sharifi, H.; Sharma, J.; Sharma, M.; Sharma, R.; She, J.; Sheikh,
613 A.; Shi, P.; Shibuya, K.; Shiferaw, M. S.; Shigematsu, M.; Shin, M.-J.; Shiri, R.; Shirkoohi, R.; Shiue,
614 I.; Shokraneh, F.; Shoman, H.; Shrim, M. G.; Shupler, M. S.; Si, S.; Siabani, S.; Sibai, A. M.; Siddiqi,
615 T. J.; Sigfusdottir, I. D.; Sigurvinssdottir, R.; Silva, D. A. S.; Silva, J. P.; Silveira, D. G. A.; Singh, J. A.;
616 Singh, N. P.; Singh, V.; Sinha, D. N.; Skiadaresi, E.; Skirbekk, V.; Smith, D. L.; Smith, M.; Sobaih, B.

- 617 H.; Sobhani, S.; Somayaji, R.; Soofi, M.; Sorensen, R. J. D.; Soriano, J. B.; Soyiri, I. N.; Spinelli, A.;
618 Sposato, L. A.; Sreeramareddy, C. T.; Srinivasan, V.; Starodubov, V. I.; Steckling, N.; Stein, D. J.;
619 Stein, M. B.; Stevanovic, G.; Stockfelt, L.; Stokes, M. A.; Sturua, L.; Subart, M. L.; Sudaryanto, A.;
620 Sufiyan, M. B.; Sulo, G.; Sunguya, B. F.; Sur, P. J.; Sykes, B. L.; Szoek, C. E. I.; Tabarés-Seisdedos,
621 R.; Tabuchi, T.; Tadakamadla, S. K.; Takahashi, K.; Tandon, N.; Tassew, S. G.; Tavakkoli, M.;
622 Taveira, N.; Tehrani-Banihashemi, A.; Tekalign, T. G.; Tekelemedhin, S. W.; Tekle, M. G.;
623 Temesgen, H.; Temsah, M.-H.; Temsah, O.; Terkawi, A. S.; Tessema, B.; Teweldemedhin, M.;
624 Thankappan, K. R.; Theis, A.; Thirunavukkarasu, S.; Thomas, H. J.; Thomas, M. L.; Thomas, N.;
625 Thurston, G. D.; Tilahun, B.; Tillmann, T.; To, Q. G.; Tobollik, M.; Tonelli, M.; Topor-Madry, R.;
626 Torre, A. E.; Tortajada-Girbés, M.; Touvier, M.; Tovani-Palone, M. R.; Towbin, J. A.; Tran, B. X.;
627 Tran, K. B.; Truelsen, T. C.; Truong, N. T.; Tsadik, A. G.; Car, L. T.; Tuzcu, E. M.; Tymeson, H. D.;
628 Tyrovolas, S.; Ukwaja, K. N.; Ullah, I.; Updike, R. L.; Usman, M. S.; Uthman, O. A.; Vaduganathan,
629 M.; Vaezi, A.; Valdez, P. R.; Donkelaar, A. V.; Varavikova, E.; Varughese, S.; Vasankari, T. J.;
630 Venkateswaran, V.; Venketasubramanian, N.; Villafaina, S.; Violante, F. S.; Vladimirov, S. K.;
631 Vlassov, V.; Vollset, S. E.; Vos, T.; Vosoughi, K.; Vu, G. T.; Vujcic, I. S.; Wagnew, F. S.; Waheed, Y.;
632 Waller, S. G.; Walson, J. L.; Wang, Y.; Wang, Y.; Wang, Y.-P.; Weiderpass, E.; Weintraub, R. G.;
633 Weldegebreal, F.; Werdecker, A.; Werkneh, A. A.; West, J. J.; Westerman, R.; Whiteford, H. A.;
634 Widecka, J.; Wijeratne, T.; Winkler, A. S.; Wiyeh, A. B.; Wiysonge, C. S.; Wolfe, C. D. A.; Wong, T.
635 Y.; Wu, S.; Xavier, D.; Xu, G.; Yadgir, S.; Yadollahpour, A.; Jabbari, S. H. Y.; Yamada, T.; Yan, L. L.;
636 Yano, Y.; Yaseri, M.; Yasin, Y. J.; Yeshaneh, A.; Yimer, E. M.; Yip, P.; Yisma, E.; Yonemoto, N.; Yoon,
637 S.-J.; Yotebieng, M.; Younis, M. Z.; Yousefifard, M.; Yu, C.; Zaidi, Z.; Zaman, S. B.; Zamani, M.;
638 Zavala-Arciniega, L.; Zhang, A. L.; Zhang, H.; Zhang, K.; Zhou, M.; Zimsen, S. R. M.; Zodpey, S.;
639 Murray, C. J. L. Global, Regional, and National Comparative Risk Assessment of 84 Behavioural,
640 Environmental and Occupational, and Metabolic Risks or Clusters of Risks for 195 Countries and
641 Territories, 1990–2017: A Systematic Analysis for the Global Burden of Disease Study 2017. *The Lancet* **2018**, 392 (10159), 1923–1994. [https://doi.org/10.1016/S0140-6736\(18\)32225-6](https://doi.org/10.1016/S0140-6736(18)32225-6).
- 642 (23) Apte, J. S.; Marshall, J. D.; Cohen, A. J.; Brauer, M. Addressing Global Mortality from Ambient
643 PM2.5. *Environ. Sci. Technol.* **2015**, 49 (13), 8057–8066. <https://doi.org/10.1021/acs.est.5b01236>.
- 644 (24) Bekö, G.; Weschler, C. J.; Wierzbicka, A.; Karottki, D. G.; Toftum, J.; Loft, S.; Clausen, G. Ultrafine
645 Particles: Exposure and Source Apportionment in 56 Danish Homes. *Environ. Sci. Technol.* **2013**,
646 47 (18), 10240–10248.
- 647 (25) Azimi, P.; Stephens, B. A Framework for Estimating the US Mortality Burden of Fine Particulate
648 Matter Exposure Attributable to Indoor and Outdoor Microenvironments. *Journal of Exposure
649 Science & Environmental Epidemiology* **2018**, 30, 271–284.
- 650 (26) Fortmann, R.; Kariher, P.; Clayton, R. Indoor Air Quality: Residential Cooking Exposures. *State of
651 California Air Resources Board* **2001**, Contract 97-330.
- 652 (27) Ji, X.; Le Bihan, O.; Ramalho, O.; Mandin, C.; D'Anna, B.; Martinon, L.; Nicolas, M.; Bard, D.;
653 Pairon, J.-C. Characterization of Particles Emitted by Incense Burning in an Experimental House.
654 *Indoor Air* **2010**, 20 (2), 147–158. <https://doi.org/10.1111/j.1600-0668.2009.00634.x>.
- 655 (28) Manoukian, A.; Quivet, E.; Temime-Roussel, B.; Nicolas, M.; Maupetit, F.; Wortham, H. Emission
656 Characteristics of Air Pollutants from Incense and Candle Burning in Indoor Atmospheres. *Environ
657 Sci Pollut Res* **2013**, 20 (7), 4659–4670. <https://doi.org/10.1007/s11356-012-1394-y>.
- 658 (29) Rim, D.; Green, M.; Wallace, L.; Persily, A.; Choi, J.-I. Evolution of Ultrafine Particle Size
659 Distributions Following Indoor Episodic Releases: Relative Importance of Coagulation, Deposition
660 and Ventilation. *Aerosol Science and Technology* **2012**, 46 (5), 494–503.
661 <https://doi.org/10.1080/02786826.2011.639317>.

- 663 (30) Schripp, T.; Kirsch, I.; Salthammer, T. Characterization of Particle Emission from Household
664 Electrical Appliances. *Science of The Total Environment* **2011**, *409* (13), 2534–2540.
665 <https://doi.org/10.1016/j.scitotenv.2011.03.033>.
- 666 (31) Farmer, D. K.; Vance, M. E.; Abbatt, J. P. D.; Abeleira, A.; Alves, M. R.; Arata, C.; Boedicker, E.;
667 Bourne, S.; Cardoso-Saldaña, F.; Corsi, R.; DeCarlo, P. F.; Goldstein, A. H.; Grassian, V. H.; Ruiz, L.
668 H.; Jimenez, J. L.; Kahan, T. F.; Katz, E. F.; Mattila, J. M.; Nazaroff, W. W.; Novoselac, A.; O'Brien, R.
669 E.; Or, V. W.; Patel, S.; Sankhyan, S.; Stevens, P. S.; Tian, Y.; Wade, M.; Wang, C.; Zhou, S.; Zhou, Y.
670 Overview of HOMEChem: House Observations of Microbial and Environmental Chemistry.
671 *Environ. Sci.: Processes Impacts* **2019**, *21* (8), 1280–1300. <https://doi.org/10.1039/C9EM00228F>.
- 672 (32) Kangasluoma, J.; Franchin, A.; Duplissy, J.; Ahonen, L.; Korhonen, F.; Attoui, M.; Mikkilä, J.;
673 Lehtipalo, K.; Vanhanen, J.; Kulmala, M.; Petäjä, T. Operation of the Airmodus A11 Nano
674 Condensation Nucleus Counter at Various Inlet Pressures and Various Operation Temperatures,
675 and Design of a New Inlet System. *Atmos. Meas. Tech.* **2016**, *9* (7), 2977–2988.
676 <https://doi.org/10.5194/amt-9-2977-2016>.
- 677 (33) Vanhanen, J.; Mikkilä, J.; Lehtipalo, K.; Sipilä, M.; Manninen, H. E.; Siivola, E.; Petäjä, T.; Kulmala,
678 M. Particle Size Magnifier for Nano-CN Detection. *Aerosol Science and Technology* **2011**, *45* (4),
679 533–542. <https://doi.org/10.1080/02786826.2010.547889>.
- 680 (34) Lehtipalo, K.; Leppä, J.; Kontkanen, J.; Kangasluoma, J.; Franchin, A.; Wimmer, D.; Schobesberger,
681 S.; Junninen, H.; Petäjä, T.; Sipilä, M.; Mikkilä, J.; Vanhanen, J.; Worsnop, D. R.; Kulmala, M.
682 Methods for Determining Particle Size Distribution and Growth Rates between 1 and 3 Nm Using
683 the Particle Size Magnifier. *Boreal Environment Research* **2014**, *19* (suppl. B), 215–236.
- 684 (35) Ferro, A. R.; Kopperud, R. J.; Hildemann, L. M. Elevated Personal Exposure to Particulate Matter
685 from Human Activities in a Residence. *J Expo Sci Environ Epidemiol* **2004**, *14* (Suppl. 1), S34–S40.
686 <https://doi.org/10.1038/sj.jea.7500356>.
- 687 (36) Pitz, M.; Cyrys, J.; Karg, E.; Wiedensohler, A.; Wichmann, H.-E.; Heinrich, J. Variability of Apparent
688 Particle Density of an Urban Aerosol. *Environ. Sci. Technol.* **2003**, *37* (19), 4336–4342.
689 <https://doi.org/10.1021/es034322p>.
- 690 (37) Singer, B. C.; Delp, W. W. Response of Consumer and Research Grade Indoor Air Quality Monitors
691 to Residential Sources of Fine Particles. *Indoor Air* **2018**, *28* (4), 624–639.
692 <https://doi.org/10.1111/ina.12463>.
- 693 (38) Wang, Z.; Delp, W. W.; Singer, B. C. Performance of Low-Cost Indoor Air Quality Monitors for
694 PM2.5 and PM10 from Residential Sources. *Building and Environment* **2020**, *171*, 106654.
- 695 (39) Licina, D.; Bhangar, S.; Brooks, B.; Baker, R.; Firek, B.; Tang, X.; Morowitz, M. J.; Banfield, J. F.;
696 Nazaroff, W. W. Concentrations and Sources of Airborne Particles in a Neonatal Intensive Care
697 Unit. *PloS one* **2016**, *11* (5), e0154991.
- 698 (40) Chen, A.; Cao, Q.; Zhou, J.; Yang, B.; Chang, V. W.-C.; Nazaroff, W. W. Indoor and Outdoor
699 Particles in an Air-Conditioned Building during and after the 2013 Haze in Singapore. *Building and*
700 *Environment* **2016**, *99*, 73–81. <https://doi.org/10.1016/j.buildenv.2016.01.002>.
- 701 (41) Bair, W. J. The ICRP Human Respiratory Tract Model for Radiological Protection. *Radiat Prot
702 Dosimetry* **1995**, *60* (4), 307–310. <https://doi.org/10.1093/oxfordjournals.rpd.a082732>.
- 703 (42) Hinds, W. C. *Aerosol Technology: Properties, Behavior, and Measurement of Airborne Particles*,
704 2nd Edition.; John Wiley & Sons, New York, 1999.
- 705 (43) U.S. EPA. *Exposure Factors Handbook 2011 Edition (Final Report)*, EPA/600/R-09/052F.; U.S.
706 Environmental Protection Agency, Washington, DC, 2011; Vol. 1.
- 707 (44) Tiwari, S.; Srivastava, A. K.; Bisht, D. S.; Parmita, P.; Srivastava, M. K.; Attri, S. D. Diurnal and
708 Seasonal Variations of Black Carbon and PM2.5 over New Delhi, India: Influence of Meteorology.
709 *Atmospheric Research* **2013**, *125–126*, 50–62. <https://doi.org/10.1016/j.atmosres.2013.01.011>.

- 710 (45) Lv, B.; Zhang, B.; Bai, Y. A Systematic Analysis of PM_{2.5} in Beijing and Its Sources from
711 2000 to 2012. *Atmospheric Environment* **2016**, *124*, 98–108.
712 <https://doi.org/10.1016/j.atmosenv.2015.09.031>.
- 713 (46) Wan, M.-P.; Wu, C.-L.; Sze To, G.-N.; Chan, T.-C.; Chao, C. Y. H. Ultrafine Particles, and PM_{2.5}
714 Generated from Cooking in Homes. *Atmospheric Environment* **2011**, *45* (34), 6141–6148.
715 <https://doi.org/10.1016/j.atmosenv.2011.08.036>.
- 716 (47) Long, C. M.; Suh, H. H.; Koutrakis, P. Characterization of Indoor Particle Sources Using Continuous
717 Mass and Size Monitors. *Journal of the Air & Waste Management Association* **2000**, *50* (7), 1236–
718 1250. <https://doi.org/10.1080/10473289.2000.10464154>.
- 719 (48) Ferro, A. R.; Kopperud, R. J.; Hildemann, L. M. Source Strengths for Indoor Human Activities That
720 Resuspend Particulate Matter. *Environ. Sci. Technol.* **2004**, *38* (6), 1759–1764.
721 <https://doi.org/10.1021/es0263893>.
- 722 (49) Licina, D.; Tian, Y.; Nazaroff, W. W. Emission Rates and the Personal Cloud Effect Associated with
723 Particle Release from the Perihuman Environment. *Indoor Air* **2017**, *27* (4), 791–802.
724 <https://doi.org/10.1111/ina.12365>.
- 725 (50) Kulmala, M.; Kontkanen, J.; Junninen, H.; Lehtipalo, K.; Manninen, H. E.; Nieminen, T.; Petäjä, T.;
726 Sipilä, M.; Schobesberger, S.; Rantala, P.; Franchin, A.; Jokinen, T.; Järvinen, E.; Äijälä, M.;
727 Kangasluoma, J.; Hakala, J.; Aalto, P. P.; Paasonen, P.; Mikkilä, J.; Vanhanen, J.; Aalto, J.; Hakola,
728 H.; Makkonen, U.; Ruuskanen, T.; Mauldin, R. L.; Duplissy, J.; Vehkamäki, H.; Bäck, J.; Kortelainen,
729 A.; Riipinen, I.; Kurtén, T.; Johnston, M. V.; Smith, J. N.; Ehn, M.; Mentel, T. F.; Lehtinen, K. E. J.;
730 Laaksonen, A.; Kerminen, V.-M.; Worsnop, D. R. Direct Observations of Atmospheric Aerosol
731 Nucleation. *Science* **2013**, *339* (6122), 943–946. <https://doi.org/10.1126/science.1227385>.
- 732 (51) Rönkkö, T.; Kuuluvainen, H.; Karjalainen, P.; Keskinen, J.; Hillamo, R.; Niemi, J. V.; Pirjola, L.;
733 Timonen, H. J.; Saarikoski, S.; Saukko, E.; Järvinen, A.; Silvennoinen, H.; Rostedt, A.; Olin, M.; Yli-
734 Ojanperä, J.; Nousiainen, P.; Kousa, A.; Maso, M. D. Traffic Is a Major Source of Atmospheric
735 Nanocluster Aerosol. *PNAS* **2017**, *114* (29), 7549–7554.
736 <https://doi.org/10.1073/pnas.1700830114>.
- 737 (52) Wallace, L. Indoor Sources of Ultrafine and Accumulation Mode Particles: Size Distributions, Size-
738 Resolved Concentrations, and Source Strengths. *Aerosol Science and Technology* **2006**, *40* (5),
739 348–360. <https://doi.org/10.1080/02786820600612250>.
- 740 (53) Wallace, L. A.; Emmerich, S. J.; Howard-Reed, C. Source Strengths of Ultrafine and Fine Particles
741 Due to Cooking with a Gas Stove. *Environ. Sci. Technol.* **2004**, *38* (8), 2304–2311.
742 <https://doi.org/10.1021/es0306260>.
- 743 (54) He, C.; Morawska, L.; Hitchins, J.; Gilbert, D. Contribution from Indoor Sources to Particle Number
744 and Mass Concentrations in Residential Houses. *Atmospheric Environment* **2004**, *38* (21), 3405–
745 3415. <https://doi.org/10.1016/j.atmosenv.2004.03.027>.
- 746 (55) Wallace, L. A.; Ott, W. R.; Weschler, C. J. Ultrafine Particles from Electric Appliances and Cooking
747 Pans: Experiments Suggesting Desorption/Nucleation of Sorbed Organics as the Primary Source.
748 *Indoor Air* **2015**, *25* (5), 536–546. <https://doi.org/10.1111/ina.12163>.
- 749 (56) Wallace, L. A.; Ott, W. R.; Weschler, C. J.; Lai, A. C. K. Desorption of SVOCs from Heated Surfaces
750 in the Form of Ultrafine Particles. *Environ. Sci. Technol.* **2017**, *51* (3), 1140–1146.
751 <https://doi.org/10.1021/acs.est.6b03248>.
- 752 (57) Odabasi, M. Halogenated Volatile Organic Compounds from the Use of Chlorine-Bleach-
753 Containing Household Products. *Environ. Sci. Technol.* **2008**, *42* (5), 1445–1451.
754 <https://doi.org/10.1021/es702355u>.
- 755 (58) Shepherd, J. L.; Corsi, R. L.; Kemp, J. Chloroform in Indoor Air and Wastewater: The Role of
756 Residential Washing Machines. *Journal of the Air & Waste Management Association* **1996**, *46* (7),
757 631–642. <https://doi.org/10.1080/10473289.1996.10467497>.

- 758 (59) Popolan-Vaida, D. M.; Liu, C.-L.; Nah, T.; Wilson, K. R.; Leone, S. R. Reaction of Chlorine Molecules
759 with Unsaturated Submicron Organic Particles. *Zeitschrift für Physikalische Chemie* **2015**, 229
760 (10–12), 1521–1540. <https://doi.org/10.1515/zpch-2015-0662>.
- 761 (60) Wang, C.; Collins, D. B.; Abbatt, J. P. D. Indoor Illumination of Terpenes and Bleach Emissions
762 Leads to Particle Formation and Growth. *Environ. Sci. Technol.* **2019**, 53 (20), 11792–11800.
763 <https://doi.org/10.1021/acs.est.9b04261>.
- 764