

UC Santa Cruz

UC Santa Cruz Previously Published Works

Title

Bayesian Sparse Multivariate Regression with Asymmetric Nonlocal Priors for Microbiome Data Analysis

Permalink

<https://escholarship.org/uc/item/8n4426xd>

Journal

Bayesian Analysis, 15(2)

ISSN

1936-0975

Authors

Shuler, Kurtis
Sison-Mangus, Marilou
Lee, Juhee

Publication Date

2020-06-01

DOI

10.1214/19-ba1164

Peer reviewed

1 **The mercury-tolerant microbiota of the zooplankton *Daphnia* aids in**
2 **host survival and maintains fecundity under mercury stress**

3

4 Jiunn C. Fong, Brandon E. De Guzman, Carl H. Lamborg and Marilou P. Sison-
5 Mangus*

6

7 Department of Ocean Sciences and Institute for Marine Sciences, University
8 of California Santa Cruz, Santa Cruz, California 95064, USA.

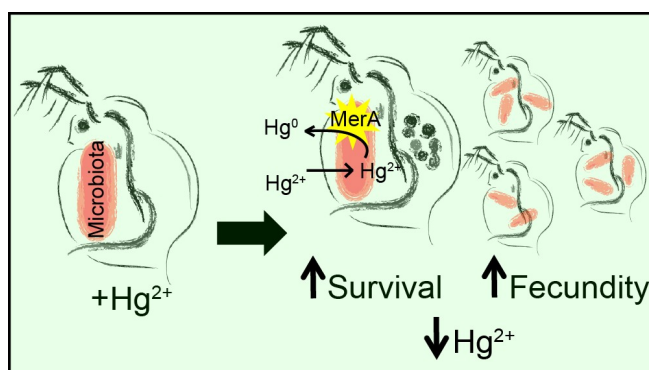
9

10 *Corresponding author: Marilou P. Sison-Mangus; E-mail:

11 msisonma@ucsc.edu

12

13 Key words: Mercury biotransformation, *Daphnia*, microbiome, *merA*, water
14 fleas, mercury stress, Yolo Bypass, symbiosis



15

16

TOC Art

17 **ABSTRACT**

18 Many aquatic organisms can thrive in polluted environments by having the
19 genetic capability to withstand sub-optimal conditions. However, the
20 contributions of microbiomes under these stressful environments are poorly
21 understood. We investigated whether a mercury-tolerant microbiota can
22 extend its phenotype to its host by ameliorating host survival and fecundity
23 under mercury-stress. We isolated microbiota members from various clones
24 of *Daphnia magna*, screened for the mercury-biotransforming *merA* gene,
25 and determined their mercury tolerance levels. We then introduced the
26 mercury-tolerant microbiota, *Pseudomonas-10*, to axenic *D. magna* and
27 quantified its *merA* gene expression, mercury reduction capability and
28 measured its impact on host survival and fecundity. Expression of *merA* gene
29 was up-regulated in *Pseudomonas-10*, both in isolation and in host-
30 association with mercury exposure. *Pseudomonas-10* is also capable of
31 significantly reducing mercury concentration in the medium. Notably,
32 mercury-exposed daphnids containing only *Pseudomonas-10* exhibited
33 higher survival and fecundity than mercury-exposed daphnids supplemented
34 with parental microbiome. Our study showed that zooplankton, such as
35 *Daphnia*, naturally harbor microbiome members that are eco-responsive and
36 tolerant to mercury exposure and can aid in host survival and maintain host
37 fecundity in a mercury-contaminated environment. This study further
38 demonstrates that under stressful environmental condition, the fitness of the
39 host can depend on the genotype and the phenotype of its microbiome.

40

41 **INTRODUCTION**

42 *Daphnia* is a ubiquitous keystone zooplankton species found in many
43 aquatic ecosystems, capable of growing in both pristine and polluted
44 environments. It can withstand many environmental stressors, hence, is
45 widely used as a model organism for numerous research (1-7), including
46 studies on the effects of temperature fluctuations (8, 9), food availability (10-
47 12), predations (13-15), and exposure to metal pollutants (such as mercury,
48 cadmium, copper, arsenic) (16-23). To date, mercury studies on *Daphnia*
49 have focused mainly on the mechanisms of uptake, accumulation and
50 elimination in *D. magna* (16, 24-26). While host responses to stressors are
51 evaluated primarily, the responses and contributions of their microbiomes
52 are often overlooked. Aside from host genetics, microbiomes can also be an
53 important component in host survival in deleterious environments. For
54 instance, the gut microbiome of desert woodrats enhances the host's ability
55 to ingest plants with toxic secondary compounds (27), and the gut
56 microbiota of a coffee berry borer can detoxify the caffeine ingested by its
57 host, allowing the beetle to subsist on the otherwise toxic caffeine-laden
58 berries (28).

59 Mercury is a well-known metal contaminant that bioaccumulates and
60 biomagnifies in aquatic food webs (29-31). Many studies on mercury
61 pollution and its neurotoxicity have been reported (30, 32-34). Three major

62 species of mercury can be found in the environment: inorganic mercury
63 (Hg^{2+}), organic methylmercury (MeHg including the mono- and dimethyl
64 forms), and elemental mercury (Hg^0). Hg^0 and Hg^{2+} are commonly released
65 into the atmosphere via anthropogenic and natural sources, while deposition
66 of atmospheric mercury leads to mercury contamination of terrestrial and
67 aquatic ecosystems (31, 33, 35). Anaerobic microorganisms harboring *hgcAB*
68 genes (encoding a corrinoid-dependent protein and an associated ferredoxin
69 protein, respectively) have been shown to methylate mercury, producing the
70 highly bioaccumulative MeHg from inorganic Hg^{2+} (36-39). Demethylation of
71 MeHg to Hg^{2+} , on the other hand, can be facilitated by microorganisms
72 containing *merB* that encodes an organomercurial lyase. Inorganic Hg^{2+} can
73 be further converted into the less toxic, extremely volatile Hg^0 by
74 microorganisms expressing the *merA* gene that encodes a mercury
75 reductase (40-43). Therefore, microorganisms harboring these mercury-
76 biotransformation genes have the potential to impact mercury speciation,
77 cycling and concentration in the environment. Although mercury tolerance
78 and detoxification in many bacteria species have been reported (42-44),
79 most of these bacteria were identified in water columns, sediments and soil
80 (45-47). In addition, most studies on biotransformation of mercury in natural
81 ecosystems focused on bacteria found in specific geographical
82 environmental habitats (45-47), with only a few studies on mercury-
83 biotransforming bacteria found in animal hosts (48, 49). The contribution of

84 these bacteria on host mercury tolerance and survival, however, has never
85 been investigated.

86 In this study, we investigated the role of microbiomes on host fitness
87 under mercury stress. We first assessed the microbiome community
88 structure of *D. magna* collected from a seasonally mercury-polluted site (Yolo
89 Bypass, California) (50). We then isolated members of the microbiota,
90 determined their mercury tolerance levels, and measured *merA* gene
91 expression and mercury reduction of the mercury-tolerant microbiota
92 member (*Pseudomonas*-10) in isolation and in association with the daphnid
93 host. We found that *Daphnia* is an environmental reservoir of mercury-
94 tolerant bacteria that could potentially biotransform mercury into less-toxic
95 form by up-regulating the expression of *merA* gene upon exposure to
96 elevated levels of mercury and reducing the concentration of mercury in the
97 medium. Most importantly, we found that a daphnid microbiota member can
98 aid in the host survival and allow the host to produce viable offspring even
99 when exposed to mercury contamination.

100

101 **MATERIALS AND METHODS**

102 ***Daphnia* animal collection and husbandry.** *D. magna* CAY (**C**alifornia
103 **Y**olo-bypass) was sampled in the inlet ponds (38°31'45.4"N, 121°36'28.9"W)
104 that are part of the Yolo Bypass Wildlife Area (Davis, California, U.S.A). Other
105 *D. magna* clones, DE-K35-linb1 (Germany) and FI-Xinb3 (Finland) (51), are
106 part of the *Daphnia* collection in our laboratory. Daphnid husbandry was

107 carried out as previously published (52). Animals were routinely cultured in
108 artificial *Daphnia* medium (ADaM) (53) at 20°C with 14:10 light:dark
109 photoperiod and fed daily with *Scenedesmus obliquus* (~2x10⁷ cells).
110 Experiments with *D. magna* were also carried in the same media and
111 conditions.

112

113 **16S rRNA fragments high-throughput sequencing and data analysis.**

114 Total genomic DNA (gDNA) ($n = 3$) was extracted using DNeasy Blood and
115 Tissue Kit (Qiagen, Germantown, MD). Equimolar DNA concentration was
116 sent to Argonne National Laboratory Sequencing Core for library preparation
117 and sequencing of the 16S rRNA V4 region (PCR-amplified with 515F and
118 806R primers, Table S1), using MiSeq Illumina sequencing platform.
119 Additional details are provided under Supporting Information B.

120

121 *D. magna* sequences were demultiplexed and checked for chimeras,
122 with low quality sequence and short sequence reads (<150 bp) removed
123 prior to post-analysis with the Quantitative Insights Into Microbial Ecology
124 (QIIME 1.8) pipeline (54). Operational taxonomic units (OTUs) were clustered
125 at 97% sequence similarity with Uclust (55). Representative sequences from
126 the clustered OTUs were picked for taxonomic identification based on RDP
127 classifier 2.2 (56). Singletons, chloroplast and mitochondria sequences were
128 filtered out of the OTU table prior to alignment of OTU representative
129 sequences with PyNAST (54). Samples were rarified at a minimum of 10
130 sequences and a maximum depth of 12,030 sequences in steps of 10. For

131 microbiome composition analysis, assignment of bacterial taxon were
132 performed using BLAST 2.22.2 (57). Information on NCBI data availability of
133 the raw sequence reads are provided in Supporting Information E.

134

135 **Bacteria isolation and identification, growth conditions, and *merA***
136 **screening.** *D. magna* microbiota were isolated from various *D. magna*
137 clones: the newly collected CAY clone and clones in the laboratory collections
138 (FI-Xinb3 and DE-K35-linb1), as well as sediment samples from Yolo Bypass.
139 Individual *D. magna* were washed three times with 1 ml of sterile ADaM,
140 homogenized in 200 μ l of sterile ADaM, plated on various agar media,
141 including LB, R2A (Teknova, Hollister, CA), 10x-diluted R2A with and without
142 N-acetylglucosamine supplement, and MacConkey (Teknova, Hollister, CA),
143 and incubated at room temperature (22-23°C) for several days. Sediment
144 samples were also plated on the same set of agar media and incubated at
145 room temperature for several days. Agar medium contains 1.5% (w/v)
146 granulated agar. Colonies exhibiting different phenotypes were repeatedly
147 streaked for single colonies. Pure microbiota isolates were cultured in R2A
148 liquid media and stored at -80°C in autoclaved glycerol.

149 Genetic identification of bacterial isolates and *merA* screening were
150 carried out by sequencing the partial PCR-amplified 16S rRNA and *merA*
151 fragments. PCR amplification was carried out with MyTaq Red (Bioline,
152 Taunton, MA) using the following PCR program: initial denaturation at 95°C
153 for 2 min, followed by 35 cycles of 95°C for 30 sec, 50°C for 30 sec, 72°C for

154 2 min and a final extension step at 72°C for 5 min. Universal 16S primer set
155 8F and 1492R (58) and *merA* primer sets MERA5 and MERA1 (44) or merA-
156 128_F and merA-993_R, were used to amplify 16S rRNA and *merA* genes,
157 respectively. Additional details are provided in Supporting Information C.
158 Primers used in this study are listed in Table S1. GenBank accession
159 numbers for the partial 16S rRNA and *merA* sequences are found in the
160 Supporting Information. Details on 16S rRNA phylogenetic analysis is
161 provided in Supporting Information D.

162

163 **Bacteria mercury minimal inhibitory concentration (MIC) assays.**

164 Mercury MIC assays of twenty-seven bacterial isolates were carried out at
165 room temperature (22-23°C) in 96-well microtiter plates containing 200 μ l of
166 R2A media with different concentrations of mercury (0.1, 0.2, 0.4, 0.6, 0.8
167 μ M, and 1 μ M to 15 μ M, with 1 μ M increment). The bacteria MIC was
168 determined using 1 μ M to 15 μ M first, followed by the lower mercury
169 concentrations (0.1, 0.2, 0.4, 0.6, 0.8 μ M and 1 μ M) for bacteria that have
170 MIC < 1 μ M. Each well containing the defined mercury concentrations was
171 inoculated with 2 μ l of bacterial inoculum and visible bacterial growth,
172 defined as increased in culture turbidity, was visually checked and measured
173 as increased in absorbance/optical density at 600 nm (OD₆₀₀). Cultures grown
174 in the absence of mercury were used as positive controls, while un-
175 inoculated R2A media were used as negative controls. The MIC is the lowest
176 mercury concentration where bacterial growth (culture turbidity) was not

177 observed with the naked eye (59), and the difference of OD₆₀₀ values
178 between measured and negative control (un-inoculated R2A media) is less
179 than 0.01 (60). The MIC assays were repeated three times and the lowest
180 mercury concentration where *no growth* occurred (59) after 3 days of
181 incubation indicates the mercury MIC for that particular bacterial isolate.
182 Mercury stock solution (1,000 µg/ml mercury in 10% nitric acid) was
183 purchased from SPEX CertiPrep (Metuchen, NJ). OD₆₀₀ was determined daily
184 using SpectraMax M2 microplate reader (Molecular Devices, Sunnyvale, CA).

185

186 **RNA extraction and cDNA synthesis.** Since *Pseudomonas*-10 exhibited
187 the highest mercury MIC (8 µM) among the *merA*-positive microbiota isolates
188 identified in this study (Table 1), this microbiota isolate was used to study
189 *merA* gene expression. To quantify *merA* gene expression in isolation,
190 *Pseudomonas*-10 was grown aerobically at room temperature (22-23°C) in LB
191 media until mid-exponential growth phase (OD₆₀₀ = 0.3). Aliquots (10 ml) of
192 the cultures were transferred to new culture flasks, followed by no mercury
193 exposure (0 µM) or exposure to 2.5 µM or 5 µM of mercury for 15 min at room
194 temperature. After treatments, 2 ml-aliquots were pelleted by centrifugation
195 (21, 200 x *g*) at room temperature for 2 min. Cell pellets were immediately
196 resuspended in Trizol (Fisher Scientific, Hampton, NH) and stored at -80°C.

197 For measuring *merA* expression of daphnid-associated *Pseudomonas*-
198 10, *D. magna* CAY harboring *Pseudomonas*-10 exposed to 50 nM mercury for
199 4 h on Day 5 (Experiment 1, *n* = 55; Experiment 2, *n* = 50) or unexposed

200 (Experiment 1, $n = 50$; Experiment 2, $n = 53$) were harvested, washed once
201 with sterile ADaM, homogenized in Trizol and stored at -80°C . The mercury
202 LD_{50} for *D. magna* is 51.5 nM (Figure S3), hence, the mercury concentration
203 of 50 nM was chosen for the mercury stress experiments in *D. magna* CAY.
204 Gnotobiotic *D. magna* were generated as described for survival assays (52).

205 RNA was extracted with Direct-zol RNA Miniprep Plus (Zymo research,
206 Irvine, CA) according to manufacturer's instructions. Eluted total RNA was
207 further treated with DNase I (Promega, Madison, WI) for 30 min at 37°C ,
208 followed by inactivation at 65°C for 10 min, and purified using Direct-zol RNA
209 Miniprep Plus. RNA concentrations were estimated using either NanoDrop
210 spectrophotometer or Qubit Fluorometer (Fisher Scientific, Hampton, NH).
211 cDNA was synthesized using SuperScript III and random hexamers
212 (Invitrogen, Carlsbad, CA), according to manufacturer's instructions. For no-
213 host RNA samples, 200 ng of RNA was used as templates for cDNA synthesis.
214 For daphnid-associated RNA samples, depletion of daphnid host RNA was
215 first carried out using Dynabeads mRNA purification Kit (Fisher Scientific,
216 Hampton, NH), where the Dynabeads Oligo (dT)₂₅ binds to host RNA that
217 contains poly-A tail. Unbound bacterial RNA was collected and purified using
218 Direct-zol RNA Miniprep Plus. cDNA synthesis was then carried out using 2-3
219 μg of RNA as templates. Reactions without reverse transcriptase were used
220 as controls (NRT).

221

222 **Quantitative real-time PCR (qRT-PCR) and semi-quantitative PCR**
223 **(semi-qPCR).** qRT-PCR reactions (10 μ l) were set up using PowerUp SYBR
224 Green Master Mix (Applied Biosystems, Foster City, CA), according to
225 manufacturer's instructions. Each reaction contains 2 μ l of cDNA, 0.4 μ M of
226 each specific primer (merA-Pse10_F and merA-Pse10_R or glnA_F and
227 glnA_R; Table S1) and 1X SYBR green master mix. qRT-PCR was performed
228 on an Applied Biosystems 7500 real-time PCR system with the following
229 thermal cycling steps: 50°C for 2 min, 95°C for 2 min, followed by 40 cycles of
230 95°C for 15 sec and 60°C for 1 min. Specificity of the *merA* and *glnA* primers
231 were tested with regular PCR and DNA gel electrophoresis, as well as in qRT-
232 PCR with a Dissociation Stage (melt curve analysis). No amplification of no
233 template control (NTC) and no reverse transcriptase (NRT) reactions served
234 as negative controls. Serially-diluted *Pseudomonas-10* gDNA samples,
235 extracted with DNeasy Blood and Tissue Kit (Qiagen, Germantown, MD),
236 were used as positive controls and for generating calibration curves. The C_T
237 values of the samples were within the linear dynamic range of the calibration
238 curves for *glnA* and *merA*, with R^2 values of >0.99 . All samples were
239 normalized to the expression of the glutamine synthetase (*glnA*)
240 housekeeping gene (61), and relative gene expression (fold change) is
241 calculated using $2^{-\Delta\Delta C_T}$ method (62). Data were log-transformed for statistical
242 analysis. The assay was carried out three times, each with three technical
243 replicates.

244 Semi-qPCR was carried out, similar to a protocol published previously
245 (63), to measure *merA* expression in daphnid-associated *Pseudomonas*-10.
246 The reactions (10 μ l) were set up using MyTaq Red master mix, 3 μ l of cDNA
247 and 1 μ M of each specific primer (*merA*-Pse10_F and *merA*-Pse10_R or *glnA*_F
248 and *glnA*_R) with the following PCR program: initial denaturation at 95°C for
249 2 min, followed by 35 cycles of 95°C for 15 sec, 55°C for 20 sec, 72°C for 1
250 min and a final extension step at 72°C for 5 min. NRT and NTC were used as
251 negative controls. Amplified products (3 μ l) were analyzed with 2% (w/v)
252 agarose gel electrophoresis and the gel images were captured using
253 ChemiDoc System (BioRad, Hercules, California). Fluorescence were
254 determined with Image Lab v5 software (Bio-Rad) and calculated by
255 subtracting background fluorescence of the corresponding NRT reactions for
256 both *merA* and *glnA*, followed by normalization to the corresponding *glnA*
257 samples. The assay was carried out twice, each with three technical
258 replicates. Statistical methods for comparison between treatments are
259 described in Supporting Information D.

260

261 ***D. magna* mercury lethal dose 50 (LD₅₀), survival and fecundity**
262 **assays.** Mercury LD₅₀ of *D. magna* was determined in non-axenic daphnids
263 cultured in 80 ml of ADaM medium containing 10, 50, 75 and 100 nM
264 mercury. Daphnids cultured in ADaM without mercury (0 μ M mercury) served
265 as controls. The mercury LD₅₀ experiments were carried out in 10% nitric
266 acid-washed experimental jars (with 0.22- μ M filter caps) containing 4

267 daphnids per experimental jar ($n = 4$ for each treatment) and were fed with
268 non-axenic *S. obliquus* ($\sim 2 \times 10^7$ cells per jar). LD₅₀ was determined using 4-
269 parameter logistic regression (Figure S3). The LD₅₀ of *D. magna* CAY after 2
270 days of mercury exposure is 51.5 nM, and as such 50 nM of mercury was
271 used as the applied concentration in the experiments that involved mercury
272 exposure in *D. magna* CAY. The LD₅₀ assays were carried out with four
273 replicates.

274 Survival assays were carried out twice in 10% nitric acid-washed
275 experimental jars (with 0.22- μ M filter caps) containing 80 ml of sterile ADaM
276 with one daphnid per jar and fed with axenic *S. obliquus* ($\sim 8 \times 10^7$ cells/jar).
277 Bacteria-free eggs were generated with antibiotic-treatment (52) using the
278 third egg clutches of reproductively synchronized F2 adults and were
279 separated into 3 groups: Bacteria-free (Bac-Free), bacteria supplemented
280 with parental microbiome (Bac-Suppl) and *Pseudomonas*-10-infected (Pse-
281 Inf). *Pseudomonas*-10 was used as it exhibits the highest MIC (8 μ M) among
282 the *merA*-positive microbiota isolates identified in this study (Table 1).
283 Mercury stress (50 nM) was introduced at Day 5. Survival were monitored
284 daily for a period of 18 days. Fecundity assays were carried out twice, set up
285 similarly as the survival assays, in 80 ml of sterile ADaM with one daphnid
286 per jar and fed with axenic *S. obliquus* ($\sim 8 \times 10^7$ cells/jar). Number of live
287 hatchlings was counted and removed from the experimental jars daily for a
288 period of 18 days. Mercury stress (50 nM) was added on Day 5. At the end of
289 the experiments, sterility of the daphnids in Bac-Free group and the

290 presence of bacteria in Bac-Suppl and Pse-Inf groups were verified by PCR
291 using 327F and 936R primer set (Table S1), targeting bacterial 16S rRNA
292 gene (52). In addition, crushed daphnids from each experimental group were
293 also plated on LB or R2A agar media (Sigma, St. Louis, MO) to verify that
294 there was no bacterial growth from the Bac-Free group, mixed bacterial
295 growth from Bac-Suppl group (indicated by the growth of bacteria exhibiting
296 various morphotypes), and pure bacterial growth from Pse-Inf group. The
297 identity of *Pseudomonas-10* from the Pse-Inf groups was further confirmed
298 via sequencing of the 16S rRNA using 8F and 1492R (58).

299

300 **Mercury biotransformation by *Pseudomonas-10*.** Total mercury (Hg^{2+}
301 and Hg^0 , MeHg and other Hg species) concentrations in the ADaM medium of
302 *Pseudomonas-10* only (Pse), bacteria-free daphnids (Bac-Free) and daphnids
303 infected with *Pseudomonas-10* (Pse-Inf) were determined by cold vapor
304 atomic absorption spectrometry using a DMA-80 (Milestone, Inc.), calibrated
305 with aqueous standards and accuracy checked against standard reference
306 material BCR-414 (EPA Standard Method 7473) (64). Experimental jars were
307 set up similar to the survival assays, except that each experimental jar
308 contains 5 daphnids. Experimental jars containing ADaM medium (80 ml)
309 only served as controls. Experimental jars of the Pse group contain ADaM
310 medium (80 ml) and *Pseudomonas-10* (200 μl of ADaM-washed bacterial
311 culture suspension diluted to $\text{OD} = \sim 0.6$). Mercury (50 nM) was added to the
312 experimental jars on Day 5. Samples were collected and filtered through

313 0.22 μm filter on Day 5, before (No Hg) and after (D5) addition of mercury,
314 and on Day 8 (D8). Mercury samples were stored at 4°C until analysis. Total
315 mercury in the samples were measured directly with cold vapor atomic
316 absorption spectrometry without any pretreatment. Data were obtained from
317 at least 3 replicates for each condition: ADaM medium ($n = 3$), Pse ($n = 3$),
318 Bac-Free ($n = 4$), Pse-Inf ($n = 4$).

319

320 **Statistical Analysis.** Data were presented as mean values with standard
321 errors, except for box plots where medians (horizontal lines within the box)
322 were shown with 25% to 75% quartiles (whiskers). For experiments
323 measuring the differences in gene expression and mercury reduction
324 between treatments, data were log-transformed (when necessary) to fit the
325 assumption of normal distribution and were tested for homogeneity of
326 variance prior to performing one-way ANOVA with Tukey's HSD post-hoc test
327 or Student's t test using JMP 14. Statistical analysis for differences in survival
328 rates was conducted using Mantel Cox log-rank test in GraphPad Prism 5.01,
329 with Bonferroni corrected p -value for multiple comparisons. Kruskal-Wallis
330 test and Wilcoxon each pair test was used to compare differences in
331 fecundity using JMP 14.0.

332

333 **RESULTS AND DISCUSSION**

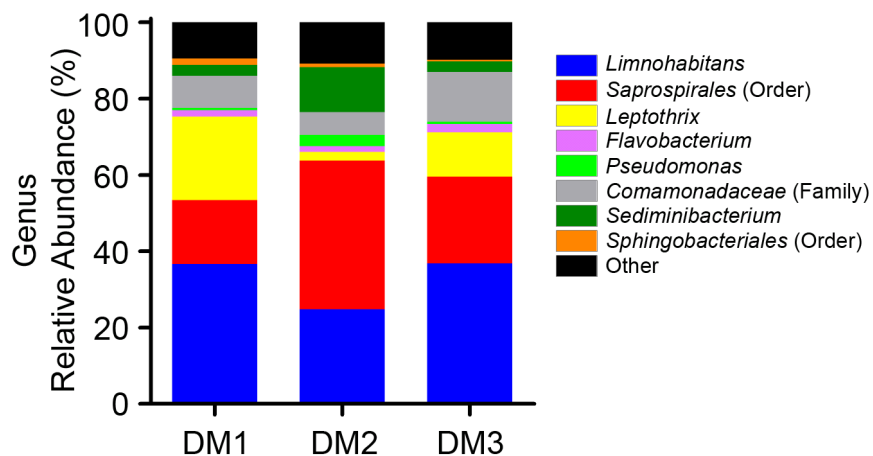
334 **Microbiome composition of *D. magna* CAY**

335 To investigate whether *D. magna* CAY (collected from **California Yolo**
336 Bypass) harbors mercury-tolerant and *merA*-positive bacteria, we first
337 determined the composition of the *D. magna* CAY microbiome via 16S rRNA
338 amplicon sequencing. A total of 1,295 OTUs were identified from the samples
339 ($n = 3$). After removing singletons, mitochondrial and chloroplast sequences,
340 416 OTUs remained and were assigned into 73 genera, 32 families, 10
341 orders, 9 classes and 1 phylum (Table S2). Only bacteria with $\geq 1\%$ average
342 relative abundance (from 3 samples) were included in the comparison shown
343 in Figure 1. The top three most abundant microbiota in *D. magna* CAY are
344 bacteria from the genus *Limnohabitans* (32.8%, average abundance), the
345 order *Saprospirales* (26.2%), and the genus *Leptothrix* (11.9%). Other
346 identified microbiota members that are $\geq 1\%$ (average abundance) include
347 bacteria from the family *Comamonadaceae* (9.2%), the genera
348 *Sediminibacterium* (5.8%), *Flavobacterium* (1.8%), *Pseudomonas* (1.3%), and
349 the order *Sphingobacteriales* (1%). Since *Pseudomonas* has been shown to
350 be pathogenic to *D. magna* (65-67), it is interesting that *Pseudomonas* was
351 found to be $>1\%$ in the microbiota of *D. magna* CAY. The presence of
352 *Pseudomonas* as part of *Daphnia* microbiome has also been reported in
353 different *Daphnia* species (51, 68, 69), suggesting that these *Pseudomonas*
354 microbiome members may be non-pathogenic, as opposed to the isolates (*P.*
355 *aeruginosa* PAO1 and strain DD1) used in the reported studies.

356 Notably, the microbial composition and structure of *D. magna* CAY
357 reported here is similar with the microbiomes of other *D. magna* coming

358 from different geographical locations (51, 68, 70), such as the dominance of
 359 *Limnohabitans* and Saprospirales group for instance, suggesting active
 360 selection of microbiome by the daphnid host. *Limnohabitans* have been
 361 reported to increase fecundity in the *Daphnia* host (71) and are abundant in
 362 host's filter apparatus (72). Indeed, host genetics have been shown to play a
 363 role in structuring and maintaining different microbiota community
 364 abundances in *Daphnia* (51) and in *Drosophila* (73-75). Collectively, our data
 365 indicates that the newly collected *D. magna* CAY clones harbor microbiota
 366 similar to other published *D. magna* clones and reinforces the idea that host
 367 factors influence the microbial composition of the *Daphnia* microbiome.

368



369

370 **Figure 1. Relative abundance of major taxonomic groups of**

371 **microbiota from *D. magna* CAY.** Relative abundance of microbiota

372 members that are $\geq 1\%$ (average abundance of all 3 samples). Genera with

373 less than 1% are grouped as "Other". Some of the microbiota were identified

374 only at the family and order levels using BLAST. *D. magna* samples ($n = 3$):

375 DM1, DM2 and DM3.

376

377 **Mercury minimal inhibitory concentration (MIC) and *merA* screening**
378 **of microbiota isolates**

379 Many bacteria are capable of mercury detoxification, commonly
380 through the expression of *mer* operons (42-44). One of the central players of
381 this widespread mercury-detoxification system is MerA, a mercuric reductase
382 that is capable of reducing a more reactive, cationic form of mercury (Hg^{2+})
383 to a relative inert, volatile monoatomic mercury vapor (Hg^0), which can
384 diffuse through the cell membrane (40-42). Cross-referencing the 73
385 assigned genera found in *D. magna* CAY with NCBI databases, 41 genera
386 were found to contain species that putatively harbor *merA* genes (Table S2).
387 We therefore hypothesize that *Daphnia* could harbor *merA*-positive bacteria
388 capable of detoxifying mercury through biotransformation. To test this
389 hypothesis, we isolated pure cultures of the microbiota members from the
390 newly collected *D. magna* CAY and two published *D. magna* clones (DE-K35-
391 linb1 and FI-Xinb3) in the laboratory collection as well as their environments
392 (ADaM culture media and the sediment from Yolo Bypass sampling site). We
393 then determined the mercury MICs of individual isolated bacteria and
394 screened them for the presence of *merA* gene. We isolated 27 bacteria from
395 different *D. magna* clones and their environments (Table 1) and determined
396 their identities and phylogeny using partial 16S rRNA sequences. The well-
397 resolved phylogenetic tree showed the phylum/class groupings of *D. magna*
398 CAY microbiota isolates (Figure 2). Several of the bacterial isolates

399 (*Acidovorax*, *Acinetobacter*, *Blastomonas*, *Exiguobacterium*,
400 *Hydrogenophaga*, and *Pseudomonas*) have homologous 16s rRNA sequences
401 identified in the *D. magna* CAY microbiome (Figure 1 and Table S2). In
402 addition, some of these isolated bacteria (*Acidovorax*, *Acinetobacter*,
403 *Aeromonas*, *Hydrogenophaga* and *Pseudomonas*) were also identified as
404 microbiome members of other *Daphnia* species (68, 69, 76). We were not
405 able to isolate and culture the dominant microbiota member, *Limnohabitans*,
406 despite numerous attempts using several different media and methods of
407 isolation.

408 Table 1 also shows the level of mercury tolerance (reported as MIC) of
409 the isolated *D. magna* microbiota. Highest mercury MIC was observed with
410 *Acidovorax*-4 (11 μ M), followed by *Curtobacterium*-20 (11 μ M) and
411 *Pseudomonas*-10 (8 μ M). Other microbiota isolates exhibited low to medium
412 MICs ranging from 0.6 μ M to 6 μ M. To date, most bacteria tested for mercury
413 tolerance are either free-living bacteria or bacteria found in the sediments
414 (45, 77-79). Host-associated bacteria that are mercury-tolerant have been
415 reported only in a few hosts, including fish gut bacteria grown in media
416 containing 12.5 μ M of mercury (48), bacteria isolated from nodules of
417 leguminous plants and marine sponges exhibiting MICs of 30 μ M and >100
418 μ M of mercury, respectively (49, 80). Hence, our study and that of others
419 indicate that host-associated microbiomes are also potential sources of
420 mercury-detoxifying bacteria.

421 To determine if mercury-tolerant bacteria contain *merA* genes, we
422 screened all 27 bacteria isolates for the presence of *merA* using published
423 primer sets (Table S1) (44, 77, 81). Initial screening identified partial *merA*
424 genes in *Acinetobacter-1* and *Hydrogenophaga-8* using primer set MERA5
425 and MERA1 (44); other primer sets did not yield amplified *merA* fragments.
426 Using the partial *merA* sequences from *Hydrogenophaga-8* and
427 *Acinetobacter-1*, we designed a primer set, merA-128_F and merA-993_R
428 (Table S1), targeting the conserved sequence regions (Figure S1) and further
429 identified three more bacteria isolates containing *merA*. In all, *merA*
430 fragments were amplified from five microbiota isolates: three Gamma-
431 proteobacteria isolates (*Pseudomonas-10*, *Pseudomonas-23*, and
432 *Acinetobacter-1*) and two Beta-proteobacteria isolates (*Variovorax-11* and
433 *Hydrogenophaga-8*). Homology sequence searches using BLAST showed high
434 sequence similarity to known *merA* sequences (Table S3 and Figure S1),
435 indicating that *Pseudomonas-10*, *Pseudomonas-23*, *Acinetobacter-1*,
436 *Variovorax-11*, and *Hydrogenophaga-8* harbor *merA* genes. It is noteworthy
437 that all these five *merA*-positive isolates also exhibited moderate to high
438 mercury MICs, ranging from 3 to 8 μ M (Table 1), hinting that these bacteria
439 isolates likely contain functional *merA* genes. The ability of primer set merA-
440 128_F and merA-993_R in amplifying *merA* fragments from both Beta-
441 proteobacteria and Gamma-proteobacteria, indicates that it can be used for
442 *merA* screening of environmental samples.

443 Although many other isolated microbiota exhibited high mercury
444 tolerance or belong to genera that putatively contain *merA*-positive species,
445 we were unable to amplify their *merA* gene, including *Acidovorax*-4 and
446 *Curtobacterium*-20. In addition, while *Pseudomonas*-10 and -23 harbor the
447 *merA* gene, we were unable to detect the presence of *merA* in the
448 congeneric *Pseudomonas*-28. It is possible that these bacteria isolates
449 contain nucleotide polymorphisms in the binding regions of the designed
450 *merA* primer sets, or that these particular species/strains do not contain
451 *merA* but exhibit mercury tolerance by other yet-to-be identified
452 mechanisms. We are continuing our effort in using new *merA* primer sets to
453 screen our remaining microbiota isolates, as well as screening for the
454 presence of *merB*, and will be part of another study.

455 MerA exists as a homodimer protein and three major conserved
456 regions/residues critical for MerA activity have been identified (42). At the C-
457 terminus, a stretch of residues containing two cysteines forms a redox-active
458 site, while the N-terminus short cysteine pair aids in Hg²⁺ binding to the C-
459 terminal redox-active site of the corresponding monomer. In addition, two
460 tyrosine residues are also involved in Hg²⁺ binding (42). While most of the
461 conserved active regions/residues are beyond the identified *merA* sequences
462 of the microbiota isolates, the first Hg²⁺-binding tyrosine is found in
463 *Pseudomonas*-10, *Hydrogenophaga*-8 and *Acinetobacter*-1 (Figure S1). The
464 well-resolved MerA tree also showed divergence of MerA between
465 *Pseudomonas* isolates and the other three microbiota isolates (Figure S1).

466 This is not surprising since *merA* found in *Pseudomonas* has been suggested
 467 to be more distantly related to the *merA* of other Gram-negative bacteria
 468 (82). While the partial MerA sequences between *Pseudomonas*-10 and -23
 469 determined in this study are identical, they exhibited distinct mercury
 470 tolerance levels. Possible sequence variations may exist in their C- and N-
 471 terminus active sites, which may explain the tolerance differences between
 472 these two *Pseudomonas* isolates. Together, our data showed that *Daphnia*
 473 microbiomes contain mercury-tolerant bacteria that harbor *merA* genes,
 474 which imply of their potential to biotransform mercury from the toxic Hg²⁺ to
 475 less toxic Hg⁰ species.

476

477 **Table 1.** List of identified *D. magna* microbiota isolates, mercury MIC and
 478 *merA* gene screening.

16S Identification ^a	rRNA Source ^b	Mercury MIC ^c (μM)	<i>merA</i> _d
<i>Acidovorax</i> -4	FI-Xinb3	11	
<i>Curtobacterium</i> - 20	CAY	11	
<i>Pseudomonas</i> -10	DE-K35-linb1	8	+
<i>Acinetobacter</i> -1	DE-K35-linb1 (ADaM)	6	+
<i>Brevundimonas</i> -6	DE-K35-linb1	6	
<i>Pseudomonas</i> -28	CAY	6	
<i>Hydrogenophaga</i> - 8	DE-K35-linb1	5	+
<i>Variovorax</i> -11	FI-Xinb3	5	+
<i>Bacillus</i> -19	CAY	5	
<i>Rhodococcus</i> -24	CAY	5	
<i>Blastomonas</i> -12	CAY	4	
<i>Runella</i> -13	CAY	4	
<i>Bacillus</i> -16	CAY	4	
<i>Bacillus</i> -5	FI-Xinb3	3	
<i>Micrococcus</i> -17	CAY	3	
<i>Pseudomonas</i> -23	CAY	3	+

<i>Shinella</i> -26	CAY	3
<i>Aeromonas</i> -27	CAY	3
<i>Arthrobacter</i> -2	DE-K35-linb1	2
<i>Aeromonas</i> -3	FI-Xinb3	2
<i>Citricoccus</i> -14	Sediment, Yolo Bypass	2
<i>Flaviumibacter</i> -21	CAY	2
<i>Curtobacterium</i> -15	CAY	1
<i>Flaviumibacter</i> -25	CAY	1
<i>Exiguobacterium</i> -7	DE-K35-linb1	0.8
<i>Flaviumibacter</i> -22	CAY	0.8
<i>Lysinibacillus</i> -18	Sediment, Yolo Bypass	0.6

480 ^a Numbers after the genus names indicated the laboratory collection
481 numbers, so as to distinguish between isolates from the same genera.

482 ^b Bacteria were isolated either from *D. magna* crushed body (CAY, DE-K35-
483 linb1 or FI-Xinb3 clones), Yolo Bypass sediment (where the *D. magna* CAY
484 clones were collected) or the ADaM culture medium.

485 ^c Lowest mercury concentration with no bacterial growth after 3 days of
486 incubation at room temperature (22-23°C).

487 ^d Presence of *merA* gene determined by PCR and sequencing, indicated with
488 + and shaded.

489

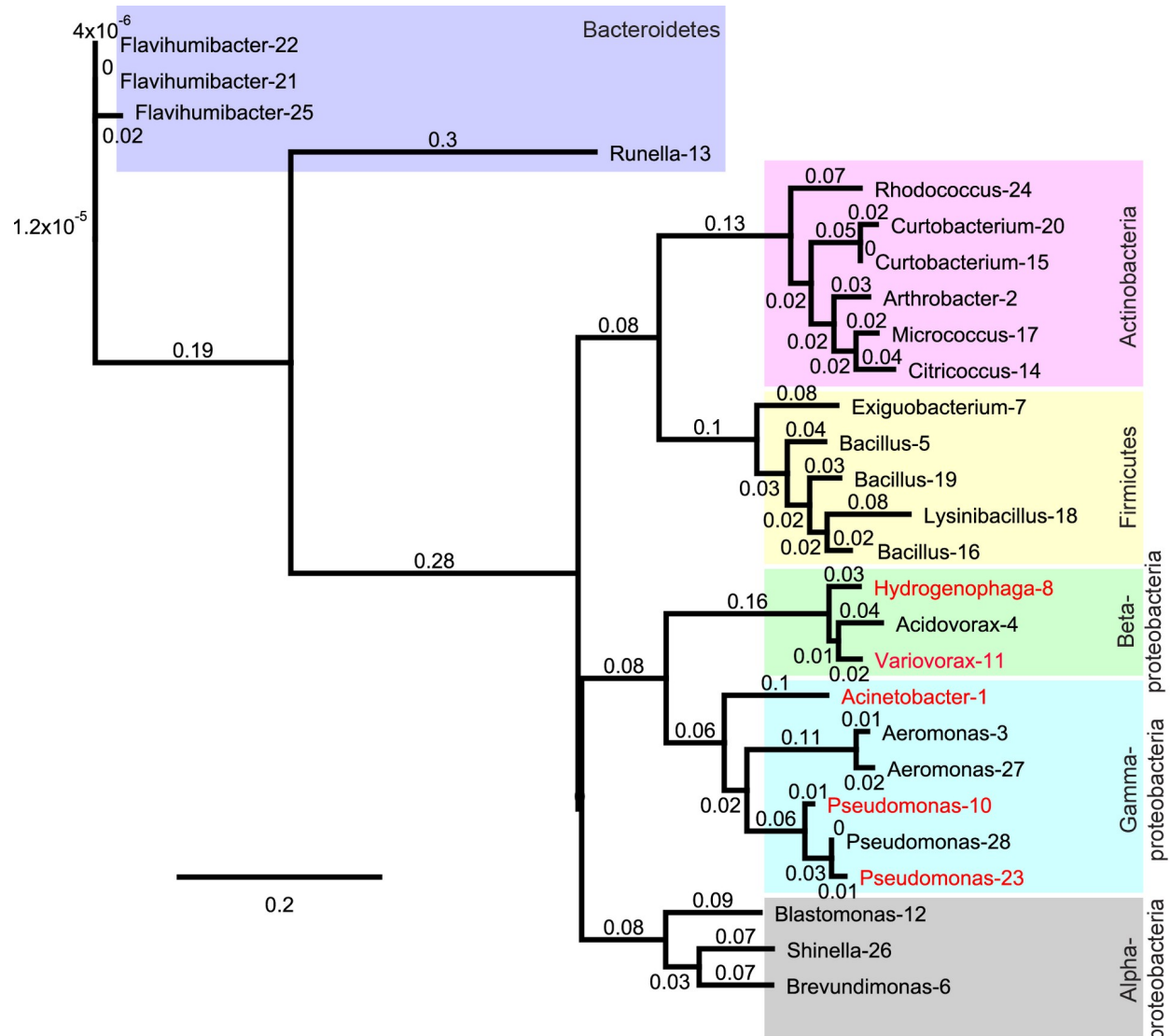
490

491

492

493

494



495
 496 **Figure 2. 16S rRNA phylogenetic tree of isolated *D. magna***
 497 **microbiota.** 16S rRNA sequences of the 27 isolates were first aligned using
 498 MUSCLE, followed by tree-construction using maximum likelihood method
 499 (PhyML), with Jukes-Cantor substitution model and 250 bootstrap replicates.
 500 *merA*-positive microbiota isolates identified in this study (Table 1) are in red.
 501 Numbers after the bacteria name represent laboratory collection ID number.
 502 Bootstrap values and scale bar are indicated as substitution per site.
 503

504 **Expression of bacterial *merA* gene in isolation and in association**
505 **with *D. magna* after mercury exposure**

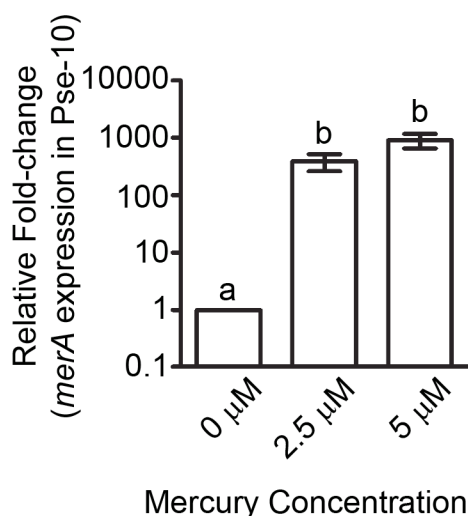
506 To test the functionality of the *merA* gene, we exposed exponentially
507 growing *Pseudomonas*-10 to 2.5 μM and 5 μM of mercury for 15 min and
508 compared *merA* expression between mercury-exposed and unexposed
509 cultures (0 μM). *Pseudomonas*-10 was chosen as it exhibits the highest
510 mercury MIC (8 μM) among the *merA*-positive microbiota isolates identified in
511 this study (Table 1). Using qRT-PCR, relative expression of *merA* was found
512 to be more than 300-fold higher in cells exposed to mercury compared to
513 unexposed cells (one-way ANOVA, $F = 130.79$, $p = 0.00001$ Figure 3). We
514 also visualized the qRT-PCR amplified products with agarose gel
515 electrophoresis and quantify the relative fluorescence (Figure S2). As
516 expected, higher fluorescence was detected in samples exposed to mercury
517 compared to the unexposed samples. Expression of *merA* is not significantly
518 different between cells exposed to 2.5 μM and 5 μM mercury (Tukey's HSD
519 Post-hoc test, $p > 0.05$, Figure 3 and Figure S2). Similarly, increases in *merA*
520 gene expressions have also been reported in other *Pseudomonas* strains,
521 where over 10-fold and 30-fold increases were observed with 2 μM and 5 μM
522 mercury exposure (respectively) in *Pseudomonas stutzeri* OX (83), and more
523 than 2-fold increase with 50 μM of mercury exposure in *Pseudomonas* strain
524 ATH-43 (78). Our data show that the mercury-tolerant and *merA*-positive
525 *Pseudomonas*-10 isolate is capable of upregulating *merA* gene expression
526 upon mercury exposure.

527 We further investigated if similar upregulation of *merA* also occurs in
528 *Pseudomonas*-10 when in association with the *D. magna* host upon exposure
529 to mercury. We first determined the lethal dose 50 (LD₅₀) of mercury in *D.*
530 *magna* CAY. To do this, we exposed 5-day-old non-axenic daphnids to
531 various concentration of mercury for 2 days and monitored their survival.
532 Using 4-parameter logistic regression, we determined that the LD₅₀ of
533 mercury is 51.5 nM (Figure S3) and as such, 50 nM was chosen as the
534 applied mercury concentration for the mercury stress experiments in *D.*
535 *magna*. We infected bacteria-free *D. magna* CAY with *Pseudomonas*-10,
536 allowed bacteria-host association to establish for 5 days, exposed the
537 *Pseudomonas*-10-infected daphnids to 50 nM of mercury for 4 hours,
538 harvested total RNA and determined *merA* gene expression. Several
539 attempts using qRT-PCR to determine *merA* expression in host-associated
540 *Pseudomonas*-10 did not yield reliable results, likely due to low abundance of
541 bacterial cDNA and interference from the daphnid host cDNA, despite
542 attempts to deplete host RNA before cDNA synthesis. As such, we carried out
543 semi-quantitative RT-PCR (semi-qPCR) by analyzing the amplified *merA* and
544 *glnA* (housekeeping gene for normalization) on agarose gel and determined
545 band intensities of the amplified products (Figure 4). Upon exposure to 50
546 nM of mercury, *merA* expression in daphnid-associated *Pseudomonas*-10 was
547 upregulated approximately 31-fold, when compared to daphnid-associated
548 *Pseudomonas*-10 without mercury treatment. This data indicates that
549 *Pseudomonas*-10 is capable of upregulating *merA* expression in response to

550 mercury stress even when in association with the daphnid host. The lower
551 expression level of *merA* in daphnid-associated *Pseudomonas*-10 (31-fold),
552 when compared to *Pseudomonas*-10 in isolation (>300-fold), is likely due to
553 the lower mercury concentration used in host-associated *Pseudomonas*-10
554 exposure.

555

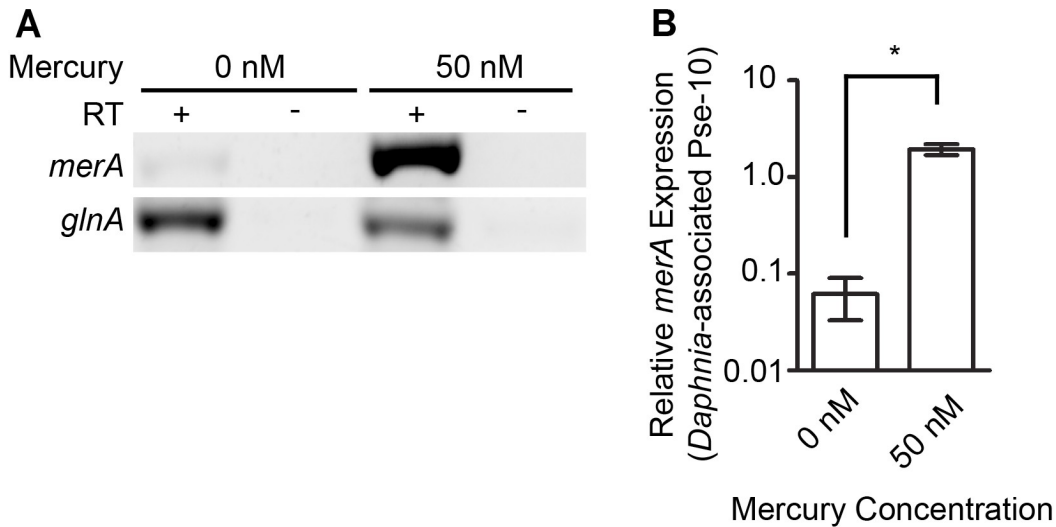
556



557

558 **Figure 3. *merA* expression in the microbiota isolate *Pseudomonas*-**
559 **10.** Relative fold-change of *merA* expression in *Pseudomonas*-10 (Pse-10)
560 bacterial cultures exposed to 0, 2.5 and 5 μ M of mercury in LB media,
561 determined from 3 independent experiments with 3 technical replicates.
562 Data were log-transformed and was analyzed using one-way ANOVA with
563 Tukey's HSD Post-hoc test. Column with different letters are significantly
564 different ($p < 0.05$).

565



566

567 **Figure 4. Expression of *merA* in the daphnid-associated**

568 ***Pseudomonas-10*. (A)** Agarose gel electrophoresis of amplified *merA* and

569 *glnA* (housekeeping gene for normalization). **(B)** Relative *merA* expression in

570 host-associated *Pseudomonas-10* exposed to 50 nM mercury, determined

571 from 2 independent experiments with 3 technical replicates. *, $p = 0.0001$

572 (Student's *t* test). RT, reverse transcriptase.

573

574 ***Pseudomonas-10* aids in host survival and maintains host fecundity**

575 **under mercury stress.**

576 Since host-associated *Pseudomonas-10* exhibited increased *merA*

577 expression upon mercury exposure, we investigated whether *Pseudomonas-*

578 *10* can contribute to host survival and fecundity under mercury stress. We

579 infected *D. magna* CAY with *Pseudomonas-10* (Pse-Inf), allowed the daphnids

580 to grow for 5 days before exposing the daphnids to 50 nM mercury, and

581 compared their survival rates to bacteria-free (Bac-Free) and parental-

582 microbiome supplemented *D. magna* CAY (Bac-Suppl) over a period of 18

583 days (Figure 5A). We positively verified the absence of bacteria in Bac-Free
584 group, the presence of *Pseudomonas* in Pse-Inf group and the presence of
585 various bacterial morphotypes in Bac-Suppl group at the end of the
586 experiment.

587 A significant difference in survival rates was observed between
588 daphnid groups (Mantel-Cox log-rank test, $\chi^2 = 26.29$, $p = 0.003$). In the
589 absence of mercury stress, Bac-Suppl daphnids (94%, $n = 29/31$) exhibited
590 similar survival rates with Pse-Inf daphnids (87%, $n = 26/30$, $\chi^2 = 0.68$, $p =$
591 0.41) and significantly higher survival rates than Bac-Free daphnids (73%, n
592 = 22/30, $\chi^2 = 4.16$, $p = 0.04$) at the end of the assay at Day 18. This
593 suggests that *Pseudomonas*-10 is a mutualistic symbiont and contributes to
594 *D. magna* survival. Similar report also showed that a *Pseudomonas* strain
595 isolated from a *D. magna* host is beneficial to the development of *D. magna*
596 resting eggs at warm conditions (84).

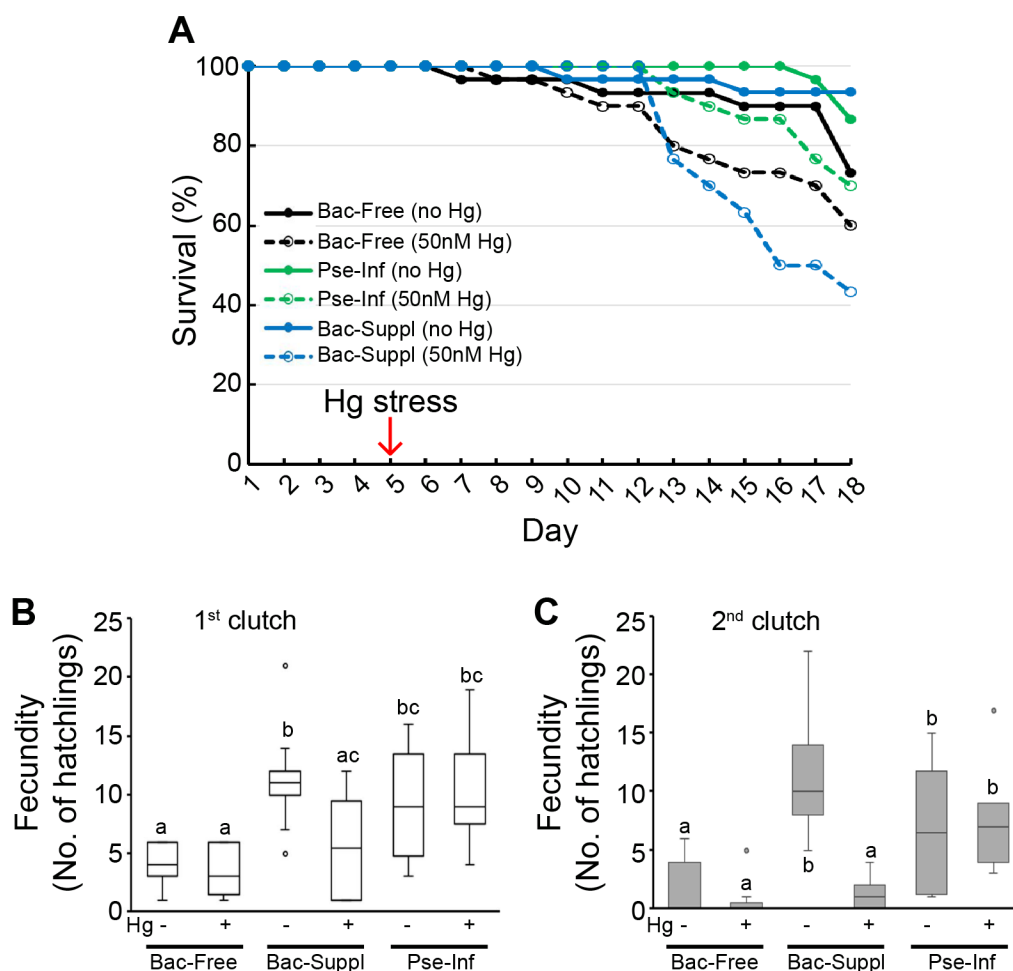
597 Under mercury stress, lower survival rates were observed for the three
598 mercury-treated groups when compared to their counterparts at the end of
599 the assay (Day 18): Bac-Suppl + Hg (43%, $n = 13/30$), Pse-Inf + Hg (70%, n
600 = 21/30) and Bac-Free + Hg (60%, $n = 18/30$). Daphnids from the Bac-Suppl
601 + Hg were most severely affected under mercury stress ($\chi^2 = 17.22$, $p <$
602 0.0001) while the survival rates of Pse-Inf + Hg daphnids were statistically
603 similar with Pse-Inf daphnids ($\chi^2 = 2.18$, $p = 0.14$). Survival rates of Bac-Free
604 daphnids were also statistically similar under mercury and without mercury
605 stress ($\chi^2 = 1.61$, $p = 0.21$). Notably, under mercury stress, daphnids

606 harboring *Pseudomonas*-10 (Pse-Inf + Hg) exhibited significantly higher
607 survival rates than Bac-Suppl + Hg ($\chi^2 = 6.19$, $p = 0.01$). These results
608 suggest that the mercury-tolerant *Pseudomonas*-10 is capable of
609 augmenting host survival under mercury stress, likely by biotransforming
610 mercury into less toxic form, in addition to the daphnid's ability to eliminate
611 mercury through excretion, egestion, molting, and neonate production (25).
612 Intriguingly, under mercury stress, daphnids supplemented with parental
613 microbiome (Bac-Suppl + Hg) exhibited the lowest survival (43% at Day 18
614 $n=13/30$). Similar results were obtained in a repeated experiment (Figure
615 S4). Possibly, the microbiota composition of the Bac-Suppl daphnids is low in
616 *merA*-positive bacteria, hence the low survival under mercury exposure, as
617 compared to the higher abundance of *merA*-positive *Pseudomonas*-10 in Pse-
618 Inf daphnids. It is also likely that under mercury stress, harboring various
619 kinds of microbiota can have a detrimental effect on the health of the
620 daphnids. Indeed, Bac-Suppl + Hg daphnids also exhibited significantly lower
621 survival rates than BacFree + Hg daphnids ($\chi^2 = 14.49$, $p = 0.0001$), implying
622 that harboring many bacteria is beneficial under normal conditions but can
623 be a burden under stressful environmental conditions. This is similarly
624 observed on grasses with and without fungal endophytes, where the
625 endophyte-free plants significantly fared better than endophytic-laden plants
626 in terms of root:shoot ratio under extreme limiting-nutrient conditions (85).

627 To further investigate the contributions of *Pseudomonas*-10 in host
628 fitness, we carried out fecundity assays and compare the number of live

629 hatchlings from first and second clutches of all daphnid groups with and
630 without mercury stress (Figure 5B, C and Figure S5). The six groups
631 significantly vary in the number of first and second clutch hatchlings
632 (Kruskal-Wallis test, $\chi^2 = 26.97$ and 37.56 , respectively, $p < 0.0001$),
633 suggesting differences between treatments. Under no mercury stress, Bac-
634 Suppl daphnids have significantly higher number of first and second clutch
635 hatchlings than Bac-Free daphnids (Wilcoxon Each Pair test, $p = 0.0004$ and
636 $p = 0.0003$, respectively), but did not differ with Pse-Inf daphnids ($p = 0.31$
637 and $p = 0.15$, respectively). This data confirms that *Pseudomonas-10* is a
638 beneficial symbiont and positively contributes to daphnid host fitness under
639 normal environmental condition. Intriguingly, untreated Bac-Suppl and Pse-
640 inf daphnids produced similar number of first and second clutch hatchlings
641 as the Pse-Inf + Hg treated daphnids ($p > 0.05$), suggesting that the
642 fecundity of the daphnids harboring mercury-tolerant *Pseudomonas-10* is
643 maintained under mercury stress, and is comparable to that of unstressed
644 daphnids. On the other hand, under mercury stress, hatchlings production of
645 Bac-Suppl + Hg treated daphnids, were significantly lower than Bac-Suppl
646 daphnids ($p = 0.007$, Figure 5A) and this decrease in fecundity is even more
647 pronounced in the second clutch ($p < 0.0001$, Figure 5C). Moreover, Bac-
648 Suppl + Hg have similar number of hatchlings with Bac-Free and Bac-Free +
649 Hg daphnids ($p > 0.05$). This indicates that prolonged exposure to mercury
650 results in lower fecundity of daphnids harboring several microbiota
651 associates. Interestingly, Bac-Suppl + Hg treated daphnids have

652 significantly lower number of second clutch hatchlings than Pse-Inf + Hg
 653 daphnids ($p = 0.0011$). This further confirms that mercury-tolerant
 654 *Pseudomonas*-10 can maintain the fecundity of the daphnid host under
 655 mercury exposure, likely by up-regulating its *merA* expression and reducing
 656 the toxicity of mercury, thereby allowing the host to withstand and survive,
 657 and maintain clonal reproduction in mercury-contaminated environment.



658
 659 **Figure 5. Survival and fecundity of gnotobiotic *D. magna* CAY**
 660 **harboring different microbiota under mercury stress. (A)** Survival of
 661 bacteria-free daphnids (Bac-Free), bacteria-supplemented daphnids
 662 harboring parental microbiota (Bac-Suppl), and *Pseudomonas*-10 infected

663 daphnids (Pse-Inf) with (50 nM Hg) and without (no Hg) mercury exposure.
664 Mercury stress (Hg stress) was introduced at Day 5, indicated with a red
665 arrow. Bac-Free (no Hg), $n = 30$; Bac-Free (50 nM Hg), $n = 30$; Bac-Suppl (no
666 Hg), $n = 31$; Bac-Suppl (50 nM Hg), $n = 30$; Pse-Inf (no Hg), $n = 30$; Pse-Inf
667 (50 nM Hg), $n = 30$. Survival assays were repeated twice (Figure S4), but only
668 one representative experiment is shown here. **(B)** Fecundity of Bac-Free,
669 Bac-Suppl and Pse-Inf daphnids with (50 nM Hg) and without (no Hg) mercury
670 stress. Boxes show the 25% to 75% quartiles, medians are shown as
671 horizontal lines (within the box), and maximum and minimum values are
672 shown as whiskers. Columns with the same letter are not significantly
673 different, analyzed using Kruskal-Wallis test with Wilcoxon each pair test for
674 pairwise comparisons. Bac-Free (no Hg), $n = 8$; Bac-Free (50 nM Hg), $n = 9$;
675 Bac-Suppl (no Hg), $n = 15$; Bac-Suppl (50 nM Hg), $n = 10$; Pse-Inf (no Hg), $n =$
676 9 ; Pse-Inf (50 nM Hg), $n = 9$. Fecundity assays were repeated twice (Figure
677 S5).

678

679 ***Pseudomonas-10* is capable of biotransforming mercury**

680 Since the *merA*-positive and mercury-tolerant *Pseudomonas-10* can up-
681 regulate *merA* expression, both in isolation and in host-association, and is
682 capable of aiding host survival and fecundity under mercury stress, we
683 wanted to investigate its ability to biotransform and reduce mercury (Figure
684 6). In a separate experiment, we measured the reduction or loss of mercury
685 by *Pseudomonas-10* in isolation (Pse) and in association with the daphnids

686 (Pse-Inf) and compared with the Bac-Free daphnids to tease apart the
687 contribution of each respective partner to mercury reduction. We did not
688 include Bac-Suppl in this study because it harbors several microbiota of
689 unknown mercury biotransformation potentials. As expected, before the
690 addition of mercury on Day 5 (No Hg), negligible background concentrations
691 of mercury were detected. Upon addition of 50nM mercury on Day 5 (D5), we
692 detected initial mercury concentrations ranging from 11.6 to 15.6 nM in all
693 groups (ADaM control, Pse, Bac-Free and Pse-Inf). The significant mercury
694 loss after initial mercury addition could be attributed to the attachment of
695 mercury to the walls of the processing vessels. Nonetheless, the
696 concentration of mercury in ADaM only control group between D5 and D8
697 showed no significant differences (Tukey's HSD, $p > 0.05$), as well as in D5
698 samples of all treatments (Tukey's HSD, $p > 0.05$), indicating that the loss of
699 mercury due to attachment to processing vessels is consistent among
700 samples. Intriguingly, after 3 days (Day 8), a drastic 4.7-fold decrease in
701 mercury concentrations (from 15.5 nM to 3.3 nM) was detected in the Pse
702 group, which contained only ADaM medium and *Pseudomonas*-10 (Tukey's
703 HSD, $p < 0.0001$). Mercury loss in the ADaM only control group between D5
704 and D8 is not significant, indicating that the reduction of mercury from
705 solution in Pse group was due to the presence of *Pseudomonas*-10. This
706 strongly suggests that *Pseudomonas*-10 biotransformed and reduced
707 mercury in the solution, most likely with the use of its MerA enzyme, which is
708 known to be the common mechanism for the reduction and

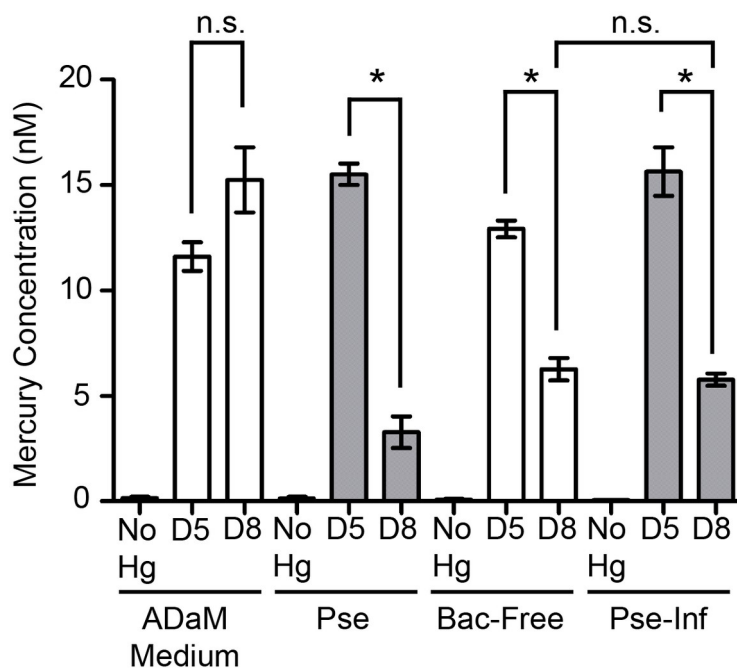
709 biotransformation of Hg²⁺ to elemental Hg⁰ (42). We also observed significant
710 reduction of mercury in the Bac-Free group after 3 days (12.9 nM to 6.3 nM,
711 2.0-fold), suggesting that the daphnids are capable of mercury uptake from
712 their environment. This is not surprising as assimilation of mercury by
713 *Daphnia* has been reported (25). Similar reduction of mercury was also
714 detected in the Pse-Inf group after 3 days (15.6 nM to 5.8 nM, 2.7-fold), but
715 mercury reduction was not significantly different from that of Bac-Free
716 group. Hence, we could not conclusively show that *Pseudomonas*-10
717 significantly contributes to mercury reduction when it is in association with
718 the daphnid host, despite its mercury biotransformation capability in
719 isolation.

720 Host genetics indeed play a major role on metal stress tolerance via
721 metallothionein genes, which act as metal-responsive systems upon metal
722 exposure. *D. magna* contain three putative genes encoding the metal-
723 binding metallothionein proteins in their genome (86, 87) and these proteins
724 are known to play an important role in metal detoxification and homeostasis
725 in many organisms, likely through binding and sequestration of various
726 metals, including mercury (88, 89). While the expression of these
727 metallothionein genes may enable the *Daphnia* host to tolerate mercury
728 stress, we showed in this study that the *merA*-positive and mercury-tolerant
729 microbiota *Pseudomonas*-10 can also reduce mercury in isolation and
730 contributes to host survival and fecundity under mercury stress, through
731 upregulation of *merA* and thus likely reducing the mercury stress

732 experienced by the host. To definitively show the contribution of
 733 *Pseudomonas*-10 in biotransforming mercury inside the host, and to further
 734 investigate its mechanistic role during host association, daphnids containing
 735 knock-outs of the host metallothionein genes and a *Pseudomonas*-10 strain
 736 harboring a *merA*-knockout, needs to be generated and will be part of
 737 another study.

738

739



740

741

742 **Figure 6. Mercury biotransformation by *Pseudomonas*-10.** Mercury
 743 concentrations (nM of Hg²⁺) measured in the ADaM medium control, with
 744 *Pseudomonas*-10 only (Pse), bacterial-free daphnids (Bac-Free) and daphnids
 745 infected with *Pseudomonas*-10 (Pse-Inf) on Day 5, before (No Hg) and after
 746 (D5) addition of mercury, and on Day 8 (D8). Experimental jars containing

747 ADaM medium only served as controls. * indicates significant difference ($p <$
748 0.0001), analyzed using one-way ANOVA with Tukey's HSD Post-hoc test,
749 while n.s. indicates no significant difference ($p > 0.05$).

750

751 Our study highlights one of the mechanisms on how an aquatic
752 organism like *Daphnia* can potentially withstand and survive environmental
753 stresses. Just like many hosts, *Daphnia* harbor several groups of bacteria
754 (with different relative abundances) in their microbiome consortium, most of
755 them with unknown functions; some microbial members maybe mutualistic,
756 commensals or parasitic depending on existing environmental conditions, or
757 the microbiome consortium may be operating as ecological guilds, with
758 various microbial members having specific contribution to the functioning of
759 the ecosystem (i.e. the host). Under normal conditions, the microbial
760 consortium collectively increase *D. magna* survival and fitness as reported
761 here and in another study (51). But under stressful environmental conditions,
762 the collective beneficial contribution of the microbial consortium to the host
763 can break down (as shown in this study) and the fitness of the host may
764 depend on the genetic and phenotypic traits of a specific microbiome
765 member/s that can positively respond to the given environmental condition.
766 *Pseudomonas* and other *merA*-containing bacteria only constitute a minor
767 percentage of the microbiome consortium in *D. magna*. In this study, we did
768 not investigate if prolonged exposure to mercury will induce changes in
769 microbiota composition or if it can lead to the enrichment of *merA*-positive

770 bacteria in *Daphnia*. It would be interesting to investigate the flexibility of
771 the symbiosis under stressful environmental conditions.

772 In conclusion, this study is one of the few that shows the extension of
773 the microbiome genetic and phenotypic traits to the fitness trait of the host,
774 resulting in a holobiont phenotype that can withstand stressful
775 environmental condition. The microbiome of *Daphnia* should therefore be
776 taken into consideration specifically in ecotoxicological research where
777 *Daphnia* is commonly used as a testing animal.

778

779 **ACKNOWLEDGEMENT**

780 The authors acknowledge the help of John Negrey and Wesley Heim
781 during *Daphnia* CAY sampling, Aigbe Woghiren and Elodie Burcklen for their
782 involvement in *Daphnia* microbiota isolations, Peter Chanseyha for his help in
783 mercury tolerance experiments and *Daphnia* husbandry and Christian Valdez
784 for his assistance in *merA* screening experiments.

785

786 **Funding**

787 The project was funded by UCSC-COR to MPSM. The funder had no role
788 in data collection and interpretation or the decision to submit the work for
789 publication.

790

791 **Disclosures**

792 The authors declare no competing financial interest.

793

794 **Supporting Information**

795 Additional materials and experimental methods, Genbank accession
796 numbers, list of primers (Table S1), assigned taxon of *D. magna* CAY
797 microbiota (Table S2), *merA* Blastn matches (Table S3), MerA protein
798 alignment and tree (Figure S1), *merA* expression in *Pseudomonas*-10 (Figure
799 S2), *Daphnia* mercury LD₅₀ assay (Figure S3), replicates of survival (Figure
800 S4) and fecundity (Figure S5) are available as Supporting Information.

801

802 REFERENCES

- 803 1. Harris, K. D.; Bartlett, N. J.; Lloyd, V. K. Daphnia as an emerging
804 epigenetic model organism. *Genet Res Int.* **2012**, 2012, 147892.
- 805 2. Ebert, D., Ecology, Epidemiology, and Evolution of Parasitism in
806 Daphnia, 2005.
- 807 3. Ebert, D. Genomics. A genome for the environment. *Science.* **2011**,
808 331 (6017), 539-540.
- 809 4. Miner, B. E.; De Meester, L.; Pfrender, M. E.; Lampert, W.; Hairston, N.
810 G., Jr. Linking genes to communities and ecosystems: Daphnia as an
811 ecogenomic model. *Proc Biol Sci.* **2012**, 279 (1735), 1873-1882.
- 812 5. Colbourne, J. K.; Pfrender, M. E.; Gilbert, D.; Thomas, W. K.; Tucker, A.;
813 Oakley, T. H.; Tokishita, S.; Aerts, A.; Arnold, G. J.; Basu, M. K.; Bauer,
814 D. J.; Caceres, C. E.; Carmel, L.; Casola, C.; Choi, J. H.; Detter, J. C.;
815 Dong, Q.; Dusheyko, S.; Eads, B. D.; Frohlich, T.; Geiler-Samerotte, K.
816 A.; Gerlach, D.; Hatcher, P.; Jogdeo, S.; Krijgsveld, J.; Kriventseva, E. V.;
817 Kultz, D.; Laforsch, C.; Lindquist, E.; Lopez, J.; Manak, J. R.; Muller, J.;
818 Pangilinan, J.; Patwardhan, R. P.; Pitluck, S.; Pritham, E. J.; Rechtsteiner,
819 A.; Rho, M.; Rogozin, I. B.; Sakarya, O.; Salamov, A.; Schaack, S.;
820 Shapiro, H.; Shiga, Y.; Skalitzky, C.; Smith, Z.; Souvorov, A.; Sung, W.;
821 Tang, Z.; Tsuchiya, D.; Tu, H.; Vos, H.; Wang, M.; Wolf, Y. I.; Yamagata,
822 H.; Yamada, T.; Ye, Y.; Shaw, J. R.; Andrews, J.; Crease, T. J.; Tang, H.;
823 Lucas, S. M.; Robertson, H. M.; Bork, P.; Koonin, E. V.; Zdobnov, E. M.;
824 Grigoriev, I. V.; Lynch, M.; Boore, J. L. The ecoresponsive genome of
825 Daphnia pulex. *Science.* **2011**, 331 (6017), 555-561.
- 826 6. Watanabe, H.; Kobayashi, K.; Kato, Y.; Oda, S.; Abe, R.; Tatarazako, N.;
827 Iguchi, T. Transcriptome profiling in crustaceans as a tool for
828 ecotoxicogenomics: Daphnia magna DNA microarray. *Cell Biol Toxicol.*
829 **2008**, 24 (6), 641-647.
- 830 7. Orsini, L.; Decaestecker, E.; De Meester, L.; Pfrender, M. E.; Colbourne,
831 J. K. Genomics in the ecological arena. *Biol Lett.* **2011**, 7 (1), 2-3.
- 832 8. Yampolsky, L. Y.; Schaer, T. M.; Ebert, D. Adaptive phenotypic plasticity
833 and local adaptation for temperature tolerance in freshwater
834 zooplankton. *Proc Biol Sci.* **2014**, 281 (1776), 20132744.
- 835 9. Henning-Lucass, N.; Cordellier, M.; Streit, B.; Schwenk, K. Phenotypic
836 plasticity in life-history traits of Daphnia galeata in response to
837 temperature - a comparison across clonal lineages separated in time.
838 *Ecol Evol.* **2016**, 6 (4), 881-891.
- 839 10. Garbutt, J. S.; Little, T. J. Maternal food quantity affects offspring
840 feeding rate in *Daphnia magna*. *Biol Lett.* **2014**, 10 (7), 20140356.
- 841 11. Bukovinszky, T.; Verschoor, A. M.; Helmsing, N. R.; Bezemer, T. M.;
842 Bakker, E. S.; Vos, M.; de Senerpont Domis, L. N. The good, the bad
843 and the plenty: interactive effects of food quality and quantity on the
844 growth of different Daphnia species. *PLoS One.* **2012**, 7 (9), e42966.

- 845 12. Pietrzak, B.; Grzesiuk, M. Food quantity shapes life history and survival
846 strategies in *Daphnia magna* (Cladocera). *Hydrobiologia*. **2010**, 643,
847 51-54.
- 848 13. Petrusek, A.; Tollrian, R.; Schwenk, K.; Haas, A.; Laforsch, C. A "crown
849 of thorns" is an inducible defense that protects *Daphnia* against an
850 ancient predator. *Proc Natl Acad Sci U S A*. **2009**, 106 (7), 2248-2252.
- 851 14. Spanier, K. I.; Leese, F.; Mayer, C.; Colbourne, J. K.; Gilbert, D.;
852 Pfrender, M. E.; Tollrian, R. Predator-induced defences in *Daphnia*
853 *pulex*: selection and evaluation of internal reference genes for gene
854 expression studies with real-time PCR. *BMC Mol Biol*. **2010**, 11, 50.
- 855 15. Schwarzenberger, A.; Courts, C.; von Elert, E. Target gene approaches:
856 Gene expression in *Daphnia magna* exposed to predator-borne
857 kairomones or to microcystin-producing and microcystin-free
858 *Microcystis aeruginosa*. *BMC Genomics*. **2009**, 10, 527.
- 859 16. Tsui, M. T.; Wang, W. X. Temperature influences on the accumulation
860 and elimination of mercury in a freshwater cladoceran, *Daphnia*
861 *magna*. *Aquat Toxicol*. **2004**, 70 (3), 245-256.
- 862 17. Tsui, M. T.; Wang, W. X. Acute toxicity of mercury to *Daphnia magna*
863 under different conditions. *Environ Sci Technol*. **2006**, 40 (12), 4025-
864 4030.
- 865 18. Altshuler, I.; Demiri, B.; Xu, S.; Constantin, A.; Yan, N. D.; Cristescu, M.
866 E. An integrated multi-disciplinary approach for studying multiple
867 stressors in freshwater ecosystems: *Daphnia* as a model organism.
868 *Integr Comp Biol*. **2011**, 51 (4), 623-633.
- 869 19. Orsini, L.; Gilbert, D.; Podicheti, R.; Jansen, M.; Brown, J. B.; Solari, O.
870 S.; Spanier, K. I.; Colbourne, J. K.; Rush, D.; Decaestecker, E.;
871 Asselman, J.; De Schamphelaere, K. A.; Ebert, D.; Haag, C. R.; Kvist, J.;
872 Laforsch, C.; Petrusek, A.; Beckerman, A. P.; Little, T. J.; Chaturvedi, A.;
873 Pfrender, M. E.; De Meester, L.; Frilander, M. J. *Daphnia magna*
874 transcriptome by RNA-Seq across 12 environmental stressors. *Sci Data*.
875 **2016**, 3, 160030.
- 876 20. Rogalski, M. A. Tainted resurrection: metal pollution is linked with
877 reduced hatching and high juvenile mortality in *Daphnia* egg banks.
878 *Ecology*. **2015**, 96 (5), 1166-1173.
- 879 21. Chen, S.; Nichols, K. M.; Poynton, H. C.; Sepulveda, M. S. MicroRNAs are
880 involved in cadmium tolerance in *Daphnia pulex*. *Aquat Toxicol*. **2016**,
881 175, 241-248.
- 882 22. Fan, W.; Ren, J.; Li, X.; Wei, C.; Xue, F.; Zhang, N. Bioaccumulation and
883 oxidative stress in *Daphnia magna* exposed to arsenite and arsenate.
884 *Environ Toxicol Chem*. **2015**, 34 (11), 2629-2635.
- 885 23. Nagato, E. G.; D'Eon J, C.; Lankadurai, B. P.; Poirier, D. G.; Reiner, E. J.;
886 Simpson, A. J.; Simpson, M. J. (1)H NMR-based metabolomics
887 investigation of *Daphnia magna* responses to sub-lethal exposure to
888 arsenic, copper and lithium. *Chemosphere*. **2013**, 93 (2), 331-337.
- 889 24. Boudou, A.; Ribeyre, F. Comparative study of the trophic transfer of
890 two mercury compounds-HgCl₂ and CH₃HgCl-between *Chlorella*

- 891 vulgaris an *Daphnia magna*. Influence of temperature. *Bull Environ*
892 *Contam Toxicol.* **1981**, 27 (5), 624-629.
- 893 25. Tsui, M. T.; Wang, W. X. Uptake and elimination routes of inorganic
894 mercury and methylmercury in *Daphnia magna*. *Environ Sci Technol.*
895 **2004**, 38 (3), 808-816.
- 896 26. Tsui, M. T.; Wang, W. X. Maternal transfer efficiency and
897 transgenerational toxicity of methylmercury in *Daphnia magna*.
898 *Environ Toxicol Chem.* **2004**, 23 (6), 1504-1511.
- 899 27. Kohl, K. D.; Weiss, R. B.; Cox, J.; Dale, C.; Dearing, M. D. Gut microbes
900 of mammalian herbivores facilitate intake of plant toxins. *Ecol Lett.*
901 **2014**, 17 (10), 1238-1246.
- 902 28. Ceja-Navarro, J. A.; Vega, F. E.; Karaoz, U.; Hao, Z.; Jenkins, S.; Lim, H.
903 C.; Kosina, P.; Infante, F.; Northen, T. R.; Brodie, E. L. Gut microbiota
904 mediate caffeine detoxification in the primary insect pest of coffee. *Nat*
905 *Commun.* **2015**, 6, 7618.
- 906 29. Watras, C. J.; Back, R. C.; Halvorsen, S.; Hudson, R. J.; Morrison, K. A.;
907 Wentz, S. P. Bioaccumulation of mercury in pelagic freshwater food
908 webs. *Sci Total Environ.* **1998**, 219 (2-3), 183-208.
- 909 30. Li, W. C.; Tse, H. F. Health risk and significance of mercury in the
910 environment. *Environ Sci Pollut Res Int.* **2015**, 22 (1), 192-201.
- 911 31. Driscoll, C. T.; Mason, R. P.; Chan, H. M.; Jacob, D. J.; Pirrone, N.
912 Mercury as a global pollutant: sources, pathways, and effects. *Environ*
913 *Sci Technol.* **2013**, 47 (10), 4967-4983.
- 914 32. Ha, E.; Basu, N.; Bose-O'Reilly, S.; Dorea, J. G.; McSorley, E.; Sakamoto,
915 M.; Chan, H. M. Current progress on understanding the impact of
916 mercury on human health. *Environ Res.* **2016**, 152, 419-433.
- 917 33. Gonzalez-Raymat, H.; Liu, G.; Liriano, C.; Li, Y.; Yin, Y.; Shi, J.; Jiang, G.;
918 Cai, Y. Elemental mercury: Its unique properties affect its behavior and
919 fate in the environment. *Environ Pollut.* **2017**, 229, 69-86.
- 920 34. Lamborg, C. H.; Hammerschmidt, C. R.; Bowman, K. L.; Swarr, G. J.;
921 Munson, K. M.; Ohnemus, D. C.; Lam, P. J.; Heimbürger, L. E.;
922 Rijkenberg, M. J.; Saito, M. A. A global ocean inventory of
923 anthropogenic mercury based on water column measurements.
924 *Nature.* **2014**, 512 (7512), 65-68.
- 925 35. Fitzgerald, W. F.; Lamborg, C. H.; Hammerschmidt, C. R. Marine
926 biogeochemical cycling of mercury. *Chem Rev.* **2007**, 107 (2), 641-
927 662.
- 928 36. Podar, M.; Gilmour, C. C.; Brandt, C. C.; Soren, A.; Brown, S. D.; Crable,
929 B. R.; Palumbo, A. V.; Somenahally, A. C.; Elias, D. A. Global prevalence
930 and distribution of genes and microorganisms involved in mercury
931 methylation. *Sci Adv.* **2015**, 1 (9), e1500675.
- 932 37. Parks, J. M.; Johs, A.; Podar, M.; Bridou, R.; Hurt, R. A., Jr.; Smith, S. D.;
933 Tomanicek, S. J.; Qian, Y.; Brown, S. D.; Brandt, C. C.; Palumbo, A. V.;
934 Smith, J. C.; Wall, J. D.; Elias, D. A.; Liang, L. The genetic basis for
935 bacterial mercury methylation. *Science.* **2013**, 339 (6125), 1332-1335.

- 936 38. Hsu-Kim, H.; Kucharzyk, K. H.; Zhang, T.; Deshusses, M. A. Mechanisms
937 regulating mercury bioavailability for methylating microorganisms in
938 the aquatic environment: a critical review. *Environ Sci Technol.* **2013**,
939 47 (6), 2441-2456.
- 940 39. Gilmour, C. C.; Podar, M.; Bullock, A. L.; Graham, A. M.; Brown, S. D.;
941 Somenahally, A. C.; Johs, A.; Hurt, R. A., Jr.; Bailey, K. L.; Elias, D. A.
942 Mercury methylation by novel microorganisms from new environments.
943 *Environ Sci Technol.* **2013**, 47 (20), 11810-11820.
- 944 40. Mathema, V. B.; Thakuri, B. C.; Sillanpaa, M. Bacterial mer operon-
945 mediated detoxification of mercurial compounds: a short review. *Arch*
946 *Microbiol.* **2011**, 193 (12), 837-844.
- 947 41. Nascimento, A. M.; Chartone-Souza, E. Operon mer: bacterial
948 resistance to mercury and potential for bioremediation of
949 contaminated environments. *Genet Mol Res.* **2003**, 2 (1), 92-101.
- 950 42. Barkay, T.; Miller, S. M.; Summers, A. O. Bacterial mercury resistance
951 from atoms to ecosystems. *FEMS Microbiol Rev.* **2003**, 27 (2-3), 355-
952 384.
- 953 43. Osborn, A. M.; Bruce, K. D.; Strike, P.; Ritchie, D. A. Distribution,
954 diversity and evolution of the bacterial mercury resistance (mer)
955 operon. *FEMS Microbiol Rev.* **1997**, 19 (4), 239-262.
- 956 44. Ojo, K. K.; Tung, D.; Luis, H.; Bernardo, M.; Leitao, J.; Roberts, M. C.
957 Gram-positive merA gene in gram-negative oral and urine bacteria.
958 *FEMS Microbiol Lett.* **2004**, 238 (2), 411-416.
- 959 45. Jan, A. T.; Azam, M.; Ali, A.; Haq, Q. M. Molecular characterization of
960 mercury resistant bacteria inhabiting polluted water bodies of different
961 geographical locations in India. *Curr Microbiol.* **2012**, 65 (1), 14-21.
- 962 46. Cabral, L.; Giovanella, P.; Gianello, C.; Bento, F. M.; Andrezza, R.;
963 Camargo, F. A. Isolation and characterization of bacteria from mercury
964 contaminated sites in Rio Grande do Sul, Brazil, and assessment of
965 methylmercury removal capability of a *Pseudomonas putida* V1 strain.
966 *Biodegradation.* **2013**, 24 (3), 319-331.
- 967 47. Figueiredo, N. L.; Canario, J.; O'Driscoll, N. J.; Duarte, A.; Carvalho, C.
968 Aerobic Mercury-resistant bacteria alter Mercury speciation and
969 retention in the Tagus Estuary (Portugal). *Ecotoxicol Environ Saf.* **2016**,
970 124, 60-67.
- 971 48. Lloyd, N. A.; Janssen, S. E.; Reinfelder, J. R.; Barkay, T. Co-selection of
972 Mercury and Multiple Antibiotic Resistances in Bacteria Exposed to
973 Mercury in the *Fundulus heteroclitus* Gut Microbiome. *Curr Microbiol.*
974 **2016**, 73 (6), 834-842.
- 975 49. Santos-Gandelman, J. F.; Giambiagi-deMarval, M.; Muricy, G.; Barkay,
976 T.; Laport, M. S. Mercury and methylmercury detoxification potential by
977 sponge-associated bacteria. *Antonie Van Leeuwenhoek.* **2014**, 106 (3),
978 585-590.
- 979 50. Alpers, C. N.; Fleck, J. A.; Marvin-DiPasquale, M.; Stricker, C. A.;
980 Stephenson, M.; Taylor, H. E. Mercury cycling in agricultural and

- 981 managed wetlands, Yolo Bypass, California: spatial and seasonal
 982 variations in water quality. *Sci Total Environ.* **2014**, 484, 276-287.
- 983 51. Sison-Mangus, M. P.; Metzger, C.; Ebert, D. Host genotype-specific
 984 microbiota do not influence the susceptibility of *D. magna* to a
 985 bacterial pathogen. *Sci Rep.* **2018**, 8 (1), 9407.
- 986 52. Sison-Mangus, M. P.; Mushegian, A. A.; Ebert, D. Water fleas require
 987 microbiota for survival, growth and reproduction. *ISME J.* **2015**, 9 (1),
 988 59-67.
- 989 53. Klüttgen, B.; Dülmer, U.; Engels, M.; Ratte, H. T. ADaM, an artificial
 990 freshwater for the culture of zooplankton. *Water Research.* **1994**, 28
 991 (3), 743-746.
- 992 54. Caporaso, J. G.; Kuczynski, J.; Stombaugh, J.; Bittinger, K.; Bushman, F.
 993 D.; Costello, E. K.; Fierer, N.; Pena, A. G.; Goodrich, J. K.; Gordon, J. I.;
 994 Huttley, G. A.; Kelley, S. T.; Knights, D.; Koenig, J. E.; Ley, R. E.;
 995 Lozupone, C. A.; McDonald, D.; Muegge, B. D.; Pirrung, M.; Reeder, J.;
 996 Sevinsky, J. R.; Turnbaugh, P. J.; Walters, W. A.; Widmann, J.;
 997 Yatsunencko, T.; Zaneveld, J.; Knight, R. QIIME allows analysis of high-
 998 throughput community sequencing data. *Nat Methods.* **2010**, 7 (5),
 999 335-336.
- 1000 55. Edgar, R. C. Search and clustering orders of magnitude faster than
 1001 BLAST. *Bioinformatics.* **2010**, 26 (19), 2460-2461.
- 1002 56. Wang, Q.; Garrity, G. M.; Tiedje, J. M.; Cole, J. R. Naive Bayesian
 1003 classifier for rapid assignment of rRNA sequences into the new
 1004 bacterial taxonomy. *Appl Environ Microbiol.* **2007**, 73 (16), 5261-5267.
- 1005 57. Altschul, S. F.; Gish, W.; Miller, W.; Myers, E. W.; Lipman, D. J. Basic
 1006 local alignment search tool. *J Mol Biol.* **1990**, 215 (3), 403-410.
- 1007 58. Weisburg, W. G.; Barns, S. M.; Pelletier, D. A.; Lane, D. J. 16S ribosomal
 1008 DNA amplification for phylogenetic study. *J Bacteriol.* **1991**, 173 (2),
 1009 697-703.
- 1010 59. Wiegand, I.; Hilpert, K.; Hancock, R. E. Agar and broth dilution methods
 1011 to determine the minimal inhibitory concentration (MIC) of
 1012 antimicrobial substances. *Nat Protoc.* **2008**, 3 (2), 163-175.
- 1013 60. Lin, M.-F.; Lin, Y.-Y.; Lan, C.-Y. Minimal inhibitory concentration (MIC)
 1014 assay for *Acinetobacter baumannii*. *Bio-protocol.* **2014**, 4 (23), e1308.
- 1015 61. Hurt, R. A.; Qiu, X.; Wu, L.; Roh, Y.; Palumbo, A. V.; Tiedje, J. M.; Zhou, J.
 1016 Simultaneous recovery of RNA and DNA from soils and sediments. *Appl*
 1017 *Environ Microbiol.* **2001**, 67 (10), 4495-4503.
- 1018 62. Pfaffl, M. W., *Quantification strategies in real-time PCR*, in *A-Z of*
 1019 *quantitativePCR*, S.A. Bustin, Editor. 2004, International University
 1020 Line: La Jolla, CA. p. 87-112.
- 1021 63. Fong, J. C.; Yildiz, F. H. Interplay between cyclic AMP-cyclic AMP
 1022 receptor protein and cyclic di-GMP signaling in *Vibrio cholerae* biofilm
 1023 formation. *J Bacteriol.* **2008**, 190 (20), 6646-6659.
- 1024 64. EPA, U. S. *Method 7473 (SW-846): Mercury in Solids and Solutions by*
 1025 *Thermal Decomposition, Amalgamation, and Atomic Absorption*
 1026 *Spectrophotometry*. Available from: <https://www.epa.gov/esam/epa->

- 1027 [method-7473-sw-846-mercury-solids-and-solutions-thermal-](#)
1028 [decomposition-amalgamation-and.](#)
- 1029 65. Le Coadic, M.; Simon, M.; Marchetti, A.; Ebert, D.; Cosson, P. *Daphnia*
1030 *magna*, a host for evaluation of bacterial virulence. *Appl Environ*
1031 *Microbiol.* **2012**, 78 (2), 593-595.
- 1032 66. Martin-Creuzburg, D.; Beck, B.; Freese, H. M. Food quality of
1033 heterotrophic bacteria for *Daphnia magna*: evidence for a limitation by
1034 sterols. *FEMS Microbiol Ecol.* **2011**, 76 (3), 592-601.
- 1035 67. Freese, H. M.; Martin-Creuzburg, D. Food quality of mixed bacteria-
1036 algae diets for *Daphnia magna*. *Hydrobiologia.* **2012**, 715 (1), 63-75.
- 1037 68. Qi, W.; Nong, G.; Preston, J. F.; Ben-Ami, F.; Ebert, D. Comparative
1038 metagenomics of *Daphnia* symbionts. *BMC Genomics.* **2009**, 10, 172.
- 1039 69. Freese, H. M.; Schink, B. Composition and stability of the microbial
1040 community inside the digestive tract of the aquatic crustacean
1041 *Daphnia magna*. *Microb Ecol.* **2011**, 62 (4), 882-894.
- 1042 70. Sullam, K. E.; Pichon, S.; Schaer, T. M. M.; Ebert, D. The Combined
1043 Effect of Temperature and Host Clonal Line on the Microbiota of a
1044 Planktonic Crustacean. *Microb Ecol.* **2018**, 76 (2), 506-517.
- 1045 71. Peerakietkhajorn, S.; Kato, Y.; Kasalicky, V.; Matsuura, T.; Watanabe, H.
1046 Betaproteobacteria Limnohabitans strains increase fecundity in the
1047 crustacean *Daphnia magna*: symbiotic relationship between major
1048 bacterioplankton and zooplankton in freshwater ecosystem. *Environ*
1049 *Microbiol.* **2016**, 18 (8), 2366-2374.
- 1050 72. Eckert, E. M.; Pernthaler, J. Bacterial epibionts of *Daphnia*: a potential
1051 route for the transfer of dissolved organic carbon in freshwater food
1052 webs. *ISME J.* **2014**, 8 (9), 1808-1819.
- 1053 73. Chaplinska, M.; Gerritsma, S.; Dini-Andreote, F.; Falcao Salles, J.;
1054 Wertheim, B. Bacterial Communities Differ among *Drosophila*
1055 *melanogaster* Populations and Affect Host Resistance against
1056 Parasitoids. *PLoS One.* **2016**, 11 (12), e0167726.
- 1057 74. Early, A. M.; Shanmugarajah, N.; Buchon, N.; Clark, A. G. *Drosophila*
1058 Genotype Influences Commensal Bacterial Levels. *PLoS One.* **2017**, 12
1059 (1), e0170332.
- 1060 75. Chaston, J. M.; Dobson, A. J.; Newell, P. D.; Douglas, A. E. Host Genetic
1061 Control of the Microbiota Mediates the *Drosophila* Nutritional
1062 Phenotype. *Appl Environ Microbiol.* **2015**, 82 (2), 671-679.
- 1063 76. Callens, M.; Macke, E.; Muylaert, K.; Bossier, P.; Lievens, B.; Waud, M.;
1064 Decaestecker, E. Food availability affects the strength of mutualistic
1065 host-microbiota interactions in *Daphnia magna*. *ISME J.* **2016**, 10 (4),
1066 911-920.
- 1067 77. Sotero-Martins, A.; de Jesus, M. S.; Lacerda, M.; Moreira, J. C.;
1068 Filgueiras, A. L.; Barrocas, P. R. A conservative region of the mercuric
1069 reductase gene (*mera*) as a molecular marker of bacterial mercury
1070 resistance. *Braz J Microbiol.* **2008**, 39 (2), 307-310.
- 1071 78. Rodriguez-Rojas, F.; Diaz-Vasquez, W.; Undabarrena, A.; Munoz-Diaz,
1072 P.; Arenas, F.; Vasquez, C. Mercury-mediated cross-resistance to

- 1073 tellurite in *Pseudomonas* spp. isolated from the Chilean Antarctic
1074 territory. *Metallomics*. **2016**, 8 (1), 108-117.
- 1075 79. Rojas, L. A.; Yanez, C.; Gonzalez, M.; Lobos, S.; Smalla, K.; Seeger, M.
1076 Characterization of the metabolically modified heavy metal-resistant
1077 *Cupriavidus metallidurans* strain MSR33 generated for mercury
1078 bioremediation. *PLoS One*. **2011**, 6 (3), e17555.
- 1079 80. Ruiz-Diez, B.; Quinones, M. A.; Fajardo, S.; Lopez, M. A.; Higuera, P.;
1080 Fernandez-Pascual, M. Mercury-resistant rhizobial bacteria isolated
1081 from nodules of leguminous plants growing in high Hg-contaminated
1082 soils. *Appl Microbiol Biotechnol*. **2012**, 96 (2), 543-554.
- 1083 81. Ni Chadhain, S. M.; Schaefer, J. K.; Crane, S.; Zylstra, G. J.; Barkay, T.
1084 Analysis of mercuric reductase (*merA*) gene diversity in an anaerobic
1085 mercury-contaminated sediment enrichment. *Environ Microbiol*. **2006**,
1086 8 (10), 1746-1752.
- 1087 82. Gilbert, M. P.; Summers, A. O. The distribution and divergence of DNA
1088 sequences related to the Tn21 and Tn501 *mer* operons. *Plasmid*. **1988**,
1089 20 (2), 127-136.
- 1090 83. Schaefer, J. K.; Yagi, J.; Reinfelder, J. R.; Cardona, T.; Ellickson, K. M.;
1091 Tel-Or, S.; Barkay, T. Role of the bacterial organomercury lyase (*MerB*)
1092 in controlling methylmercury accumulation in mercury-contaminated
1093 natural waters. *Environ Sci Technol*. **2004**, 38 (16), 4304-4311.
- 1094 84. Mushegian, A. A.; Burcklen, E.; Schar, T. M.; Ebert, D. Temperature-
1095 dependent benefits of bacterial exposure in embryonic development of
1096 *Daphnia magna* resting eggs. *J Exp Biol*. **2016**, 219 (Pt 6), 897-904.
- 1097 85. Cheplick, G. P. Costs of fungal endophyte infection in *Lolium perenne*
1098 genotypes from Eurasia and North Africa under extreme resource
1099 limitation. *Environmental and experimental botany*. **2007**, 60, 202-
1100 210.
- 1101 86. Poynton, H. C.; Varshavsky, J. R.; Chang, B.; Caviglioglio, G.; Chan, S.;
1102 Holman, P. S.; Loguinov, A. V.; Bauer, D. J.; Komachi, K.; Theil, E. C.;
1103 Perkins, E. J.; Hughes, O.; Vulpe, C. D. *Daphnia magna*
1104 ecotoxicogenomics provides mechanistic insights into metal toxicity.
1105 *Environ Sci Technol*. **2007**, 41 (3), 1044-1050.
- 1106 87. Asselman, J.; Shaw, J. R.; Glaholt, S. P.; Colbourne, J. K.; De
1107 Schampelaere, K. A. Transcription patterns of genes encoding four
1108 metallothionein homologs in *Daphnia pulex* exposed to copper and
1109 cadmium are time- and homolog-dependent. *Aquat Toxicol*. **2013**, 142-
1110 143, 422-430.
- 1111 88. Mao, H.; Wang, D. H.; Yang, W. X. The involvement of metallothionein
1112 in the development of aquatic invertebrate. *Aquat Toxicol*. **2012**, 110-
1113 111, 208-213.
- 1114 89. Amiard, J. C.; Amiard-Triquet, C.; Barka, S.; Pellerin, J.; Rainbow, P. S.
1115 Metallothioneins in aquatic invertebrates: their role in metal
1116 detoxification and their use as biomarkers. *Aquat Toxicol*. **2006**, 76
1117 (2), 160-202.
1118

