# UC Santa Barbara

**UC Santa Barbara Previously Published Works** 

# Title

Method to decompose uncertainties in LCA results into contributing factors

**Permalink** https://escholarship.org/uc/item/6j13n8zt

**Journal** The International Journal of Life Cycle Assessment, 26(5)

**ISSN** 0948-3349

**Authors** Qin, Yuwei Suh, Sangwon

Publication Date 2021-05-01

**DOI** 10.1007/s11367-020-01850-5

Peer reviewed

# Method to Decompose Uncertainties in LCA Results into Contributing Factors

- 3
- 4

5 Yuwei Qin<sup>1,2</sup>, Sangwon Suh<sup>1\*</sup>

6

7 <sup>1</sup>Bren School of Environmental Science and Management, 2400 Bren Hall, University of

8 California, Santa Barbara, CA 93106, USA; <sup>2</sup>Department of Civil and Environmental

9 Engineering, University of California, Berkeley, CA 94720, USA

10 \*Corresponding Author: Sangwon Suh; Tel: +1-805-893-7185; Fax: +1-805-893-7612; Email:

- 11 suh@bren.ucsb.edu.
- 12

# **13 ABSTRACT**

# 14 Purpose

15 Understanding uncertainty is essential in using life cycle assessment (LCA) to support decisions.

16 Monte Carlo simulation (MCS) is widely used to characterize the variability in LCA results, be

17 them life cycle inventory (LCI), category indicator results, normalized results, or weighted

18 results. In this study, we present a new method to decompose MCS results into underlying

19 contributors using the logarithmic mean Divisia index (LMDI) decomposition method with a

20 case study on natural gas focusing on two impact categories: global warming and USETox

21 human health impacts.22

# 23 Methods

24 First, after each run of MCS, the difference in simulated and deterministic results is calculated

and the difference is decomposed using the LMDI decomposition method, which returns the

26 contribution of each factor to the difference of the run. After repeating this for 1,000 MCS runs,

27 the statistical properties of the contributions by each factor are analyzed. The method quantifies

28 the contribution of underlying variables, such as characterization factors and LCI items, to the

29 overall variability of the result, such as characterized results.

30

# 31 Results

32 The method presented can decompose the variabilities in LCI, characterized, normalized, or

33 weighted results into LCI items, characterization factors, normalization references, weighting

34 factors, or any subset of them. As an illustrative example, a case study on natural gas LCA was

35 conducted, and the variabilities in characterized results were decomposed into underlying LCI

36 items and characterization factors. The results show that LCI and characterization phases

37 contribute 59% and 41%, respectively, to the uncertainty of the characterized result for global

38 warming. For the human health impact category, LCIs and characterization factors contribute

**39** 32% and 68% to the overall uncertainty, respectively.

## 41 Conclusions and discussion

Using this approach, LCA practitioners can decompose the overall variability in the results to the
underlying contributors under MCS setting, which can help prioritize the parameters that need
further refinement to reduce overall uncertainty in the results. The method reliably estimates the
uncertainty contributions of the variables with large variabilities without the need for large
computational resources, and it can be applied to any stage of an LCA calculation including
normalization and weighting, or to other fields than LCA such as material flow analysis and risk
assessment.

- 50 Keywords: uncertainty analysis, uncertainty contribution, life cycle assessment, Monte Carlo51 simulation, LMDI method
- 52 53

# 54 **1. INTRODUCTION**

55 Life Cycle Assessment (LCA) is a method to quantify the environmental impacts of a 56 product system (ISO 2006; Finnveden et al. 2009). Understanding uncertainty in LCA results is 57 essential in supporting decisions that use them (Geisler et al. 2005; Basson and Petrie 2007; 58 Lloyd and Ries 2008). Quantitative uncertainty analysis has been implemented in many LCA 59 studies (Lo et al. 2005; Bojacá and Schrevens 2010; Clavreul et al. 2012). Two common forms 60 of uncertainty propagation in LCA are (1) the sampling method and (2) the analytical method 61 (Heijungs 1996; Björklund 2002; Heijungs and Huijbregts 2004). In general, the sampling 62 approach can provide more statistics than the analytical approach while requiring much more 63 computer time for large systems, such as the ecoinvent database (Heijungs and Lenzen 2014). 64 Among them, Monte Carlo simulation (MCS), a sampling method, is one of the most widely 65 used methods to characterize the variability in LCA (Huijbregts 2002; Sonnemann et al. 2003; 66 Beltran et al. 2018). Increasing LCA software tools support MCS (SimaPro 2016; OpenLCA 67 2018). Recently, the distribution functions of the entire ecoinvent LCI database have been 68 estimated using MCS (Qin and Suh 2017). 69 Typical MCS results show the distribution of the overall calculation, be them LCI,

69 Typical MCS results show the distribution of the overall calculation, be them LCI,
70 characterized result, normalized result, or weighted result. However, such distributions do not
71 indicate which factor, for example, LCI item, characterization factor, normalization reference, or
72 weighting factor, contributes the most to the overall uncertainty. Sensitivity analysis can be used

73 to quantify the relative importance of the parameters to overall uncertainty. Local sensitivity 74 analysis or one-at-a-time (OAT) technique varies one parameter for each run and measures the 75 change in the result relative to the change in the parameter (Hamby 1994; Hughes et al. 2013). 76 Global sensitivity analysis, on the other hand, simultaneously varies all uncertain parameters and 77 considers parameter interactions to obtain the input-output mapping (Saltelli et al. 2008; 78 Cucurachi et al. 2016). Global sensitivity analysis identifies the most influential parameters that 79 contribute to the output uncertainty by measuring the relative importance of the model 80 parameters, and the method has been utilized to estimate relative contributions of parameters to 81 overall uncertainty (Geisler et al. 2005; de Koning et al. 2010; Mutel et al. 2013; Heijungs and 82 Lenzen 2014; Wei et al. 2015; Groen et al. 2017; Igos et al. 2019; Patouillard et al. 2019). 83 Parsing out overall distribution results to contributing factors becomes computationally intensive 84 under an MCS setting (Ye and Hill 2017). Furthermore, although global sensitivity analysis 85 provides the ranking and the sensitivity indices of contributing parameters, the index values 86 cannot be interpreted as the measure of uncertainty contributed to the overall results (Xu and 87 Gertner 2011). Attempts to avoid using rank orders and indices for uncertainty contributions 88 have been made using regression or correlation analyses (Heijungs and Lenzen 2014; Groen et 89 al. 2017; Igos et al. 2019), which require an even larger number of samples, exacerbating the 90 problem of computational intensity.

91 In this paper, we present a new method to decompose the overall uncertainty of an LCA 92 study derived from MCS. The method is then applied to a case study on a natural gas LCA. We 93 compare our method against previously reported methods of analyzing uncertainty contributions. 94 In general, the change in the overall results of a model involving multiple variables can be 95 allocated over contributing factors using decomposition analysis methods. Oaxaca (1973) and 96 Blinder (1973) developed a decomposition method to analyze the wage differences by race and 97 sex, which is now a standard method in applied economics (Jann 2008). The basic idea of the 98 Oaxaca-Blinder decomposition has since been improved and expanded over the past three 99 decades and evolved into various decomposition analysis approaches (Boyd et al. 1988; Fortin et 100 al. 2011; Su and Ang 2012). Two main approaches of decomposition analysis are index 101 decomposition analysis (IDA) and structural decomposition analysis (SDA) (Hoekstra and Van 102 den Bergh 2003). IDA uses aggregate sector information, which only assesses the impact of

103 direct effects, while SDA uses the economic input–output analysis framework allowing

104 decomposition of both direct and indirect effects such as changes in economic structures,

105 international sourcing, and changes in consumption patterns and volumes (Dietzenbacher and

106 Hoekstra 2002; Hoekstra and Van Den Bergh 2002; Hoekstra et al. 2016). In our study, we chose

107 IDA, as our objective here is to demonstrate the method to decompose the overall uncertainty

108 into the four phases of LCA calculation without involving the analysis of structural effects.

109 However, SDA can also be applied under the same framework that we are proposing here.

110 Index decomposition analysis was first developed to study the impact of structural changes 111 on energy use by industry in the late 1970s (Kako 1978; Jenne and Cattell 1983; Marlay 1984). 112 IDA has been used to quantify the impact of different factors on the change of energy intensity 113 and extended to many regions and various application areas such as transportation, electricity 114 generation and environmental study (Ang et al. 1998; Paul and Bhattacharya 2004; Malla 2009; 115 Al-Ghandoor et al. 2010). For example, Zhang et al. (2009) used the IDA to decompose the 116 influence of energy-related factors in  $CO_2$  emission reduction in China.

Ang (2004) reviewed IDA studies and concluded that the logarithmic mean Divisia index (LMDI) method is the most preferable decomposition method. LMDI leaves no residuals in the analysis and performs well even with multiple variables and zeros in the dataset (Ang 2004; Ang and Liu 2007; Meng et al. 2018). The LMDI has been widely used in economy-wide studies and also in the energy field (Boyd et al. 1987; Ang and Liu 2001; Timilsina and Shrestha 2009;

**122** Baležentis et al. 2011; Jeong and Kim 2013; González et al. 2014).

123 The LMDI method in these applications was used as a method to decompose changes in the 124 total results over time. We use the LMDI to decompose the variability of the results. The LMDI 125 method was originally applied for the decomposition of variabilities in results over time. In our 126 case, however, we apply the same decomposition method, while the variability in the results is 127 not from the changes over time but from the perturbations in MCS. For example, the overall 128 variability in a characterized result, say climate change impact measured by global warming 129 potential (GWP) 100, of an electric vehicle can be decomposed into the variabilities in 130 underlying factors such as the amount of cobalt or lithium needed to produce a unit of the 131 electric vehicle.

This work aims to provide a methodology for quantifying the contribution of each variable
in an LCA model to the overall variability of the model results using the LMDI method. To our
best knowledge, this paper represents the first attempt to apply the technique of index
decomposition analysis to uncertainty analysis.

136 This paper is organized as follows: a detailed methodology description of the proposed 137 method is presented in the Methods section. We then apply the method using a case study in a 138 subsection, Case study. The next section, Results, presents the findings of the case study. The 139 Discussion and Conclusions section discusses the implications of the proposed method and 140 concludes the paper.

#### 141 **2. METHODS**

#### 142 **2.1 Basic Principle**

143 Decomposing MCS results into contributing factors is challenging because the outcome of 144 an MCS, which is a distribution, is a result of simultaneous sampling of all the variables involved 145 in a model. Therefore, parsing out the contribution by each variable directly to the overall shape 146 of the distribution is not feasible based on the MCS result alone. However, the outcome of each 147 MCS run deviates from the outcome of the deterministic model generated from default values for 148 all parameters. This difference between the deterministic and simulated results can be attributed 149 to the differences between the default and sampled values of the underlying variables using a 150 decomposition method. The result of decomposition can then be used as a measure of 151 contribution by each variable to the variability of the overall result. The contribution by each 152 variable will vary between model runs, as different values will be sampled for each run according 153 to the stochastic properties of those variables. By sampling the variables, running the model, and 154 decomposing the variability in the results for a sufficiently large number of times, it is possible to 155 estimate the distribution of the contribution by each variable to the overall variability in the 156 results. This basic idea is presented in more detail in steps from the following subsection.

#### **157 2.2** Calculation steps for uncertainty contribution analysis

158 Six main steps are involved in the proposed method to calculate the uncertainty

159 contributions of underlying factors in an LCA. Fig. 1 shows a summary of the procedure. For the

160 sake of simplicity, we will use the characterized result,  $h_i$  as an example. Calculation steps for

- 161 normalized and weighted results are presented later in this section.
- 162

163



Fig. 1. Flow diagram of the use of the LMDI method in decomposing the uncertainty in the
 LCA results.

166 Fig. 2 (a) illustrates the procedure graphically. For each run from k = 1 to k = n, the overall

167 difference in the deterministic and simulated results,  $\Delta h_i$  is decomposed into underlying factors

```
168 that are denoted as m_1, m_2, \dots, m_l (LCI items) and c_{i,1}, c_{i,2}, \dots, c_{i,l} (characterization factors). By
```

169 running the decomposition for *n* times, the distribution of the contribution to  $\Delta h_i$  by each factor

170 can be derived through the steps explained in the following section (Fig. 2 (b)). The average of

171 the contribution can be illustrated in a pie chart (Fig. 2 (c)).





Fig. 2. Graphic presentation of the uncertainty decomposition method using the LMDI
 decomposition method combined with Monte Carlo samples.

## 176 Step 1. Deterministic result calculation

177 The first step is to calculate the deterministic characterized result. The characterized LCA178 result is calculated through

$$h_i = \sum_j c_{i,j} m_j \tag{1}$$

180 where,  $h_i$  is the characterized result for characterization model *i*;

181  $c_{ij}$  is the characterization factor for the elementary flow *j* in impact category *i*;

- 182  $m_j$  is the inventory for the elementary flow j.
- 183 Step 2. Simulated result calculation

184 The second step is to simulate LCI items and characterization factors by randomly selecting

185 values from their specified distributions and store both simulated values and the characterized

**186** LCA result,  $h_i^k$ . In the study, k represents simulation runs. I.e.,  $h_i^k$  is the kth simulation of the

characterized result for impact category *i*. Deterministic value for *i*th characterized result is noted 187 as  $h_i^0$ , where k = 0. 188 189 The equation for calculating the characterized result in impact category *i* in simulation *k* is 190 provided as follows:  $h_{i}^{k} = \sum_{j} h_{i,j}^{k} = \sum_{j} c_{i,j}^{k} m_{j}^{k}$ 191 (2)where  $h_i^k$  is the characterized result for impact category *i* in simulation *k*; 192  $h_{i,j}^{k}$  is the characterized result of elementary flow j for impact category i in 193 194 simulation k;  $c_{i,j}^{k}$  is the characterization factor for the elementary flow j in impact category i in 195 196 simulation k;  $m_j^k$  is the life cycle inventory for the elementary flow *j* in simulation *k*. 197 198 Step 3. Difference calculation 199 The third step is to calculate the difference between the simulated LCA result and the 200 deterministic LCA result for each simulation, and the difference is considered the change in the 201 LCA results:  $\Delta h_i^k = h_i^k - h_i^0$ 202 (3) 203 **Step 4. Decomposition of the difference** 204 The next step is to apply the LMDI decomposition method to find the contribution of each LCI item and characterization factor in the change of the LCA result for each simulation. The 205 difference between the simulated and deterministic characterized results,  $h_i^k - h_i^0$ , can be 206

207 decomposed into the influence of LCI items and characterization factors,  $c_{i,j}$  and  $m_j$ ,

208 respectively.

The calculation of aggregate changes from  $h_i^0$  in the deterministic result to  $h_i^k$  in simulation *k* followed the LMDI approach by Ang (2005, 2015). The additive decomposition suggests:

$$211 \qquad \Delta h_i^k = h_i^k - h_i^0 = \Delta h_{ic}^k + \Delta h_{\Im}^k \tag{4}$$

212 where  $\Delta h_{ic}^{k}$  is the change in characterized result for impact category *i* in simulation *k* 213 attributable to the variabilities in characterization factors;

- 214  $\Delta h_3^k$  the change in the characterized result for impact category *i* in simulation *k* 215 attributable to the variabilities in LCI items.
- 216

217 In multiplicative decomposition method, the difference can be decomposed:

218 
$$D_{h_i}^k = h_i^k / h_i^0 = D_{ic}^k D_{\Im}^k$$
 (5)

- 219 where  $D_{ic}^{k}$  is the changes in characterized result for impact category *i* in simulation *k* 220 attributable to the variabilities in characterization factors;
- 221  $D_{\Im}^{k}$  is the changes in characterized results for impact category *i* in simulation *k* 222 attributable to the variabilities in LCI items.

223 Using the LMDI approach,  $\Delta h_{ic}^k$  and  $\Delta h_{\Im}^k$  can be decomposed by additive decomposition:

224 
$$\Delta h_{ic}^{k} = \sum_{j} L(h_{i,j}^{k}, h_{i,j}^{0}) \ln\left(\frac{c_{i,j}^{k}}{c_{i,j}^{0}}\right) = \sum_{j} \frac{h_{i,j}^{k} - h_{i,j}^{0}}{\ln h_{i,j}^{k} - \ln h_{i,j}^{0}} \ln\left(\frac{c_{i,j}^{k}}{c_{i,j}^{0}}\right)$$

226 
$$\Delta h_{\Im}^{k} = \sum_{j} L\left(h_{i,j}^{k}, h_{i,j}^{0}\right) \ln\left(\frac{m_{j}^{k}}{m_{j}^{0}}\right) = \sum_{j} \frac{h_{i,j}^{k} - h_{i,j}^{0}}{\ln h_{i,j}^{k} - \ln h_{i,j}^{0}} \ln\left(\frac{m_{j}^{k}}{m_{j}^{0}}\right)$$
(7)

- 227
- By multiplicative decomposition:
- $229 \qquad D_{ic}^{k} = \exp i i \qquad (8)$
- $230 \qquad D_3^k = \exp i i \qquad (9)$
- 231 where L(a, b) = (a-b)i is the logarithmic mean (Ang, 2004).

## 232 Step 5. Repeat steps 2-4

This step repeats steps 2 to 4 for n=1,000 times, and the results of each run are stored, so that, once completed, the statistical properties of the decomposition results are analyzed in the next step.

236 Step 6. Analysis of the distribution

The final step is to calculate the average contribution of each LCI item and characterization

**238** factor to the change in the LCA result.

For additive decomposition:

240 
$$\Delta \dot{h}_{ic} = \frac{\sum_{k} \Delta h_{ic}^{k}}{n}$$

241 (10)  
242 
$$\frac{\sum_{k} \Delta h_{\mathfrak{I}}}{\Delta h_{\mathfrak{I}}} = \frac{\sum_{k} \Delta h_{\mathfrak{I}}^{k}}{n}$$
(11)

For multiplicative decomposition:

244 
$$D_{ic}^{k} = \frac{\sum_{k} D_{ic}^{k}}{n}$$

245 (12)  
246 
$$\overline{D_{3}^{k}} = \frac{\sum_{k} D_{3}^{k}}{n}$$
 (13)

The LMDI decomposition method in Step 4 can be extended to find the uncertainty

contribution in normalized and weighted results. The LMDI decomposition method for weightedresults is presented in the following section.

250 **2.3** The L

#### 2.3 The LMDI method for weighted results

Conceptually, the same steps described in the previous section can be applied to any stage of
LCA calculation including LCI, characterized results, normalized results, and weighted results.
Shown below is an application of the method to weighted results, where the contributions of the
LCI items, characterization factors, normalization references, and weighting factors can be
calculated.

Weighted results are calculated using Eq. 14 (Step 1):

257 
$$W = \sum_{i} w_{i}(h_{i}i/n_{i}) = \sum_{i} w_{i}(h_{i}i/n_{i}) = \sum_{i} w_{i}h_{i}q_{i} = \sum_{i,j} w_{i}c_{i,j}m_{j}q_{i}ii$$

258 (14)

259where W is the normalized and weighted result;260
$$w_i$$
 is the weighting factor for impact category  $i$ ;261 $h_i$  is characterized result of impact category  $i$ ;262 $n_i$  is the normalization reference for impact category  $i$ ;263 $q_i$  is the inverse of the normalization reference,  $n_{i,}$ , for impact category  $i$ ;264 $c_{i,j}$  is the characterization factor for the elementary flow  $j$  in impact category  $i$ ;

 $m_i$  is the LCI item for the elementary flow *j*.

266 267 Using Eq. 15, deterministic and simulated results are calculated (Step 2), and the difference 268 between the two is derived (Step 3). Then the difference is decomposed into inventory items, 269 characterization factors, normalization references, and weighting factors such that (Step 4): 270  $\Delta W^{k} = W^{k} - W^{0} = \sum_{i} \Delta W^{k}_{iw} + \Delta W^{k}_{ic} + \Delta W^{k}_{3} + \Delta W^{k}_{iq} \text{ (additive decomposition)}$ 271 (15)272 where  $\Delta W$  is the change in the normalized and weighted result;  $\Delta W_{iw}^{k}$  is the change in normalized and weighted results for impact category *i* in 273 274 simulation k attributable to the variability in weighting factors; 275  $\Delta W_{ic}^{k}$  is the change in normalized and weighted results for impact category *i* in simulation *k* attributable to the variabilities in characterization factors; 276 277  $\Delta W_{\mathfrak{I}}^k$  is the change in normalized and weighted results for impact category *i* in simulation *k* attributable to the variabilities in LCI items; 278  $\Delta W_{iq}^k$  is the effect of the variabilities in inversed normalization reference contributed 279 to the change in normalized and weighted results for impact category *i* in simulation *k*. 280 281  $D_W^k = W^k / W^0 = \sum_{i} D_{iw}^k D_{ic}^k D_{\Im}^k D_{iq}^k$  (multiplicative decomposition) 282 (16)where  $D_W$  is the change in the normalized and weighted result; 283  $D_{iw}^{k}$  is the change in normalized and weighted results for impact category *i* in 284 285 simulation k attributable to the variability in weighting factor; 286  $D_{ic}^{k}$  is the change in normalized and weighted results for impact category *i* in 287 simulation k attributable to the variabilities in characterization factors;  $D_{\mathfrak{I}}^{k}$  is the change in normalized and weighted results for impact category *i* in 288 289 simulation k attributable to the variabilities in LCI items; 290  $D_{iq}^{k}$  is the change in normalized and weighted results for impact category *i* in 291 simulation k attributable to the effect of the variabilities in inversed normalization 292 references. 293

294 Under additive decomposition, the terms on the right-hand-side of Eq. 15 are calculated by:

295 
$$\Delta W_{w}^{k} = \sum_{i,j} L(W_{i,j}^{k}, W_{i,j}^{0}) \ln\left(\frac{w_{i}^{k}}{w_{i}^{0}}\right) = \sum_{i,j} \frac{W_{i,j}^{k} - W_{i,j}^{0}}{\ln W_{i,j}^{k} - \ln W_{i,j}^{0}} \ln\left(\frac{w_{i}^{k}}{w_{i}^{k}}\right)$$
(17)

296 
$$\Delta W_c^k = \sum_{i,j} L(W_{i,j}^k, W_{i,j}^0) \ln\left(\frac{c_{i,j}^k}{c_{i,j}^0}\right) = \sum_{i,j} \frac{W_{i,j}^k - W_{i,j}^0}{\ln W_{i,j}^k - \ln W_{i,j}^0} \ln\left(\frac{c_{i,j}^k}{c_{i,j}^0}\right)$$

297 (18)

298 
$$\Delta W_m^k = \sum_{i,j} L(W_{i,j}^k, W_{i,j}^0) \ln\left(\frac{m_j^k}{m_j^0}\right) = \sum_{i,j} \frac{W_{i,j}^k - W_{i,j}^0}{\ln W_{i,j}^k - \ln W_{i,j}^0} \ln\left(\frac{m_j^k}{m_j^0}\right)$$

299 (19)

$$300 \qquad \Delta W_{q}^{k} = \sum_{i,j} L(W_{i,j}^{k}, W_{i,j}^{0}) \ln\left(\frac{q_{i}^{k}}{q_{i}^{0}}\right) = \sum_{i,j} \frac{W_{i,j}^{k} - W_{i,j}^{0}}{\ln W_{i,j}^{k} - \ln W_{i,j}^{0}} \ln\left(\frac{q_{i}^{k}}{q_{i}^{0}}\right)$$
(20)

301 Under multiplicative decomposition, the terms on the right-hand-side of Eq. 16 are calculated by:302

303 
$$D_{w}^{k} = \exp\left(\sum_{i,j} \frac{L\left(W_{i,j}^{k}, W_{i,j}^{0}\right)}{L\left(W^{k}, W^{0}\right)} \ln\left(\frac{w_{i}^{k}}{w_{i}^{0}}\right)\right) = \exp i i$$
(21)

$$304 D_c^k = \exp\left(\sum_{i,j} \frac{L(W_{i,j}^k, W_{i,j}^0)}{L(W^k, W^0)} \ln\left(\frac{c_{i,j}^k}{c_{i,j}^0}\right)\right) = \exp \dot{\iota} \dot{\iota}$$
(22)

$$305 D_m^k = \exp\left(\sum_{i,j} \frac{L(W_{i,j}^k, W_{i,j}^0)}{L(W^k, W^0)} \ln\left(\frac{m_j^k}{m_j^0}\right)\right) = \exp i i (23)$$

$$306 \qquad D_q^k = \exp\left(\sum_{i,j} \frac{L(W_{i,j}^k, W_{i,j}^0)}{L(W^k, W^0)} \ln\left(\frac{q_i^k}{q_i^0}\right)\right) = \exp i i \qquad (24)$$

307

## 3082.4 Case study

The additive decomposition method presented in the previous section is applied to a natural

310 gas LCA based on the ecoinvent database, version 3.1 (Allocation, default system model),

311 (Wernet et al. 2016) to demonstrate the applicability of the method. Natural gas is the largest

source (33%) of electricity generation in the U.S and will remain the primary energy source in

313 the near future (EIA 2019). Characterization factors chosen for the comparison in the study are

**314** GWP 100 from IPCC 2013 (Stocker 2014) and carcinogenic human toxicity impact from

- 315 USEtox (Rosenbaum et al. 2008) because global warming impact is a time and space-insensitive
- 316 model while the human health impact is time and space-sensitive.
- 317 The process flow diagram of natural gas production in the U.S. is presented in Fig. 3.
- 318



decomposition method, which returns the contribution of each factor to the difference of the run.

Then, we analyzed the statistical properties of the contributions by each factor after 1,000 MCSruns.

#### **334 3. RESULTS**

The deterministic values for life cycle greenhouse gas emissions (GHG) and carcinogenic human toxicity impacts of 1 m<sup>3</sup> of natural gas in the U.S. were, 0.45 kg of CO<sub>2</sub>e and 1.27e-08 comparative toxic units (CTU<sub>h</sub>), respectively. Fig. 4 shows the distribution of the simulated characterized results for the two impact categories. The average global warming impact of natural gas was 0.46 kg of CO<sub>2</sub>e, and the corresponding standard deviation was 0.067 kg of

CO<sub>2</sub>e. The average carcinogenic human toxicity impact was 1.39e-08 CTU<sub>h</sub>, and the
corresponding standard deviation was 4.46e-09 CTU<sub>h</sub>.





Fig. 5. (a) The distributions of the uncertainty contributions of the most influential factors in
 the climate change impact of natural gas production. (b) The average contribution of each factor
 to the overall uncertainty. The results are subject to change based on the selections of the
 Pedigree scores in the study.

361 The USEtox carcinogenic human health results indicated that 32% and 68% of the 362 uncertainty can be attributed to LCI and characterization factors, respectively (Fig. 6). Among the LCI items, Chromium VI, Cr<sup>+</sup>, contributed to 30.3% of the overall uncertainty. Arsenic, 363 364 nickel, and the remainder of the LCI contributed to 1.2%, 0.4%, and 0.4%, respectively to the 365 overall uncertainty in the characterized LCA result. Among the characterization factors, 366 Chromium VI contributed 68% of the overall uncertainty. Arsenic, nickel, and the rest of the LCI 367 contributed to 3.0%, 0.8%, and 1.1%, respectively to the uncertainty in the characterized LCA 368 result.

369



370

Fig. 6. (a) The distributions of the uncertainty contributions of the most influential factors in
the carcinogenic human health impact of natural gas production. (b) The average contribution of
each factor to the overall uncertainty. The results are subject to change based on the selections of
the Pedigree scores in the study.

Both cases in climate change and human health carcinogen impacts suggested that the top 2 or 3 factors in LCI and characterization factors contributed to the majority (>90%) of the uncertainty, and the rest of LCI and characterization factors only had little (<10%) influence on the overall uncertainty of the characterized results. We have tested the multiplicative approach to the characterized results, and the relative contributions by underlying factors between additive and multiplicative approaches were identical. Improving the reliability of those top contributors, therefore, would reduce the uncertainty of the characterized results more effectively.

## **4. CONCLUSIONS AND DISCUSSION**

This paper introduces a method to quantify the contributions of underlying variables in LCA to overall variability in the result. The proposed method uses the LMDI decomposition method combined with Monte Carlo simulation, with minimal additional needs of computational resources. To our knowledge, the method proposed in this paper is the first attempt to decompose the overall variability of an LCA derived from MCS into the variabilities of underlying parameters using an index decomposition approach.

390	Table 1 summarizes previously reported approaches and our method (LMDI decomposition)
391	for analyzing uncertainty contributions drawing mainly from recent papers that reviewed
392	multiple approaches (Groen et al., 2017; Igos et al. 2019). These approaches are compared
393	against (1) the uncertainty propagation methods used, (2) the ability to explain small and larger
394	input variabilities, (3) reliability in the results, and (4) computation time for sampling and
395	calculation. According to Groen et al. (2017), global sensitivity methods require a sampling size
396	(N) of $10^6$ or larger for reliable results. Igos et al (2019) concluded that Sobol' indices method
397	provides more reliable results than other methods, while it requires extensive computation time.
398	Finally, as shown in Table 1, all but one (Sobol' total effect) of the existing approaches are not
399	able to generate reliable uncertainty contributions for large input variability. The results from the
400	LMDI decomposition method presented in this paper are close to those of the top contributors
401	from Sobol' total effect approach, which is recognized as the most reliable approach in the
402	literature, while substantially reducing the computation time. The results of the LMDI
403	decomposition method, Sobol' total effect approach, and the OAT method are summarized in the
404	Supplementary Information. Therefore, we believe that the LMDI decomposition method offers
405	high reliability with reasonable computation time and is suitable for both small and large input
406	variabilities.

Table 1.	Key chara	cteristics of	common	sensitivity	methods.
	•/			•/	

Approach	Reference	Uncertainty propagation	Explain small input variability <sup>a</sup>	Explain large input variability <sup>a</sup>	Reliability <sup>a</sup>	Computation time <sup>a</sup>
Local sensitivity analysis						
One-at-a-time analysis	(Hamby 1994)	Sampling	Yes	Yes	Medium	Long
Perturbation analysis	(Heijungs and	Analytical	Yes	No	Medium	Short
	Kleijn 2001)					
Global sensitivity analysis						
Key issue analysis	(Heijungs 2010)	Analytical	Yes	No	Low	Short
Standardized regression coefficient	(Huijbregts et al. 2001)	Sampling	Yes	No	Medium	Medium
Spearman correlation coefficient	(Sonnemann et al. 2003)	Sampling	Yes	No	Medium	Medium
Sobol' main effect	(Sobol 2001)	Sampling	Yes	No	High	Long
Sobol' total effect	(Saltelli et al. 2010)	Sampling	Yes	Yes	High	Long

Random balance design	(Tarantola et al.	Sampling	No	No	Low	Medium
	2006)					
LMDI decomposition	This paper	Sampling	Yes	Yes	High	Medium

408 <sup>a</sup> Drawn mostly from Groen et al (2017) and Igos et al (2019). The computation time is estimated from the
409 calculation using the entire ecoinvent database.

410

411 Our case study showed a use case of the method proposed in this paper. A deterministic 412 LCA result showed that 1 m<sup>3</sup> of natural gas generates 0.45 kg of CO<sub>2</sub>e and 1.27e-08 CTU<sub>h</sub> of 413 characterized impacts for climate change and carcinogenic human toxicity categories throughout 414 its life-cycle. The standard deviations of the distributions of climate change and carcinogenic 415 human toxicity impacts were 0.067 kg of CO<sub>2</sub>e and 4.46e-09 CTU<sub>h</sub>, respectively. These 416 distributions were then decomposed into underlying factors using the LMDI decomposition 417 method as proposed in this paper. The results show that methane was the largest contributor to 418 the overall variability of the characterized result of climate change, and Chromium VI was the 419 largest contributor to the overall variability of the characterized result of carcinogenic human 420 toxicity. Future data collection and refinement efforts can focus on these categories to more 421 effectively reduce the overall variability of the results. 422 It is notable that our method and the case study only considered parametric uncertainty, 423 which is the most commonly addressed uncertainty type in LCA studies (Lloyd and Ries 2008).

424 These results are based on the uncertainty estimates from the Pedigree matrix for both LCI and 425 characterization factors. Whether the Pedigree method is an appropriate approach to quantifying 426 the variabilities of parameters in LCA is still debated (Qin et al. 2020). The method presented in 427 this paper is agnostic about the method of variability estimation or the type of distribution

428 functions used.

The method presented can be used in other fields of science, where quantifying the influence of underlying variables on the overall variability of the results is useful. For example, the proposed method can be used to quantify the uncertainty contribution of population, affluence, and technology to the impact of human activities on the environment using the IPAT equation (York et al. 2003; Ma et al. 2017). Likewise, the proposed method can be used to analyze uncertainty contributions in Kaya identity, where the total GHG emissions are expressed as a product of GHG emissions intensity of energy, energy intensity of fuels, fuel consumption 436 intensity of products, final consumption per capita, and population (Jung et al. 2012; Pachauri et
437 al. 2014). The results of the proposed method of uncertainty contribution analysis can help policy
438 and decision makers better understand the uncertainty in the results and prioritize the research

439 effort to reduce the overall uncertainty.

440

## 441 ACKNOWLEDGEMENT

442 We thank Dr. Reinout Heijungs for his constructive review comments. We are grateful for

the financial support from the Assistance Agreement No. 83557901 awarded by the U.S.

444 Environmental Protection Agency to University of California Santa Barbara. This paper has not

been formally reviewed by EPA. The views expressed in this document are solely those of the

446 authors and do not necessarily reflect those of the Agency. EPA does not endorse any products or

447 commercial services mentioned in this publication.

448

### 449 **REFERENCES**

- Al-Ghandoor A, Al-Hinti I, Mukattash A, Al-Abdallat Y (2010) Decomposition analysis of
   electricity use in the Jordanian industrial sector. Int J Sustain Energy 29:233–244
- Ang BW (2015) LMDI decomposition approach: a guide for implementation. Energy Policy
   86:233–238
- Ang BW (2004) Decomposition analysis for policymaking in energy: which is the preferred
   method? Energy Policy 32:1131–1139
- Ang BW (2005) The LMDI approach to decomposition analysis: a practical guide. Energy Policy
   33:867–871
- Ang BW, Liu FL (2001) A new energy decomposition method: perfect in decomposition and consistent in aggregation. Energy 26:537–548
- Ang BW, Liu N (2007) Handling zero values in the logarithmic mean Divisia index
   decomposition approach. Energy Policy 35:238–246
- Ang BW, Zhang FQ, Choi K-H (1998) Factorizing changes in energy and environmental
   indicators through decomposition. Energy 23:489–495

464 Baležentis A, Baležentis T, Streimikiene D (2011) The energy intensity in Lithuania during 465 1995–2009: A LMDI approach. Energy Policy 39:7322–7334 466 Basson L, Petrie JG (2007) An integrated approach for the consideration of uncertainty in 467 decision making supported by Life Cycle Assessment. Environ Model Softw 22:167–176 468 Beltran MM, Pomponi F, Guinée JB, Heijungs R (2018) Uncertainty Analysis in Embodied 469 Carbon Assessments: What Are the Implications of Its Omission? In: Embodied Carbon 470 in Buildings. Springer, pp 3–21 471 Björklund AE (2002) Survey of approaches to improve reliability in lca. Int. J. Life Cycle 472 Assess. 7:64–72 473 Blinder AS (1973) Wage discrimination: reduced form and structural estimates. J Hum Resour 474 436-455 475 Bojacá CR, Schrevens E (2010) Parameter uncertainty in LCA: stochastic sampling under 476 correlation. Int J Life Cycle Assess 15:238-246 477 Boyd G, McDonald JF, Ross M, Hanson DA (1987) Separating the changing composition of US 478 manufacturing production from energy efficiency improvements: a Divisia index 479 approach. Energy J 8:77–96 480 Boyd GA, Hanson DA, Sterner T (1988) Decomposition of changes in energy intensity: a 481 comparison of the Divisia index and other methods. Energy Econ 10:309–312 482 Clavreul J, Guyonnet D, Christensen TH (2012) Quantifying uncertainty in LCA-modelling of 483 waste management systems. Waste Manag 32:2482–2495 484 Cucurachi S, Borgonovo E, Heijungs R (2016) A protocol for the global sensitivity analysis of 485 impact assessment models in life cycle assessment. Risk Anal 36:357-377 486 de Koning A, Schowanek D, Dewaele J, et al (2010) Uncertainties in a carbon footprint model 487 for detergents; quantifying the confidence in a comparative result. Int J Life Cycle Assess 488 15:79 489 Dietzenbacher E, Hoekstra R (2002) The RAS structural decomposition approach. In: Trade, 490 Networks and Hierarchies. Springer, pp 179–199 491 EIA (2019) Short-Term Energy Outlook. U.S. Energy Information Administration, Washington, 492 D.C. 493 Finnveden G, Hauschild MZ, Ekvall T, et al (2009) Recent developments in life cycle 494 assessment. J Environ Manage 91:1–21 495 Fortin N, Lemieux T, Firpo S (2011) Decomposition methods in economics. In: Handbook of 496 labor economics. Elsevier, pp 1–102

497 Geisler G, Hellweg S, Hungerbühler K (2005) Uncertainty Analysis in Life Cycle Assessment 498 (LCA): Case Study on Plant-Protection Products and Implications for Decision Making 499 (9 pp + 3 pp). Int J Life Cycle Assess 10:184–192. 500 https://doi.org/10.1065/lca2004.09.178 501 González PF, Landajo M, Presno MJ (2014) Tracking European Union CO2 emissions through 502 LMDI (logarithmic-mean Divisia index) decomposition. The activity revaluation 503 approach. Energy 73:741–750 504 Groen EA, Bokkers EA, Heijungs R, de Boer IJ (2017) Methods for global sensitivity analysis in 505 life cycle assessment. Int J Life Cycle Assess 22:1125–1137 506 Gustafson P, Srinivasan C, Wasserman L (1996) Local sensitivity analysis. Bayesian Stat 5:197-507 210 508 Hamby DM (1994) A review of techniques for parameter sensitivity analysis of environmental 509 models. Environ Monit Assess 32:135-154 510 Heijungs R (1996) Identification of key issues for further investigation in improving the 511 reliability of life-cycle assessments. J Clean Prod 4:159-166 512 Heijungs R (2010) Sensitivity coefficients for matrix-based LCA. Int J Life Cycle Assess 513 15:511-520. https://doi.org/10.1007/s11367-010-0158-5 514 Heijungs R, Huijbregts MA (2004) A review of approaches to treat uncertainty in LCA. Orlando 515 Fla Elsevier 516 Heijungs R, Kleijn R (2001) Numerical approaches towards life cycle interpretation five 517 examples. Int J Life Cycle Assess 6:141-148 518 Heijungs R, Lenzen M (2014) Error propagation methods for LCA-a comparison. Int J Life 519 Cycle Assess 19:1445–1461 520 Hoekstra R, Michel B, Suh S (2016) The emission cost of international sourcing: using structural 521 decomposition analysis to calculate the contribution of international sourcing to CO 2 -522 emission growth. Econ Syst Res 28:151-167. 523 https://doi.org/10.1080/09535314.2016.1166099 524 Hoekstra R, Van den Bergh JC (2003) Comparing structural decomposition analysis and index. 525 Energy Econ 25:39–64 526 Hoekstra R, Van Den Bergh JC (2002) Structural decomposition analysis of physical flows in the 527 economy. Environ Resour Econ 23:357-378 528 Hughes M, Palmer J, Cheng V, Shipworth D (2013) Sensitivity and uncertainty analysis of 529 England's housing energy model. Build Res Inf 41(2):156-67.

- Huijbregts M (2002) Uncertainty and variability in environmental life-cycle assessment. Int J
   Life Cycle Assess 7:173–173
- Huijbregts MA, Norris G, Bretz R, et al (2001) Framework for modelling data uncertainty in life
   cycle inventories. Int J Life Cycle Assess 6:127–132
- Igos E, Benetto E, Meyer R, et al (2019) How to treat uncertainties in life cycle assessment
   studies? Int J Life Cycle Assess 24:794–807
- ISO (2006) 14040: Environmental management–life cycle assessment–principles and framework. International Organization for Standardization.
- Jann B (2008) The Blinder–Oaxaca decomposition for linear regression models. Stata J 8:453–
   479
- Jenne CA, Cattell RK (1983) Structural change and energy efficiency in industry. Energy Econ
   5:114–123
- Jeong K, Kim S (2013) LMDI decomposition analysis of greenhouse gas emissions in the
   Korean manufacturing sector. Energy Policy 62:1245–1253
- 544 Jung S, An K-J, Dodbiba G, Fujita T (2012) Regional energy-related carbon emission
  545 characteristics and potential mitigation in eco-industrial parks in South Korea:
  546 Logarithmic mean Divisia index analysis based on the Kaya identity. Energy 46:231–241
- 547 Kako T (1978) Decomposition analysis of derived demand for factor inputs: The case of rice
   548 production in Japan. Am J Agric Econ 60:628–635
- 549 Lloyd SM, Ries R (2008) Characterizing, Propagating, and Analyzing Uncertainty in Life-Cycle
  550 Assessment: A Survey of Quantitative Approaches. J Ind Ecol 11:161–179.
  551 https://doi.org/10.1162/jiec.2007.1136
- Lo S-C, Ma H, Lo S-L (2005) Quantifying and reducing uncertainty in life cycle assessment
   using the Bayesian Monte Carlo method. Sci Total Environ 340:23–33
- Ma M, Yan R, Du Y, et al (2017) A methodology to assess China's building energy savings at
   the national level: an IPAT–LMDI model approach. J Clean Prod 143:784–793
- Malla S (2009) CO2 emissions from electricity generation in seven Asia-Pacific and North
  American countries: A decomposition analysis. Energy Policy 37:1–9.
  https://doi.org/10.1016/j.enpol.2008.08.010
- 559 Marlay RC (1984) Trends in industrial use of energy. Science 226:1277–1283

Meng Z, Wang H, Wang B (2018) Empirical Analysis of Carbon Emission Accounting and
 Influencing Factors of Energy Consumption in China. Int J Environ Res Public Health
 15:2467

- Mutel CL, de Baan L, Hellweg S (2013) Two-step sensitivity testing of parametrized and
   regionalized life cycle assessments: methodology and case study. Environ Sci Technol
   47:5660–5667
- 566 Oaxaca R (1973) Male-female wage differentials in urban labor markets. Int Econ Rev 693–709
- 567 OpenLCA (2018) User Manual. GreenDelta, Germany
- Pachauri RK, Allen MR, Barros VR, et al (2014) Climate change 2014: synthesis report.
  Contribution of Working Groups I, II and III to the fifth assessment report of the Intergovernmental Panel on Climate Change. IPCC
- 571 Patouillard L, Collet P, Lesage P, et al (2019) Prioritizing regionalization efforts in life cycle
  572 assessment through global sensitivity analysis: a sector meta-analysis based on ecoinvent
  573 v3. Int J Life Cycle Assess 24:2238–2254
- 574 Paul S, Bhattacharya RN (2004) CO2 emission from energy use in India: a decomposition
   575 analysis. Energy Policy 32:585–593
- 576 Qin Y, Cucurachi S, Suh S (2020) Perceived uncertainties of characterization in LCA: a survey.
  577 Int J Life Cycle Assess 25: 1846-1858.
- Qin Y, Suh S (2017) What distribution function do life cycle inventories follow? Int J Life Cycle
   Assess 22:1138–1145. https://doi.org/10.1007/s11367-016-1224-4
- Rosenbaum RK, Bachmann TM, Gold LS, et al (2008) USEtox—the UNEP-SETAC toxicity
   model: recommended characterisation factors for human toxicity and freshwater
   ecotoxicity in life cycle impact assessment. Int J Life Cycle Assess 13:532–546
- 583 Saltelli A, Annoni P, Azzini I, et al (2010) Variance based sensitivity analysis of model output.
  584 Design and estimator for the total sensitivity index. Comput Phys Commun 181:259–270
- 585 Saltelli A, Ratto M, Andres T, et al (2008) Global sensitivity analysis: the primer. John Wiley &
  586 Sons
- 587 SimaPro (2016) User Manual. PRé Consultants, Netherlands
- Sobol IM (2001) Global sensitivity indices for nonlinear mathematical models and their Monte
   Carlo estimates. Math Comput Simul 55:271–280
- Sonnemann GW, Schuhmacher M, Castells F (2003) Uncertainty assessment by a Monte Carlo
   simulation in a life cycle inventory of electricity produced by a waste incinerator. J Clean
   Prod 11:279–292
- 593 Stocker T (2014) Climate change 2013: the physical science basis: Working Group I contribution
   594 to the Fifth assessment report of the Intergovernmental Panel on Climate Change.
   595 Cambridge University Press

- Su B, Ang BW (2012) Structural decomposition analysis applied to energy and emissions: some methodological developments. Energy Econ 34:177–188
- Tarantola S, Gatelli D, Mara TA (2006) Random balance designs for the estimation of first order
   global sensitivity indices. Reliab Eng Syst Saf 91:717–727
- Timilsina GR, Shrestha A (2009) Factors affecting transport sector CO2 emissions growth in
   Latin American and Caribbean countries: an LMDI decomposition analysis. Int J Energy
   Res 33:396–414
- Wang H, Ang BW, Su B (2017) Assessing drivers of economy-wide energy use and emissions:
   IDA versus SDA. Energy Policy 107:585–599
- Wei W, Larrey-Lassalle P, Faure T, et al (2015) How to conduct a proper sensitivity analysis in
  life cycle assessment: taking into account correlations within LCI data and interactions
  within the LCA calculation model. Environ Sci Technol 49:377–385
- Weidema BP, Bauer C, Hischier R, et al (2013) Overview and methodology: Data quality
   guideline for the econvent database version 3. Swiss Centre for Life Cycle Inventories
- 610 Wernet G, Bauer C, Steubing B, et al (2016) The ecoinvent database version 3 (part I): overview
  611 and methodology. Int J Life Cycle Assess 21:1218–1230
- Ku C, Gertner G (2011) Understanding and comparisons of different sampling approaches for the
   Fourier Amplitudes Sensitivity Test (FAST). Comput Stat Data Anal 55:184–198
- Ye M, Hill MC (2017) Global sensitivity analysis for uncertain parameters, models, and
   scenarios. In: Sensitivity Analysis in Earth Observation Modelling. Elsevier, pp 177–210
- 616 York R, Rosa EA, Dietz T (2003) STIRPAT, IPAT and ImPACT: analytic tools for unpacking
   617 the driving forces of environmental impacts. Ecol Econ 46:351–365
- 618 Zhang M, Mu H, Ning Y, Song Y (2009) Decomposition of energy-related CO2 emission over
   619 1991–2006 in China. Ecol Econ 68:2122–2128