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Journal

Environmental Science and Technology, 58(49)

ISSN

0013-936X

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Publication Date

2024-12-10

DOI

10.1021/acs.est.4c06805

Peer reviewed

Pyrolysis of Two Perfluoroalkanesulfonates (PFSA) and PFSA-Laden Granular Activated Carbon (GAC): Decomposition Mechanisms and the Role of GAC

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Cite This: *Environ. Sci. Technol.* 2024, 58, 21850–21860



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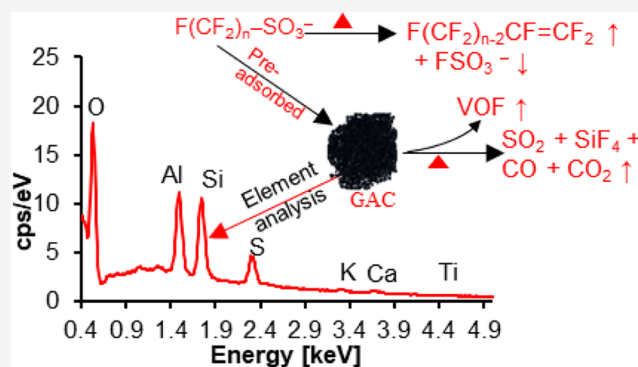
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ABSTRACT: Thermal treatment of perfluoroalkyl and polyfluoroalkyl substances (PFASs) presents a promising opportunity to halt the PFAS cycle. However, how co-occurring materials such as granular activated carbon (GAC) influence thermal decomposition products of PFASs, and underlying mechanisms remain unclear. We studied the pyrolysis of two potassium salts of perfluoroalkanesulfonates (PFSA, $C_nF_{2n+1}SO_3K$), perfluorobutanesulfonate (PFBS-K), and perfluorooctanesulfonate (PFOS-K), with or without GAC. PFBS-K is more stable than PFOS-K for pure standards, but when it is adsorbed onto GAC, its thermal stabilities and decomposition behaviors are similar. Temperatures and heating rates can significantly influence the decomposition mechanisms and products for pure standards, while these effects are less pronounced when PFSA are adsorbed onto GAC. We further studied the underlying decomposition mechanisms. Pure standards of $C_nF_{2n+1}SO_3K$ can decompose directly in their condense phase by reactions: $F(CF_2)_nSO_3K \rightarrow F(CF_2)_{n-2}CF=CF_2 + KFSO_3$ or $F(CF_2)_nSO_3K \rightarrow F(CF_2)_n^- + K^+ + SO_3$. GAC appears to facilitate breakage of the C–S bond to release SO_2 at temperatures as low as 280 °C. GAC promotes fluorine mineralization through functional reactive sites. SiO_2 is particularly important for the surface-mediated mineralization of PFASs into SiF_4 . These findings offer valuable insights into optimizing thermal treatment strategies for PFAS-contaminated waste.

KEYWORDS: PFBS, PFOS, PFAS, sorbent treatment, pyrolysis, surface-mediated decomposition, functional reactive site, silica



1. INTRODUCTION

Per- and polyfluoroalkyl substances (PFASs) are persistent man-made chemicals that have been widely detected in natural waters and industrial wastes,^{1–4} and have raised substantial concern for ecosystems and human health.^{5–9} When the source of drinking water is contaminated, PFASs cannot be efficiently removed in conventional water treatment plants.^{10,11} A few technologies have emerged to degrade PFASs for water treatment, such as electrochemical degradation,^{12,13} ultraviolet/sulfite/iodine,^{14–18} and plasma-based degradation;^{19–21} however, most of them have only been tested in laboratory settings or still in the effort of scaling up.^{22,23} PFAS adsorption from water by granular activated carbon (GAC) or ion exchange (IX) resins is one of the most established approach that can be deployed in full-scale because of its economic effects, compatibility, and regeneration potential.^{24–26}

When PFASs are concentrated on GAC, the spent GAC requires further treatment to prevent the PFASs from reentering the environment.^{27,28} Thermal reactivation is a well-established process to treat spent GAC,^{29–31} and most

PFASs can be removed in the thermal treatment.^{29,32–34} The reactivation involves two stages: (1) a pyrolytic process to remove and destroy adsorbates in an inert atmosphere; and (2) an oxidative process to reactivate the GAC by gasification reactions in steam or carbon dioxide atmosphere around 800 °C.^{30,35} Adsorbed PFASs can be thoroughly removed (>99%) and partial mineralization of fluorine (30–72%) can be obtained in the pyrolytic stage.^{36,37} Perfluoroalkanesulfonates (PFSA) require higher temperatures (400 °C) than perfluoroalkyl carboxylic acids (PFCAs) and emerging carboxylate structures such as GenX.^{32,34} Although PFSA are one of the most common and thermally resistant PFAS categories found in the environment, current mechanistic

Received: July 5, 2024

Revised: October 31, 2024

Accepted: November 18, 2024

Published: November 26, 2024



understanding about the thermal treatment of PFASs is limited.

Prior studies of the thermal decomposition of PFASs focused on the acidic form ($C_nF_{2n+1}SO_3H$) as pure standards at specific temperatures. For example, pyrolysis of perfluorooctanesulfonic acid (PFOS-H) pure standards has been reported to release HF, SO_2 , COF_2 , C_2F_4 , and perfluorooctanoyl fluoride ($C_7F_{15}COF$) as major products in computational and experimental studies.^{38–41} However, earlier understandings from $C_nF_{2n+1}SO_3H$ may not be directly applicable to the anions ($C_nF_{2n+1}SO_3^-$) existing in the environment due to the dissociation caused by pH, considering the low pK_a (typically <2) of PFASs.⁴² In addition, GAC could alter the PFAS thermal decomposition mechanisms and enhance the mineralization rates in the thermal treatment processes,^{33,37} yet the underlying mechanisms have not been fully understood. Furthermore, thermal treatment of GAC involve the ramp heating step, but the thermal behaviors of PFASs during ramp heating process have not been studied.^{27,30} Finally, there has been little experimental research on the thermal decomposition of shorter-chain PFASs, such as PFBS, leaving a significant knowledge gap regarding the influence of the perfluorocarbon chain on thermal decomposition mechanisms.

The objectives of this study are to investigate: (1) the thermal decomposition products and mechanisms of deprotonated PFASs using a combination of experimental and computational approaches; (2) the thermal behaviors of pure PFSA standards and PFSA-laden GAC to understand the role of GAC in PFSA thermal decomposition and mineralization; (3) the influence of temperature and heating methods (e.g., ramp heating or isothermal conditions used in GAC regeneration) on product formation; and (4) the influence of perfluorocarbon chain length on PFSA thermal decomposition. The results of this study contribute to an improved understanding of thermal treatment for PFAS remediation, which can aid in the development of more cost-effective and sustainable strategies for PFAS waste management.

2. MATERIALS AND METHODS

2.1. Chemicals. Four PFASs were used in this study: perfluorobutanesulfonic acid (PFBS-H, 98%, TCI America), potassium perfluorobutanesulfonate (PFBS-K, 98%, TCI America), perfluorooctanesulfonic acid (PFOS-H, 95%, Matrix Scientific), and potassium perfluorooctanesulfonate (PFOS-K, 98%, Matrix Scientific). Potassium fluorosulfate ($KFSO_3$, 99.5%) was purchased from Thermo Scientific. FILTRASORB 400 GAC (Calgon Carbon Corporation, Pittsburgh, PA) was used in this study for its high PFAS sorption capability and common use in commercial systems.^{27,43–45} PFSA-laden GAC samples were prepared by bench adsorption experiments. Detailed chemical information and preparations of GAC samples can be found [Supporting Information](#) and in [Table S1](#).

2.2. Experimental Methods. Samples were ramp heated or heated in isothermal conditions. Ramp heating process specifically means that the furnace and sample are heated simultaneously from room temperature with a rate of $10\text{ }^\circ\text{C min}^{-1}$, while heating in isothermal conditions means a furnace is preheated and maintained to 500 or 600 $^\circ\text{C}$ before and after samples are introduced in this study. Ramp heating of ~ 10 mg pure PFASs standards or ~ 30 mg PFSA-laden GAC in aluminum crucibles was conducted with a simultaneous thermal analyzer (STA, NETZSCH STA 449 F1 Jupiter) equipped with a silicon carbide furnace at Iowa State

University. The temperatures increased from 40 to 1100 $^\circ\text{C}$, with a rate of $10\text{ }^\circ\text{C min}^{-1}$. Thermogravimetry (TGA) and differential scanning calorimetry (DSC) data were acquired simultaneously. The STA was coupled with a quadrupole mass spectrometer (NETZSCH Aeolos QMS 403 D) and a Fourier-transform infrared (FTIR, Bruker ALPHA II) spectrometer to monitor evolved gaseous products that were correlated with TGA and DSC signals. In the following sections, STA-MS/FTIR refers to the simultaneous thermal analysis of TGA/DSC and analysis of gaseous products with MS/FTIR. Additional experiments were conducted to confirm the reproducibility of STA-MS/FTIR analysis (SI). An MTI tube furnace (OTF-1200X) equipped with a ceramic tube was also used for additional ramp heating experiments to exclude the interference of silicon. The detailed parameters of the thermal treatment instrument can be found in [Figure S1](#).

To aid in the interpretation of the STA-MS/FTIR results, we conducted pyrolysis experiments under isothermal conditions (preheated at 500 and 600 $^\circ\text{C}$) using an electric tube furnace modified from a Shimadzu solid sample module (SSM-5000A). SSM-5000A was fitted with a quartz tube or stainless steel 316 tube, which was flowed by ultrahigh purity N_2 at a flow rate of 100 mL min^{-1} . An FTIR spectrometer (Nicolet Magna-IR 560) at a resolution of 2 cm^{-1} equipped with a 2 m optical-path gas cell was connected to the outflow of furnaces to monitor the evolved gases in real time. The system was purged by N_2 more than 1 h at 500 or 600 $^\circ\text{C}$ and monitored by FTIR to exclude the interferences of impurities or carryovers before pyrolysis experiments started. The experiments were conducted two to three times to confirm the findings. Gaseous products were identified through spectral comparison with standard references from published papers or databases.^{46–50} The OMNIC software (Thermo Scientific) was used to perform FTIR spectra subtraction. Qualitative analysis for pyrolysis residuals was conducted with attenuated total reflectance FTIR (ATR-FTIR, ThermoFisher Scientific Nicolet iS50) and ultrahigh-resolution mass spectrometry (UHPLC-HRMS, Q Exactive Plus, Thermo Scientific). Details are provided in [Supporting Information](#). The UHPLC-HRMS instrumental settings and product identification procedures have been published elsewhere.⁵¹

Scanning electron microscopy (SEM, TESCAN, Vega3 SBH) coupled with energy dispersive X-ray spectroscopy (EDS, Bruker, XFlash7T) was used to characterize the elemental composition of the GAC surfaces pre- and postpyrolysis experiments.

2.3. Calculation of Theoretical Potential Energy Profiles. Thermal degradation mechanisms of PFBS and PFOS were studied through quantum chemical calculations based on density functional theory (DFT) and transition state theory (TST). The influence of the crystal structures on the decomposition is not considered in this study. The suite of Gaussian 16 programs was used to optimize the geometries of reactants, products, intermediates, and transition states using the Boese and Martin's τ -dependent hybrid functional (BMK)⁵² and 6–31+G(2d,p) basis set.⁵³ Frequency analysis was conducted by the same method, and the electronic energies were computed using double-hybrid functional DSD-PBEP86,⁵⁴ in combination with Grimme's third-generation empirical dispersion implemented by Becke-Johnson damping (D3BJ),⁵⁵ and the triple- ζ basis set def2-TZVP.⁵⁶ Intrinsic reaction coordinate calculations were carried out to verify the transition state connectivity.⁵⁷ Previous studies have demon-

strated that the calculated decomposition pathways using these functional and basis sets are satisfactory^{41,58} and consistent with experimental results.^{38–40,59} Instead of targeting the precise values of energy barriers, our computation can qualitatively point out feasible pathways during the thermal decomposition of PFASs. When PFBS/PFOS were preadsorbed onto GAC, their decomposition pathways were primarily derived from experimental results rather than computation, given that the DFT calculations were performed intrinsically in vacuum without any interaction with surrounding environments.

3. RESULTS AND DISCUSSION

3.1. Pyrolysis of Pure PFBS-K and PFOS-K Standards.

3.1.1. Simultaneous Thermal Analysis and Gaseous Products in Ramp Heating. Pure standards of ~10 mg PFBS-K/PFOS-K were heated from 40 to 1000 °C for STA-MS/FTIR analysis. The TGA and DSC curves, mass spectra, and FTIR at 488 °C of gaseous products are shown in Figure 1. The TGA curve of PFBS-K did not significantly change until

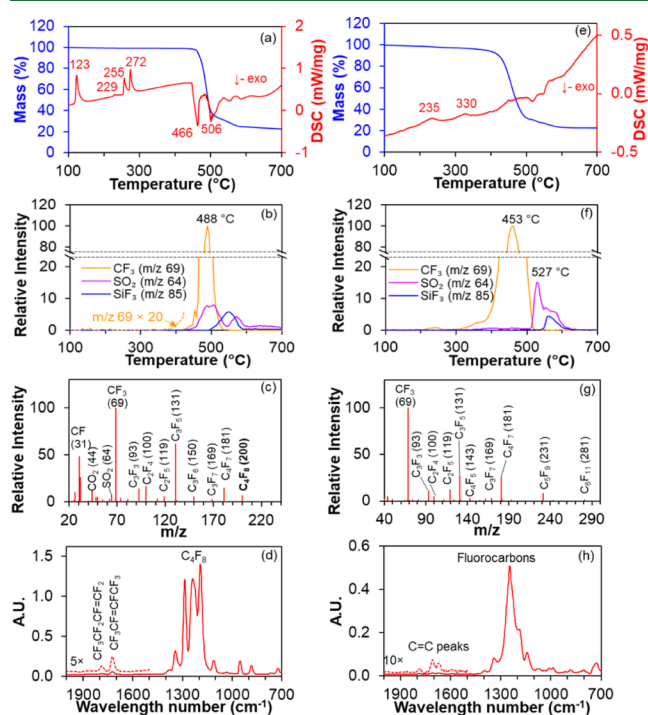


Figure 1. STA-MS/FTIR analysis of PFBS-K (a–d) and PFOS-K (e–h), respectively: (a) and (e) TGA/DSC curve, (b) and (f) formation trends of selected gaseous products represented by EI fragments, (c) and (g) mass spectrum of gaseous products at 488 °C EI (70 eV), and (d) and (h) FTIR spectrum (resolution, 4 cm⁻¹) of gaseous products at 488 °C for PFBS-K and 453 °C PFOS-K, respectively. Reference FTIR spectra of CF₃CF₂CF=CF₂ and CF₃CF=CFCF₃ can be found in published literature.^{46,60} More mass spectra and FTIR spectra can be found in Figure S2 (PFBS-K) and Figure S5 (PFOS-K).

430 °C (98.92% mass left). However, PFBS-K started to yield organofluorine gases at temperatures as low as ~380 °C, as indicated by the increase in the CF₃ (*m/z* 69) signal in gaseous products detected by MS (Figure 1b). Thus, PFBS-K can be considered thermally unstable above 380 °C.

The significant mass loss (61.45%) of PFBS-K between 448 °C (98.46% remaining) and 506 °C (37.01% remaining) indicates that thermal decomposition occurred in its

condensed phase and decomposed into gaseous products before reaching the boiling point (transitioning into the gas phase). At this stage, the DSC curve shows a distinct exothermic peak at 466 °C, which further excludes the evaporation of PFBS-K because evaporation is an endothermic process. The observed mass loss can be attributed to the formation of gaseous products, such as organofluorine (represented by CF₃ *m/z* 69) and SO₂ (*m/z* 64), as shown in Figure 1b and 1c. By comparing with reference FTIR spectra published in literature,^{46,60} the products at 488 °C include CF₃CF=CFCF₃ (2-C₄F₈) and CF₃CF₂CF=CF₂ (1-C₄F₈) (Figure 1d). Additional FTIR and mass spectra obtained at various temperatures are shown in Figure S2. The identified products from FTIR spectra were further supported by MS. For example, all EI fragments of 1-C₄F₈ and 2-C₄F₈, such as *m/z* 200 (C₄F₈⁺), *m/z* 181 (C₄F₇⁺), and *m/z* 131 (C₃F₅⁺) were detected and agreed well with the NIST references (Figure S3).

The second stage of mass loss, occurring between temperatures of 506 °C (37.01% remaining) and 588 °C (24.74% remaining), exhibits a decrease in rate compared to the initial stage. The DSC curve suggests that the second stage consisted of multiple processes, characterized by indistinguishable peaks between the temperature from 506 to 650 °C. The gaseous products detected by MS and FTIR (Figure S2) indicated the presence of SO₂F₂ (Figure S5), SiF₄, and SO₂ (Figure 1b). One possible explanation is that the second stage involves the further decomposition of residual products such as KFSO₃ (see Sections 3.1.2 and 3.2.2) from the first stage when temperatures increased. This is evidenced by the shift of the dominant gaseous products from organofluorine or olefins to inorganic gases such as SO₂ and SiF₄. SiF₄ is likely produced through the reaction of intermediates with SiC, which is the reactor material of the tube furnace. SiF₄ is frequently observed during the thermal decomposition of PFASs when silicon-based materials such as quartz and glass are present.^{49,61,62} Thus, the thermal decomposition of PFBS-K does not follow a single pathway but rather involves multiple pathways and processes.

The TGA curve of PFOS-K (Figure 1e) exhibits similarities to that of PFBS-K. The first stage of mass loss occurred between temperatures of 400 and 506 °C, resulting in the release of organofluorine gases, including olefins (Figure 1h). Both the FTIR spectrum (Figure 1h) and mass spectrum (Figure 1g) can demonstrate that the released gases were fluorocarbons. At 448 °C, two peaks at 1782.79 and 1709.78 cm⁻¹ in FTIR spectra, corresponding to C=C bonds (Figure 1h), most likely represent formations of perfluoro-1-octene (CF₃(CF₂)₅CF=CF₂, 1-C₈F₁₆), and perfluoro-2-octene (CF₃(CF₂)₄CF=CFCF₃, 2-C₈F₁₆) (S4), similarly with decomposition of PFBS-K (Table 1). The second stage of mass loss occurred between 506 and 600 °C, releasing inorganic gases such as SO₂, SiF₄, and SO₂F₂ (Figure S4). At temperatures exceeding 500 °C, the primary gaseous products were SO₂ and SiF₄, while organofluorine products were no longer detected (Figure 1f). Similar trends of TGA curves and decomposition products (e.g., corresponding olefins, SO₂, SiF₄) imply that PFOS-K underwent similar decomposition processes as PFBS-K.

However, the thermal decomposition of PFOS-K also exhibits unique characteristics (Table 1). First, the onset of PFOS-K thermal decomposition occurs at lower temperatures (300 °C) compared to that of PFBS-K (380 °C). The DSC peaks at ~330 °C (Figure 1e) and associated VOF detection

Table 1. Comparisons of the Thermal Behaviors of PFBS and PFOS at Different Conditions

Conditions	Similarities	Differences
PFBS-K vs PFOS-K	Trends of TGA curves; Decomposition pathways (reaction 1) and products (olefins, SO ₂ , SiF ₄ , C ₂ F ₄ , KFSO ₃ , SO ₂ F ₂).	PFOS-K is less thermally stable; PFBS-K has multiple endothermic peaks before mass loss; Different decomposition mechanism to form SO ₂ ; Pyrolysis of PFOS-K produced C ₂ F ₆ at 600 °C.
GAC laden with PFBS vs PFOS	TGA curves; Exothermic reactions at ~448 °C; Thermal stabilities; Decomposition products (olefins, SO ₂ , SiF ₄ , C ₂ F ₄).	PFOS yields longer perfluorocarbon chain products.
PFSA-K vs PFSA-laden GAC	Decomposition products (olefins, SO ₂ , SiF ₄ , C ₂ F ₄).	GAC promotes pyrolysis of PFSA to (1) yield more SO ₂ but less VOF; (2) form SiF ₄ ; (3) produce hydrofluorocarbons; (4) reduce their thermal stabilities; (5) change decomposition mechanisms and thermal behaviors.

(e.g., CF₃⁺, *m/z* 69, Figure 1f) indicate the formation of gaseous products. In addition, the DSC curve for PFOS-K did not show any discernible exothermic peaks during the initial stage of mass loss. Furthermore, in contrast to PFBS-K, the pyrolysis of PFOS-K results in asynchronous formation peaks of VOF (e.g., CF₃⁺ *m/z* 69) and SO₂ (*m/z* 64), with the later formation peak of SO₂ observed at 527 °C (Figure 1f). PFBS-K and PFOS-K belong to PFSA with similar molecular structures, which were tested under the same conditions. Different thermal behaviors, including different thermal stabilities, DSC curves, and timings of produce gaseous products, were observed in ramp heating. One possible explanation is that crystal structures⁶³ caused by perfluorocarbon chain lengths (C4 vs C8) can influence thermal behaviors of potassium salts of PFSA.

3.1.2. Product Analysis at 500 and 600 °C (Isothermal Conditions). During the ramping heating process, PFBS-K and PFOS-K were decomposed below 500 °C. However, the effects of the temperatures on the resulting decomposition products remain unclear, as the heating rate may alter the kinetic behavior and influence the pyrolysis process. We conducted additional research on the pyrolysis of PFBS-K and PFOS-K under isothermal conditions at 500 and 600 °C, respectively. FTIR was used to monitor the gaseous products, with the corresponding spectra presented in Figure 2. Based on the FTIR spectra, we identified a series of products: (1) at 500 °C, pure PFBS-K produced 1-C₄F₈ (C=C stretching at 1792.69 cm⁻¹), 2-C₄F₈ (C=C stretching at 1730.37 cm⁻¹), C₂F₄, SO₂, and some unknown compounds; (2) at 600 °C, the primary product of PFBS-K was CF₃CF=CFCF₃, with minor amounts

of C₂F₄ and SO₂ also being produced. Similar products were also detected during the pyrolysis of PFOS-K: (1) at 500 °C, a C=C peak at 1792.08 cm⁻¹ and a C=C peak at 1718.60 cm⁻¹ (Figure 2c) suggested the formation of 1-C₈F₁₆ and 2-C₈F₁₆ (S4), respectively; (2) SO₂ and C₂F₄ were detected at 500 and 600 °C. The formations of 1-C₈F₁₆ and 2-C₈F₁₆ were also confirmed in another recent publication.⁶⁴ In addition, formations of hexafluoropropene (C₃F₆), perfluorobutane (C₄F₁₀), perfluoro-1-heptene(1-C₇F₁₄), perfluoro-1-hexene (1-C₆F₁₂), perfluoro-2-hexene (1-C₆F₁₂) in pyrolysis of PFOS-K were also reported.⁶⁴

Notably, the decomposition products of PFOS-K can be significantly influenced by the temperature changes. New products such as C₂F₆ and some unknown perfluorocarbons were identified at 600 °C (Figure 2d). The formation of C₂F₆ and C₂F₄, small molecules, provided strong evidence of C–C cleavage and formation of perfluorocarbon radicals in thermal decomposition. The variations in products demonstrate the significant impact of temperatures on decomposition mechanisms.

After 2 h of pyrolysis of PFBS-K and PFOS-K at 500 °C, black char residuals were found in the aluminum crucibles. Another study found black chars in the pyrolysis of octafluorocyclobutane, which were assumed to be carbon.⁶¹ ATR-FTIR was used to analyze the products. We found that residuals (74.2% mass left) from PFBS-K pyrolysis contained undecomposed PFBS-K and some new organofluorine, whereas black char (68.6% mass left) from PFOS-K contained K₂SO₄ (Figure S8). The residual products were also dissolved in methanol, diluted with water, and analyzed by UHPLC-HRMS. Undecomposed PFBS and PFOS were found in residuals at 500 °C, showing that 2 h at this temperature are insufficient to completely destroy PFSA potassium salts. KFSO₃ was detected in the pyrolysis of both PFBS-K and PFOS-K at 500 °C (Figure S9). FSO₃⁻ could be a common product, along with the formation of perfluoro-1-alkene in the pyrolysis of PFSA salts by reaction 1. When the two compounds (5–7 mg) were pyrolyzed at 600 °C for 1 h, less than 1 mg white residuals were found in aluminum crucibles. Undecomposed PFOS and PFBS were detected again from residuals by UHPLC-HRMS. FSO₃⁻ was detected only in the residuals of PFOS-K but not in PFBS-K at 600 °C.



3.2. Pyrolysis of PFBS- or PFOS-Laden GAC Samples.

3.2.1. Simultaneous Thermal Analysis and Gaseous Products in Ramp Heating. To understand the role of GAC in the thermal decomposition of PFBS and PFOS on spent GAC samples, we prepared ~30 mg of GAC equilibrated with either PFBS or PFOS for STA-MS/FTIR analysis. The TGA/DSC curves and gaseous products are shown in Figure 3. Based

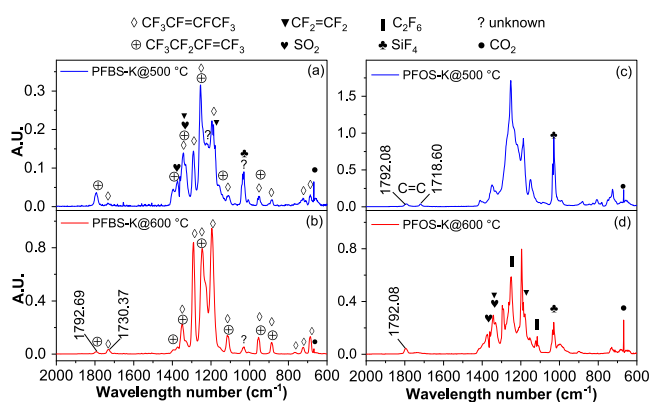


Figure 2. FTIR spectra (resolution, 2 cm⁻¹) of gaseous products in the pyrolysis of (a) PFBS-K at 500 °C, (b) PFBS-K at 600 °C, (c) PFOS-K at 500 °C, (d) PFOS-K at 600 °C. Reference FTIR spectra of CF₃CF₂CF=CF₂ and CF₃CF=CFCF₃ can be found in published literature.^{46,60} Subtracted spectra from features attributable to C₂F₄ (CF₂=CF₂) and C₂F₆ can be found in Figures S6 and S7.

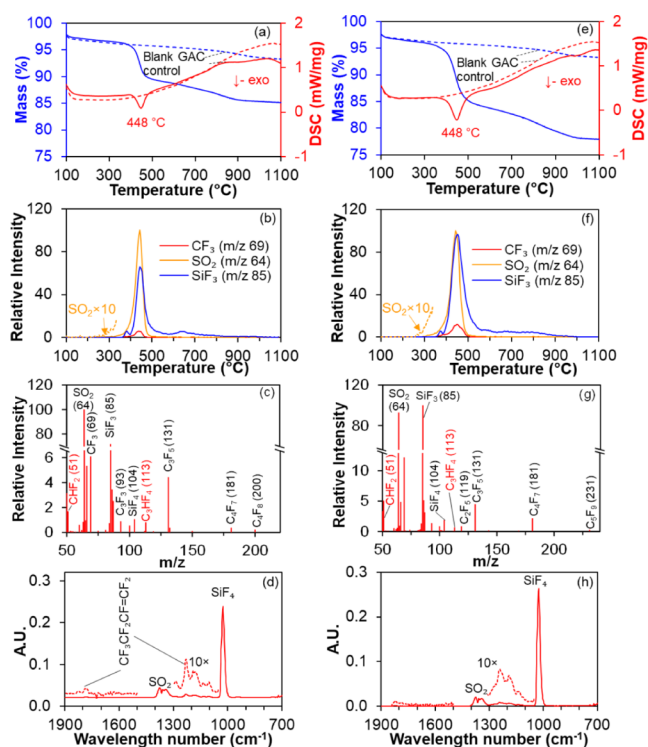


Figure 3. STA-MS/IR of GAC laden with PFBS (a–d) or PFOS (e–h). (a) and (e) TGA/DSC curve (dashed lines represent blank GAC); (b,f) formation trends of selected gaseous products represented by EI fragments; (c,g) mass spectrum of gaseous products at 448 °C (EI, 70 eV); and (d,h) FTIR spectrum (resolution, 4 cm^{-1}) of gaseous products at 448 °C for GAC laden with PFBS and PFOS, respectively. More mass spectra and FTIR spectra at different temperatures can be found in Figures S10–11. The gaseous products of blank GAC pyrolysis can be found in Figure S12.

on the TGA curves of PFBS or PFOS-laden GAC samples, the first stage of mass loss related to PFSA is between 280 °C (the temperature of SO_2 being detected) and 470 °C, becoming faster after 400 °C with an exothermic DSC peak at 448 °C (Figure 3a and e). From 280 to 470 °C, the mass losses were 6.59% and 9.59% for PFBS- and PFOS-laden GAC, which is consistent with the preadsorbed PFBS (6.34%) and PFOS (11.76%) mass concentration, respectively. Evolved gaseous products include SO_2 , SiF_4 , CO_2 , CO , and minor organofluorine compounds (represented by CF_3^+ m/z 69) (Figure 3c, d, g, and h). Two EI fragments, m/z 51 and m/z 113 (Figure 3c, g), can be assigned to CHF_2^+ and C_3HF_4^+ , implying that hydrogenation may occur during the pyrolysis of PFBS- or PFOS-laden GAC. Notably, CHF_2^+ was also reported as a gaseous product when PFOS was thermally treated with a sludge matrix.⁶⁵ The second stage of mass loss occurred between 470 and 900 °C (85.62% remaining for PFBS-laden GAC and 79.45% for PFOS-laden GAC), with significant release of SiF_4 . When compared to pure PFSA-K, the second stage is much longer for GAC samples, suggesting GAC may retain or adsorb fluorine products produced in the first decomposition stage. In contrast to pure PFSA potassium salt standards, the TGA/DSC curves and major decomposition products (i.e., olefins, SiF_4 , and SO_2) for PFBS- or PFOS-laden GAC samples are similar, which suggests that the effects of perfluorocarbon chain length to the thermal behaviors of PFSA are less pronounced when PFSA are adsorbed onto GAC.

The release of SO_2 indicates that PFBS and PFOS start to decompose at temperatures as low as 280 °C when being adsorbed onto GAC. Two new signals, m/z 81 (possibly HO-SO_2^+) and m/z 96 (possibly SO_4^+), were also detected in the initial decomposition stage of both PFBS- and PFOS-laden GAC (Figure S12). How m/z 81 and m/z 96 were formed is not clear but could be related to the initialization of PFSA decomposition. The signals of m/z 81 and m/z 96 attained a peak at 363 °C, and their FTIR spectra are presented in Figures S10 and S11. m/z 81 and m/z 96 were not detected in STA-MS/FTIR of blank GAC, while low levels of SO_2 were observed above 350 °C (Figure S13).

3.2.2. Product Analysis at 500 and 600 °C (Isothermal Conditions). The temperatures for the pyrolysis of PFSA-laden GAC were set to 500 or 600 °C to be consistent with conditions for pure standards, allowing us to examine the effects of GAC by comparing their decomposition products. When PFBS or PFOS was preadsorbed onto GAC and pyrolyzed at 500 or 600 °C, considerable amounts of SO_2 were produced, and a large amount of organofluorine was converted into SiF_4 (Figure 4). 1- C_4F_8 and 2- C_4F_8 , C_2F_4 , SO_2 , and SiF_4 ,

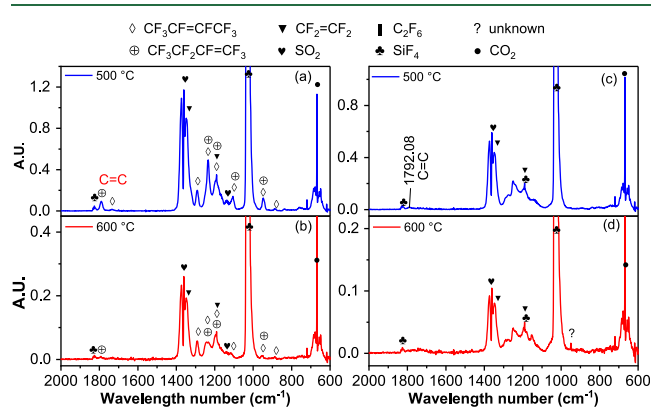


Figure 4. FTIR spectra (resolution, 2 cm^{-1}) of gaseous products in the pyrolysis of GAC laden with (a) PFBS at 500 °C, (b) PFBS at 600 °C, (c) PFOS at 500 °C, and (d) PFOS at 600 °C. Subtracted spectra from features attributable to SiF_4 , SO_2 , and C_2F_4 ($\text{CF}_2=\text{CF}_2$) can be found in Figures S14 and S15.

detected in the pyrolysis of pure PFBS-K, were also found in the pyrolysis of PFBS-laden GAC (Figure S14). 1- C_4F_8 underwent a transformation to 2- C_4F_8 when heated to 600 °C (Figure 4b), which is demonstrated by the FTIR spectrum of 1- C_4F_8 (Figure S16) obtained by subtracting the gaseous products at 600 °C from those at 500 °C. This clearly indicated that the difference of gaseous products in the pyrolysis of PFBS-laden GAC at 600 °C versus 500 °C is 1- C_4F_8 . The VOF products except C_2F_4 of PFOS-laden GAC could not be identified based on FTIR spectrum (Figure 4c), only a $\text{C}=\text{C}$ bond was found at 500 °C.

Temperatures have less influence on the decomposition products when PFOS and PFBS are adsorbed into GAC. Based on FTIR spectra, the distribution of gaseous products was more different in pyrolysis of pure PFBS/PFOS-K (Figure 2) than PFBS/PFOS-laden GAC (Figure 4) between 500 and 600 °C. For example, more C_2F_4 was produced at 500 °C than 600 °C in the pyrolysis of PFBS-K, and C_2F_6 was detected at 600 °C but was not detected at 500 °C. These differences were not observed in the pyrolysis of PFBS/PFOS-laden GAC. One possible reason is that GAC altered the decomposition

mechanisms when PFBS or PFOS are adsorbed onto the GAC surface. Another reason could be because crystal structures of salt can influence the decomposition pathways.

After PFBS- or PFOS-laden GAC were pyrolyzed for 1 h at 500 or 600 °C, the GAC was extracted by ammonium acetate (NH₄Ac) methanol solution (S2) but no PFBS/PFOS or FSO₃⁻ was detected by UHPLC-HRMS, which is consistent with the fact that GAC can alter and facilitate decomposition of PFASs.

3.3. Influence of GAC on Fluorine Conversion to SiF₄

When we compared the thermal decomposition products of pure PFBS/PFOS standards and PFBS/PFOS-laden GAC samples (Table 1), we noticed clear differences in the distribution of the products, particularly in the formation of SiF₄. The presence of GAC resulted in a decrease in the temperature required for SiF₄ formation from 500 (Figure 1) to 350 °C (Figure 3) during the STA-MS/FTIR analysis. One potential explanation is that GAC contains silicon to facilitate the conversion of PFASs into SiF₄.

To investigate the role of GAC in the formation of SiF₄, PFOS-K and fresh GAC were externally mixed (rather than adsorption) in a ceramic boat and ramp heated in a ceramic tube furnace from 100 to 900 °C at a rate of 10 °C/min in the N₂ atmosphere. A ceramic tube was used to exclude silicon interference from the reactor materials. FTIR was used to monitor gaseous products in real time, and the resulting spectra are displayed in Figure S17. The gaseous decomposition products (e.g., SO₂ and VOF) were detected around 475 °C, and then SiF₄ was formed at around 525 °C, which is consistent with the phenomena of pure PFOS-K pyrolysis in STA-MS/FTIR analysis (Figure 1f). The intensity of SiF₄ signal became much higher than VOF from ~525 to 900 °C, which is similar to the pyrolysis of PFOS-laden GAC in STA-MS/FTIR analysis (Figure 3f). The pyrolysis of mixture and PFOS-K confirms the finding that GAC can influence the distribution of products in pyrolysis of PFASs. During the ramp heating of the PFOS-K/GAC mixture, PFOS-K first decomposed into VOF including perfluorinated olefins, and then parts of gaseous products were released. Subsequently, some perfluorinated olefins could undergo reactions with quartz on the GAC surface to form SiF₄, which is consistent with the phenomenon discussed in Section 3.2. An additional experiment was conducted using a stainless-steel 316 tube at 500 °C, and SiF₄ was detected again by FTIR, demonstrating the GAC-mediated surface interactions for SiF₄ formation.

SEM-EDS analysis revealed that F400 GAC contains elements O, Al, Si, S, K, Ca, and Ti, in addition to C (Figure 5). Based on the safety data sheet (SDS) of F400 GAC, F400 contains quartz, titanium dioxide, aluminum oxide, calcium sulfate, and cobalt. These compounds onto GAC could

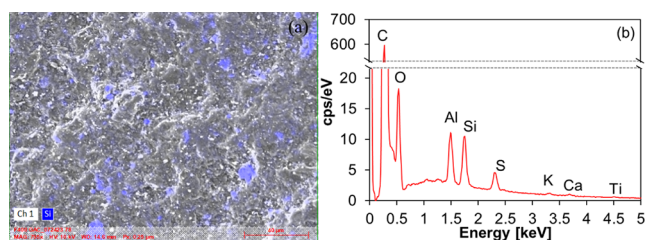


Figure 5. (a) SEM images of Si mapping of F400 GAC, and (b) Elemental analysis of clean F400 GAC surface by SEM-EDS.

potentially act as catalysts for the decomposition⁶⁶ or facilitate the conversion of PFASs into inorganic fluorine such as SiF₄. Sasi and Xiao et al. found that nonactivated charcoals/biochars did not accelerate thermal decomposition of PFASs, indicating that insignificant role of π electron-rich, polyaromatic effects of carbon on pyrolysis of PFASs.³³ The detailed mechanisms contributing to the promotion of PFSA breakdown remain unclear. However, our study confirms that quartz facilitates the mineralization of PFASs by generating SiF₄.

In the ramp heating of mixture of PFOS-K and GAC, SO₂F₂ was found above 600 °C, and CF₄ and C₂F₆ were found above 800 °C (Figure S17). These two recalcitrant gaseous products^{67–69} appear to be prevalent at high temperatures (>700 °C) in pyrolysis, which was also observed in the pyrolysis of perfluorocarboxylic acids at 780 °C.⁴⁹

3.4. Comparative Analysis and Decomposition Mechanisms. **3.4.1. Comparisons of Thermal Stabilities and Behaviors.** While thermal stabilities of pure PFASs have been previously reported,^{33,34,70} we found that thermal stabilities and behaviors of PFASs vary under different conditions. The similarities and differences of PFASs under different conditions are summarized in Table 1. PFBS-K and PFOS-K were stable below 380 and 300 °C for pure standards, respectively. However, when they were adsorbed onto GAC, thermal decomposition and the release of SO₂ occurred at temperatures as low as 280 °C for both of them. These observations highlight the significance of co-occurring materials in modifying the thermal degradation of PFASs, which needs to be considered when thermally treating PFAS-contaminated wastes.

Heating methods can influence the thermal behavior of the pure PFSA salts. During the ramp heating, both PFBS-K and PFOS-K lost more than 60% mass before reaching 500 °C in less than 50 min (Figure 1), but PFBS-K and PFOS-K only lost 25.7% and 41.4% mass in a constant heating process at 500 °C in 2 h, respectively. Faster mass loss in ramp heating could be because PFASs-K are a solid–solid phase change materials that can change its crystalline structure from one lattice configuration to another as temperature increases, generating higher-energy structures,⁷¹ as supported by endothermic peaks in the DSC curve without mass change (Figure 1). Higher-energy structures are easier to decompose, resulting in a faster mass loss in ramp heating.

Vaporization is common in thermal treatment. However, it is more probable for deprotonated PFBS and PFOS to undergo decomposition rather than vaporization. This is due to the low pK_a (<2) values of PFASs,⁴² indicating that PFASs exist in an ionized form in the environmental matrices. Protonated PFBS-H and PFOS-H have a higher propensity to undergo vaporization and a transition into the gas phase. During the ramp heating, these two neutral compounds evaporated before reaching a temperature of 201 °C (Figure S18), considering PFASs-H are thermally stable below 200 °C.^{32–34,34–40,72}

3.4.2. Decomposition Mechanisms. To gain a better understanding of potential thermal decomposition mechanisms, we performed DFT- and TST-based quantum chemical calculations to evaluate the potential surface energies for the unimolecular thermal decomposition of PFSA-K (PFBS-K shown in Figure 6a as an example and PFOS decomposition shown in Figure S19). As shown in Figure 6b, we proposed three decomposition pathways for the PFSA potassium salt (F(CF₂)_n–SO₃K) based on computationally derived mechanisms and the identified thermal decomposition products

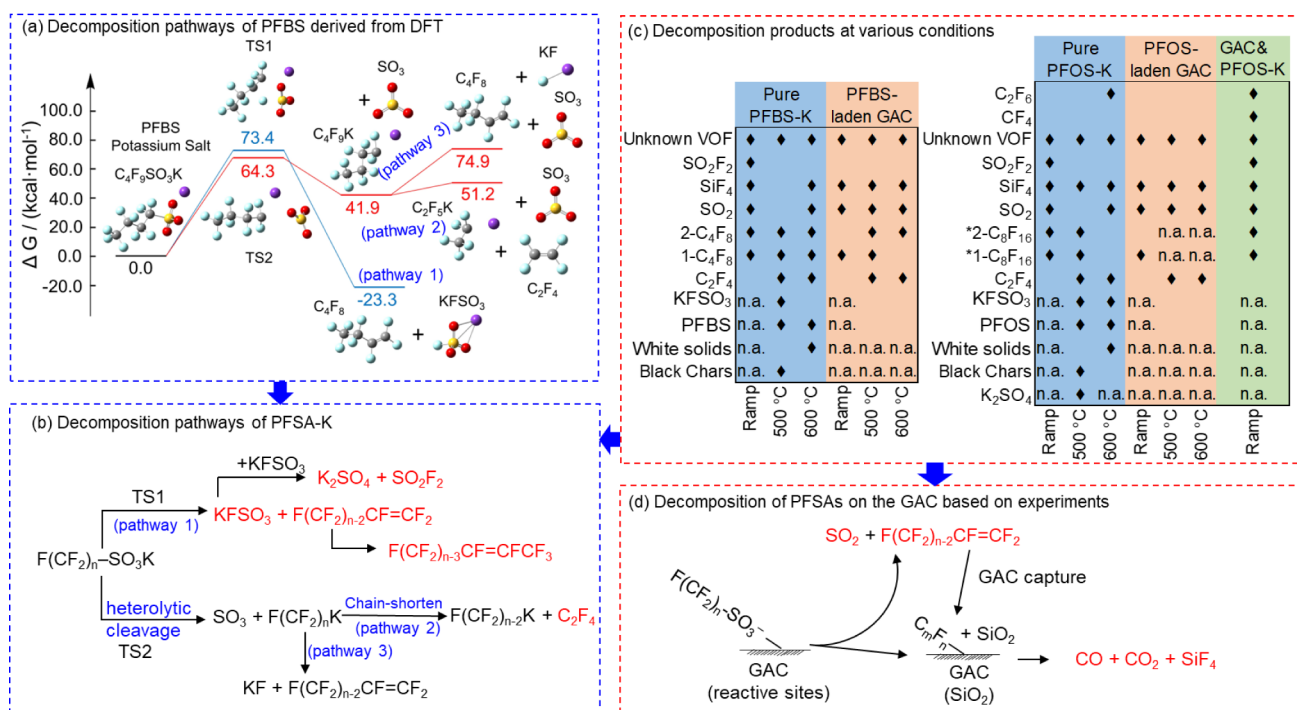


Figure 6. (a) Calculated Gibbs free energy barriers of thermal degradation of PFBS potassium salt at 800 K as an example, (b) unimolecular decomposition mechanisms of deprotonated PFSA-K derived from both DFT calculations and experimental results, (c) decomposition products of PFBS/PFOS in various conditions, and (d) the decomposition of PFSA-K on the surface of GAC (derived from experimental results). Final products identified by experiments ($n = 4, 8$) are highlighted by red texts. See S4.

(Figure 6c). The first pathway is an exothermic elimination reaction by the transition state (TS1) to yield perfluoro-1-alkene ($F(CF_2)_{n-2}CF=CF_2$) and $KFSO_3$ (pathway 1). The exothermic reaction is also supported by the experimental results, in which the formation of $CF_3CF_2CF=CF_2$ along with a DSC exothermic peak was found in pyrolysis PFBS-K. $F(CF_2)_{n-2}CF=CF_2$ is not thermally stable and can be converted into perfluoro-2-alkene ($F(CF_2)_{n-3}CF=CF_2$) (Figure S20). Residual $KFSO_3$ can be thermally decomposed into K_2SO_4 and SO_2F_2 ,⁷³ which are further demonstrated by the experimental detection (Figure 6c). Another competing channel is the dissociation of the C–S bond, which is a heterolytic cleavage via a variational transition state (TS2) (Figure 6), resulting in $F(CF_2)_nK$ and SO_3 . $F(CF_2)_nK$ can be shortened by pathway 2 (C–C cleavage) with the release of $CF_2=CF_2$, a product detected in the pyrolysis of both PFBS and PFOS. Decomposition of $F(CF_2)_nK$ by pathway 3 (C–F cleavage) is another channel to yield $F(CF_2)_{n-2}CF=CF_2$ (Figure 6a).

The calculated decomposition of $C_nF_{2n+1}SO_3K$ is consistent with experimental results of $C_nF_{2n+1}SO_3K$. We further calculated the decomposition of $C_4F_9SO_3^-$ (Figure S21), and the results are similar to those of $C_4F_9SO_3K$, demonstrating that K^+ minorly influences the decomposition mechanisms. However, some other counterions (e.g., NH_4^+ , Al^{3+} and transition metals) could significantly change the decomposition mechanisms.^{74,75} Notably, the influence of morphology or crystal structures that could change the decomposition kinetics and mechanisms as documented⁷⁶ is not considered in the simulation. The three proposed decomposition pathways did not exclude other mechanisms, as indicated by some unknown products identified in the experiments (Figure 6c).

When PFSA-K are adsorbed onto GAC, the interactions between GAC (or its functionalized surface) and PFSA-K can alter the thermal decomposition pathways to form SO_2 and SiF_4 even at low temperatures (<400 °C). Based on the observed products, we proposed that GAC can enhance the thermal decomposition of adsorbed PFSA-K through chemical interactions between active sites or compounds on the GAC surface and PFSA-K (Figure 6d), which can (1) facilitate the breakdown of the C–S bond, resulting in the degradation of the sulfonate group into SO_2 , and (2) mediate surface reactions with SiO_2 to form SiF_4 .

3.5. Implications. This study provides critical environmental implications for thermal treatment of PFAS. First, the lowest temperatures (e.g., 280 °C) and the most effective temperature ranges (e.g., 389–470 °C) were determined for PFSA-laden GAC decomposition in thermal regeneration of spent GAC, which is meaningful for optimization of operation parameters.^{29,35,35} Second, the form of PFAS molecules (e.g., $C_nF_{2n+1}SO_3^-$ vs $C_nF_{2n+1}SO_3H$) and environmental matrix (e.g., GAC) can significantly influence and even alter the decomposition mechanisms and products. The change in mechanisms from unimolecular decomposition in pure PFAS standards to surface-mediated reactions in PFAS-laden GAC underscores the complex nature of PFAS in thermal treatment processes. We strongly suggest that the influence of environmental matrices/conditions (e.g., pH, co-occurring materials) should be considered for future research about thermal treatment of PFAS. Third, we identified some important products such as SiF_4 , SO_2F_2 , SO_2 , $KFSO_3$, K_2SO_4 , and a few olefins based on IR and MS technologies, but some VOF products remain unknown due to instrument limitations in our analysis. Gas chromatography/mass spectrometry (GC/MS) could be used to measure these unknown VOF after

separation. Finally, we found some materials, such as quartz, that can promote the mineralization of PFASs into SiF₄, which is meaningful for catalyst development.

Pyrolysis is an effective remediation strategy for PFAS-impacted materials such as GAC, soil, and biosolids,^{77–81} but closing fluorine mass balance is still a challenge in thermal treatment due to unknown fluorine products.^{36,37} This study revealed fluorine products, such as VOF, SiF₄, SO₂F₂, and KFSO₃, which contribute to the goal of closing the F mass balance. The VOF has high global warming potential and perfluoroalkenes can be oxidized into PFCAs in the atmosphere, which impose significant environmental risks.^{46,82,83} Thus, the study about detailed characterization of evolved VOF in pyrolysis of PFASs is warranted for further combustion treatment as a strategy for VOF control.^{32,67,84}

■ ASSOCIATED CONTENT

SI Supporting Information

The Supporting Information is available free of charge at <https://pubs.acs.org/doi/10.1021/acs.est.4c06805>.

Preparation of PFAS-laden GAC samples, chemical information, qualitative analysis method, diagrams of thermal analyzers and tube furnace, mass spectra, FTIR spectra, formation trends of EI fragments of gaseous products, detection of FSO₃⁻, TGA/DSC curves, theoretical potential energy profiles (PDF)

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Notes

The authors declare no competing financial interest.

■ ACKNOWLEDGMENTS

This research was supported by the DoD Strategic Environmental Research and Development Program (Project ER21-1191) and the UCR OASIS Internal Funding Award. M.L. was supported by an NSF graduate research fellowship. W.W. and A.C. were supported in part by an NRSA T32 training grant (T32ES018827). We greatly appreciate Dr. Brett Boote's assistance in the simultaneous thermal analysis at Iowa State University. Computations were performed using the computer clusters and data storage resources of the High-Performance Computer Cluster at the University of California, Riverside, which were funded by grants from the National Science Foundation (MRI-2215705 and MRI-1429826) and the National Institutes of Health (1S10OD016290-01A1).

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