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Niche differentiation, reproductive interference, and range expansion

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Abstract

Understanding species distributions and predicting future range shifts requires considering all relevant abiotic factors and biotic interactions. Resource competition has received the most attention, but reproductive interference is another widespread biotic interaction that could influence species ranges. Rubyspot damselflies (*Hetaerina* spp.) exhibit a biogeographic pattern consistent with the hypothesis that reproductive interference has limited range expansion. Here, we use ecological niche models to evaluate whether this pattern could have instead been caused by niche differentiation. We found evidence for climatic niche differentiation, but the species that encounters the least reproductive interference has one of the narrowest and most peripheral niches. These findings strengthen the case that reproductive interference has limited range expansion and also provide a counterexample to the idea that release from negative species interactions triggers niche expansion. We propose that release from reproductive interference enables species to expand in range while specializing on the habitats most suitable for breeding.

KEYWORDS

behavioural interference, ecological niche model, ecological release, habitat suitability model, niche breadth, niche overlap, niche similarity, Odonata, reproductive interference, species distribution model

INTRODUCTION

Understanding how species ranges have been shaped by physiological tolerances and biotic interactions is crucial for predicting how species ranges will shift in the future (Lancaster, 2022; Sexton et al., 2009). Physiological tolerances can be measured experimentally or inferred statistically from the relationships between environmental variables and occurrence data. While ‘biotic interactions’ is a large and diverse category, the vast majority of theoretical and empirical research has focused on trophic interactions and exploitative resource competition, which are closely related phenomena. Exploitative competition can exclude species from portions of their fundamental niche by reducing resources below the level required for the population to increase when rare (Godsoe et al., 2017). Other negative species interactions

have often been assumed to have effects similar to exploitative competition, but theoretical and empirical research has shown that not to be the case. Direct interactions, such as aggressive and reproductive interference, can shape species distributions in ways that would not be predicted by niche theory (Grether et al., 2017; Grether & Okamoto, 2022; Kishi & Nakazawa, 2013; Patterson & Drury, 2023). Interspecific interference does not consistently favour species with higher intrinsic growth rates and can cause Allee effects that prevent species from persisting at low densities in environments that could support much higher densities (Case et al., 2005). If interference is asymmetrical, the species harmed least could prevail even if it has a lower intrinsic growth rate in the absence of interference (Grether & Okamoto, 2022; Kishi & Nakazawa, 2013; Ribeiro & Spielman, 1986). If interference is symmetrical, the first

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species established in an area could become a barrier to the range expansion of other species (Case et al., 2005). Although progress has been made, empirical research on the real-world consequences of interspecific interference lags far behind theory and laboratory experiments.

Reproductive interference is a widespread form of interspecific interference that occurs in all kinds of sexually reproducing organisms (Gröning & Hochkirch, 2008; Takahashi et al., 2016; Ting & Cutter, 2018). In animals, reproductive interference is typically caused by persistent mate recognition errors, which can be costly for both species and yet impossible for selection to eliminate (Drury et al., 2019). Theoretical and experimental work has shown that reproductive interference can cause Allee effects and impede range expansion, particularly in taxa with limited dispersal capabilities (Bargielowski & Lounibos, 2016; Case et al., 2005; Kishi et al., 2009; Kishi & Nakazawa, 2013; Kuno, 1992; Kyogoku, 2015; Kyogoku & Nishida, 2012; Kyogoku & Wheatcroft, 2020; Noriyuki & Osawa, 2016). However, the best evidence that reproductive interference has shaped species ranges comes from studies of invasive species actively displacing other species (Bargielowski et al., 2013; Bargielowski & Lounibos, 2016; Butler & Stein, 1985; Liu et al., 2007; Söderbäck, 1995; Sun et al., 2014). Determining whether reproductive interference has limited species ranges is more challenging.

Rubyspot damselflies (*Hetaerina* spp.; Garrison, 1990) are a promising study system in this regard because previous research has revealed a biogeographic pattern consistent with the hypothesis that reproductive interference has limited range expansion in North America (Drury et al., 2015, 2019; Grether et al., 2020). Briefly, the species that encounters the least reproductive interference has the largest latitudinal range and overlaps with the most congeners, while the species with the largest longitudinal range is allopatric across most of its range, and thus has also largely escaped from reproductive interference. Field experiments and landscape genomic analyses will be required to test this hypothesis directly. Here, we use existing data to evaluate whether niche differentiation could account for the same biogeographic pattern. If niche differentiation cannot account for the large differences between species in range size and overlap, the case for reproductive interference limiting species ranges would be strengthened.

A few conceptual and methodological distinctions are needed to clarify our approach to testing predictions of the niche differentiation hypothesis. The various ecological niche concepts all have in common the idea that a species' ecological niche pertains to environmental space (Godsoe et al., 2017; Sillero et al., 2021; Vandermeer, 1972). As such, niche axes are environmental variables, not geographic or geo-environmental dimensions. Niche differentiation refers to changes in the relationship between environmental variables and population growth rates that occurred after two or more

species diverged from a common ancestor. Species with broader niches are capable of persisting in a broader range of environments and thus tend to have larger geographic ranges (Brown, 1984; Carscadden et al., 2020; Kambach et al., 2019; Slatyer et al., 2013). In taxonomic groups in which niche differentiation has occurred and most species are habitat specialists, a habitat generalist with a relatively broad niche (Figure 1a), or one near the ancestral group centroid (Figure 1b), would be expected to have the largest geographic range and the most geographic overlap with the other species (Suárez-Mota & Villaseñor, 2020; Tomašových et al., 2017).

Ecological niche modelling is widely used to characterize and compare species' realized niches (i.e., the portions of fundamental niches currently occupied) and to make predictions about species distributions in regions that are incompletely surveyed or in the future based on climate change projections (Ahmadi et al., 2018; Boys et al., 2021; Buermann et al., 2008; Guisan et al., 2013; Inman et al., 2019; Melo-Merino et al., 2020; Novella-Fernandez et al., 2021; Pearson & Dawson, 2003; Peterson & Holt, 2003). While the focus of such studies is usually on geographic space, ecological niche models (ENMs) can also be used to make inferences about niche differentiation (Ahmadi et al., 2018; Evans & Jacquemyn, 2022; Warren et al., 2019, 2021). To avoid ambiguity, we focus here on ENMs constructed using Maxent (Phillips et al., 2017), but the same concepts apply to other correlative ENMs. Maxent is a machine-learning, presence-background method that often outperforms presence-only and presence-absence methods (Elith et al., 2011; Merow et al., 2013; Valavi et al., 2022; Wisz et al., 2008) and is widely used for species comparisons (Jaime et al., 2015; Namyatova, 2020; Wellenreuther et al., 2012). The basic output is an equation for predicting the suitability of any combination of values of the environmental variables in the training set. Suitability can be projected to the geographic extent or to an n -dimensional environmental hypercube, where n is the number of environmental variables (Figure 1c–e; Warren et al., 2019). The environmental hypercube is closely analogous to Hutchinson's (1957) n -dimensional hypervolume, except that instead of being the discrete set of environments where the species has a positive intrinsic growth rate, all environments in the hypercube are weighted by their estimated suitability. Statistics commonly referred to as 'niche breadth' and 'niche similarity' (or 'niche overlap') can be calculated using the hypercube or the geographic projection (Figure 1d,e; Warren et al., 2021).

Importantly, niche metrics based on geographic projections are not suitable for studying niche differentiation because they conflate geography and environment (Figure 1d,e). Consider the question of whether species differences in niche breadth can explain differences in range size. It would be circular to test whether range size correlates with niche breadth calculated using geographic

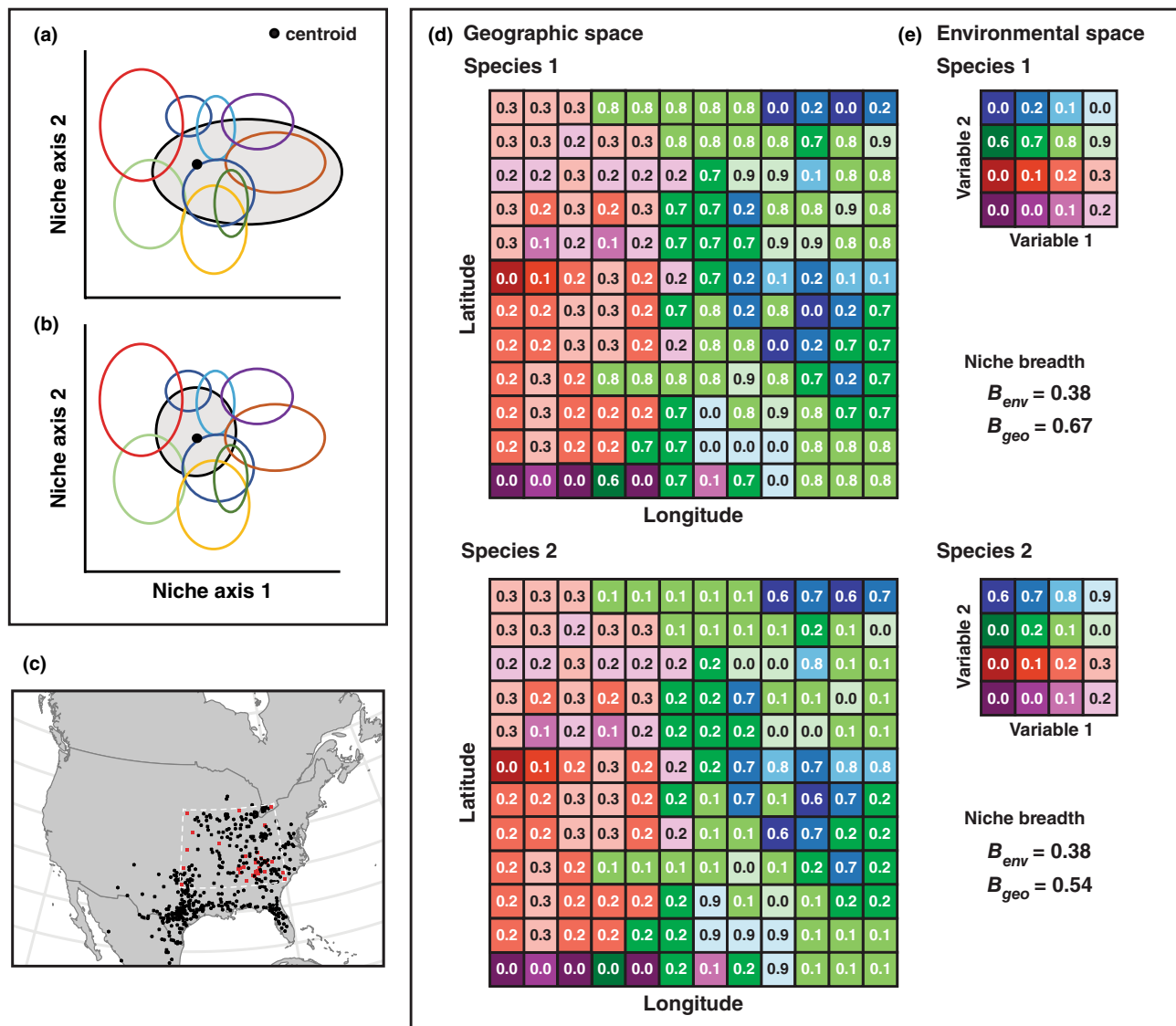


FIGURE 1 Conceptual diagrams. Upper left: illustration showing that one species' niche (grey ellipse) could overlap the niches of all other species in the group by being (a) especially broad or (b) close to the group centroid (multivariate average). Lower left (c): map depicting the shared geographic extent of two species (within the dashed lines), with red and black symbols representing occurrence records. Right: hypothetical examples of ENM suitability values projected to (d) geographic space and (e) environmental space for two species. For illustration purposes, the environmental space consists of two variables and 16 environments (colours) and the geographic space consists of 144 locations. Each environment is represented by one cell in environmental space and by multiple cells in geographic space. The numbers in the cells represent suitability values. Green environments have higher suitability for species 1 and blue environments have higher suitability for species 2, but the species have the same distribution of suitability values and therefore the same niche breadth in environmental space (B_{env}). Green environments are more common on the geographic landscape than blue environments, and thus species 1 has greater niche breadth than species 2 in geographic space (B_{geo}). Niche breadth was calculated with Levin's normalized niche breadth equation (B2).

projections that are based on occurrence samples from across the species' ranges. For this question, it only makes sense to use estimates of niche breadth in environmental space. Likewise, estimates of niche similarity based on geographic projections confound niche overlap with geographic overlap, and thus should not be used to evaluate whether niche similarity predicts geographic overlap. While these points might seem obvious, they have not surfaced in our reading of the niche modelling literature, and most papers with niche statistics based on ENMs use the geographic versions (Evans & Jacquemyn, 2022).

To evaluate whether niche differentiation can account for large differences in range size and geographic overlap, we constructed ENMs for 17 of the 18 species of calopterygid damselflies in North America and made several types of niche comparisons in environmental space. Surprisingly, we found that the species with the largest latitudinal range and the most geographic overlap with other species has one of the narrowest ecological niches and is no closer in niche space to the other species than expected by chance. These results strongly suggest that something other than niche differentiation

accounts for this species' expansive range, and therefore strengthen the case for reproductive interference limiting species ranges. This study also provides a counterexample to the idea that release from negative species interactions precipitates niche expansion (Herrmann et al., 2021; Lancaster, 2022).

MATERIALS AND METHODS

Study system

The life cycle of rubyspot damselflies is centred around areas with suitable larval habitat in perennial springs, streams, or rivers (Corbet, 1999; Delgado, 2002; Johnson, 1973; Novelo-Gutiérrez, 2000). At the adult stage, females return repeatedly to the larval habitat to oviposit in submerged vegetation, and males compete for perching sites above the water surface and attempt to clasp arriving females (Córdoba-Aguilar & González-Tokman, 2014). Most sympatric species experience high levels of reproductive interference because the females are too similar phenotypically for males to reliably distinguish between them, the only striking exception being the smoky rubyspot damselfly, *Hetaerina titia* (Drury) (Drury et al., 2015, 2019; Grether et al., 2020). Female *H. titia* have darker wings than females of other species and consequently are pursued and clasped at lower rates by heterospecific males (Drury et al., 2015, 2019). Release from reproductive interference could explain why *H. titia* has the largest latitudinal range and occurs in sympatry with the most congeners (Figure 2). The species with the largest longitudinal range, *H. americana* (Fabricius), was recently split into two sister species based on genetic data (Standing et al., 2022; Vega-Sánchez et al., 2019, 2020), but the occurrence records are indistinguishable; we refer to this monophyletic subclade as '*H. americana* spp.'. The melanic wing pigmentation that makes *H. titia* distinct from congeners increases its resemblance to *Calopteryx* (Figure 3; Córdoba-Aguilar et al., 2007), a younger genus in the same family (Calopterygidae; Standing et al., 2022). Our niche comparisons include 12 of the 13 species of *Hetaerina* and all five species of *Calopteryx* that occur in North America.

Occurrence data and range overlap

Species occurrence data were obtained from the Global Biodiversity Information Facility in December 2020 (<https://doi.org/10.15468/dl.wmxwvj>; <https://doi.org/10.15468/dl.cd58ag>) and trimmed to a rectangular area (in degrees of longitude and latitude) slightly beyond the geographic range of *H. titia* (−130, −60, 0, 55) to encompass the ranges of all species of Calopterygidae with which *H. titia* overlaps. We refer to this area as the 'full extent', to distinguish it from the rectangular areas of overlap

between individual species pairs, which we refer to as 'shared extents' (Figure 1c). To estimate species' range sizes, we used R package 'alphahull' (Pateiro-López & Rodríguez-Casal, 2010; The R Foundation for Statistical Computing, R version 4.1.2).

Environmental variables

Variable selection is a crucial part of ecological niche modelling (Sillero et al., 2021; Warren et al., 2014). Maxent requires environmental data across the full geographic extent at locations where the species is known to occur and randomly selected background points (Merow et al., 2013; Phillips & Dudík, 2008). The standard approach for most applications is to use a combination of bioclimatic and land cover layers, but it is important to consider the natural history of the species and the goals of the study (Phillips et al., 2006). It might be supposed that aquatic variables would be required to construct ENMs for insects with aquatic larvae, but the abiotic variables that govern aquatic insect assemblages (e.g., light levels, water chemistry, water clarity, water temperature, disturbance regimes, organic matter, sedimentation) are strongly influenced by and therefore covary with climate, elevation, and land cover (Burgherr & Ward, 2001; Faria et al., 2021; Null et al., 2013; Yoshimura, 2012). Bioclimatic layers based on air temperature and precipitation have often been used along with elevation and land cover layers to construct ENMs for aquatic plants (Alahuhta et al., 2011; Gillard et al., 2017; Lumbreras et al., 2013), freshwater fish (Comte & Grenouillet, 2015), and aquatic insects (Kusch, 2015; Megna et al., 2021; Shah et al., 2015; Sundar et al., 2021) including Odonata (Bhowmik & Schäfer, 2015; Cancellario et al., 2022). Using more proximal environmental variables does not necessarily result in better niche models; e.g., for modelling the distributions of aquatic invasive species in lakes across the United States, the WorldClim air temperature layers consistently outperformed a new global dataset of lake surface water temperatures (Burner et al., 2023).

Scale is another important consideration (Elith & Leathwick, 2009; Phillips et al., 2006). Testing the hypothesis that niche differentiation accounts for differences between species' ranges on a continental scale requires environmental variables that could differ between species' ranges on a continental scale. An environmental layer for the presence/absence of flowing water would be useful for generating high-resolution habitat suitability maps but not for studying niche differentiation, because flowing water is a habitat requirement for all calopterygid damselflies in North America (Westfall & May, 1996). Syntopic species of rubyspot damselflies often differ in microhabitat use (e.g., current speed, canopy cover and stream width) (Anderson & Grether, 2011; McEachin et al., 2021), which suggests that microhabitat data would be useful for explaining the distributions and

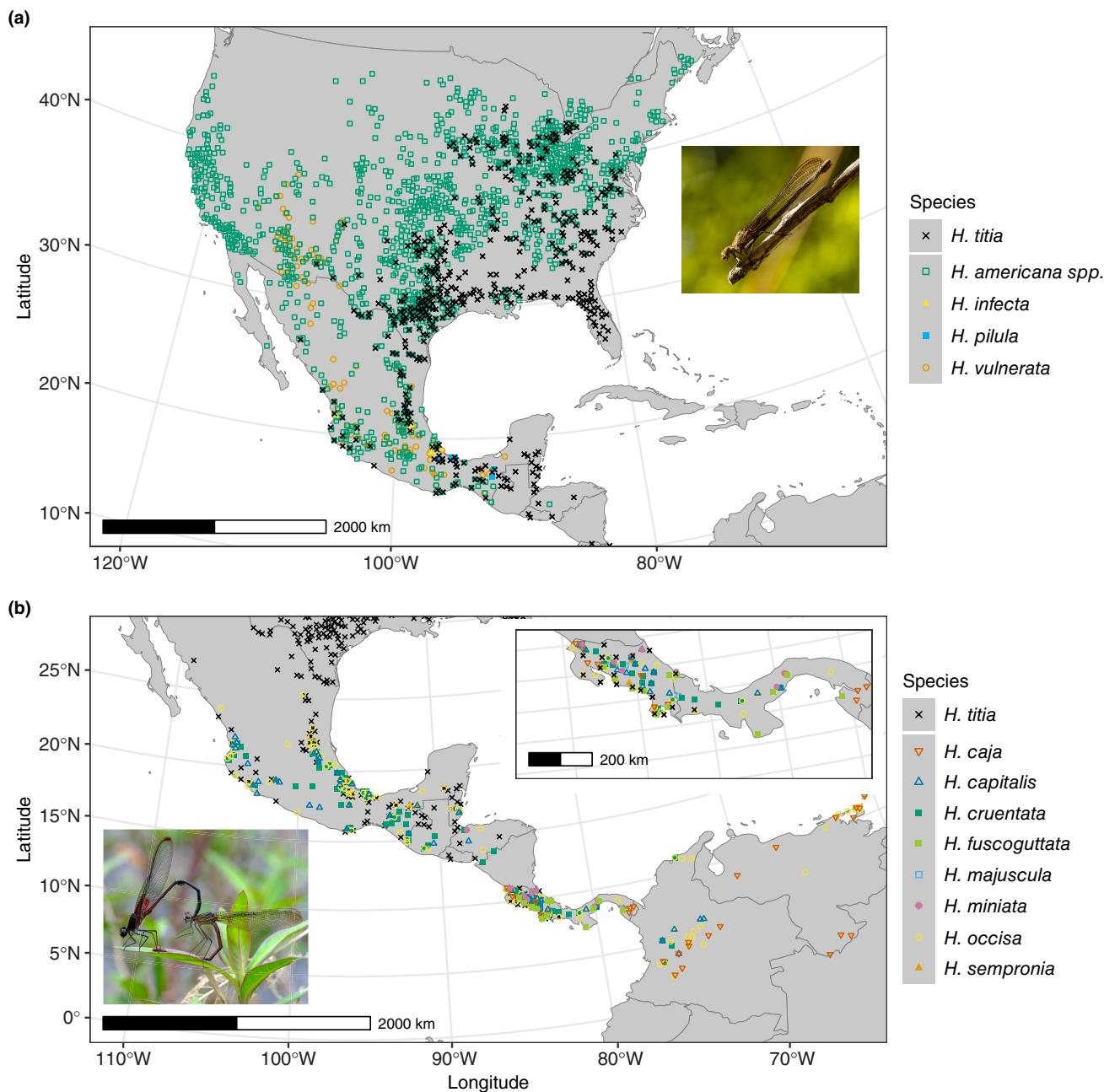


FIGURE 2 Occurrence maps for all species of rubyspot damselflies (*Hetaerina* spp.) within the range of *H. titia*, separated into (a) northern and (b) southern groups to facilitate viewing. The inset map in (b) is a close-up view of Costa Rica and Panama. Points for *H. titia* were plotted on top of the other species' points in (a) and the inset map. Occurrence records were thinned to a minimum separation distance of 25 km for plotting. Inset photos: (a) female *H. americana* hunting; (b) mating pair of *H. cruentata*. Mapped with Albers equal-area conic projection for North America using R package 'sf' version 1.0-13. Photo credits: G.F. Grether.

relative densities of species within streams, as has been shown in other aquatic insects (Goss et al., 2020), but microhabitat differences could not account for species range differences at the continental scale.

Based on the above considerations, we followed the standard approach and constructed ENMs using bioclimatic and land cover layers. Air temperature, precipitation, and elevation data were obtained from WorldClim version 2.1 (Fick & Hijmans, 2017) and averaged over the years 1970–2000. Percent tree cover and land cover

type (Table S1) were obtained from MODIS Terra data 2008 (Geospatial Information Authority of Japan). Normalized difference vegetation index (NDVI) was obtained from NASA Earthdata using 'MODISstp' (Busetto & Raghetti, 2016). Land cover type was aggregated from 15 to 30 arc-seconds (~1 km) to match the spatial resolution of the other variables.

Prior to niche modelling, we cropped the environmental raster files to the full extent using the 'crop' function in R package 'raster'. To avoid problems with highly

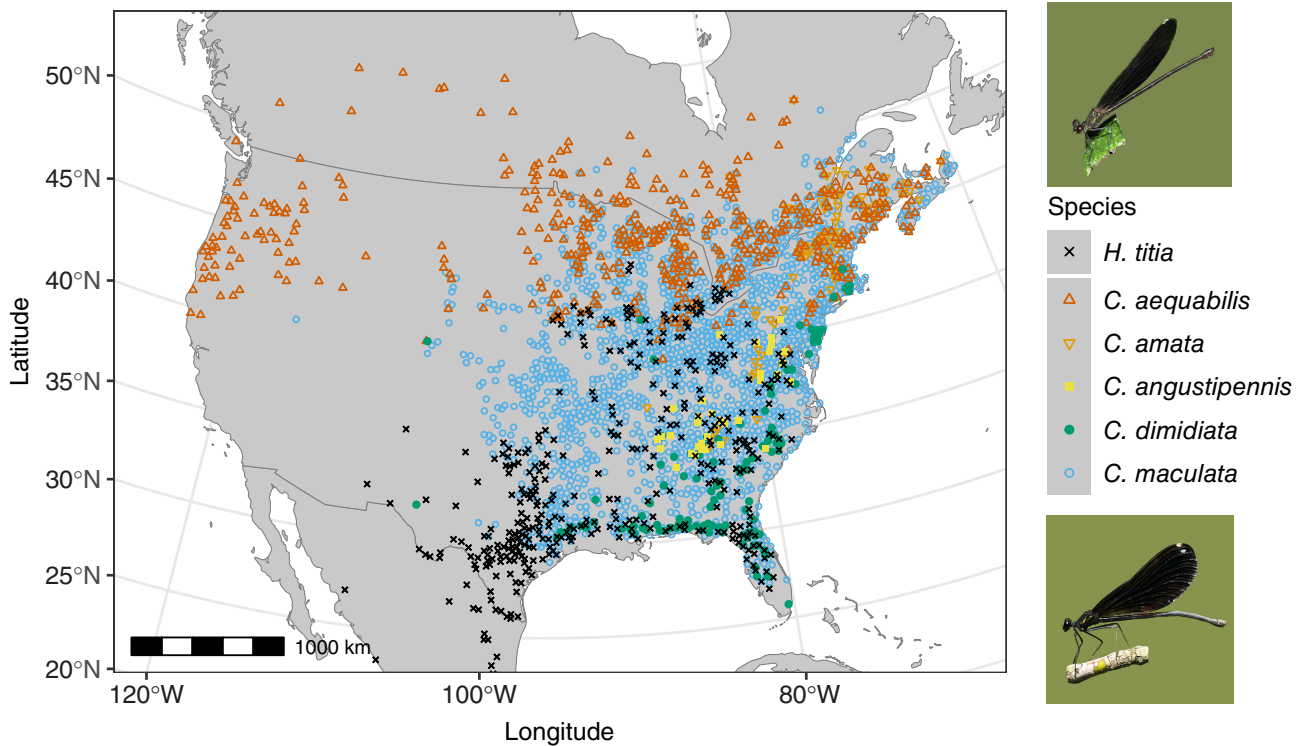


FIGURE 3 Occurrence maps for all species of *Calopteryx* within the range of *H. titia*. Points for *H. titia* were plotted on top of the other species' points. Upper photo: female *H. titia*. Lower photo: female *C. maculata*. Photo credits: G.F. Grether. For mapping details, see [Figure 1](#).

correlated predictor variables (Sillero et al., 2021), we removed variables with variance inflation factors (VIFs) greater than 2.5 (Hair et al., 2019) using the step-wise procedure 'vifstep' in R package 'usdm' (Naimi et al., 2014). The final variable set included 10 variables: NDVI, tree cover, elevation, the categorical variable landcover type, and six bioclimatic variables (BIO2, BIO7, BIO8, BIO15, BIO18 and BIO19) (Table S2).

Principal component analysis

To visualize differences between species in environmental space, we carried out a principal component analysis (PCA) of variation among occurrence sites in the continuous environmental variables with the 'prcomp' function in R.

Niche models

We constructed ENMs for species with 15 or more occurrence records, after thinning, with Maxent 3.4.1 in R package 'dismo' (Hijmans et al., 2021; Phillips et al., 2017). This sample size threshold enabled all ENMs to be based on the same environmental variable transformations, and thus to be directly comparable (Morales et al., 2017), but excluded *H. pilula* (Calvert). We thinned the data to one occurrence per km² per species using 'spThin' (Aiello-Lammens et al., 2015) and

used the following Maxent settings: 10⁴ background points; regularization parameter of 1; 500 iterations; linear, quadratic, product, and hinge features; complementary log–log (cloglog) output; and 10 replicates per model with cross-validation and 30% random test percentage (Fithian et al., 2015; Phillips et al., 2017; Phillips & Dudík, 2008). Model performance can be evaluated by comparing the area under the curve (AUC) to the null AUC (Phillips & Dudík, 2008). We computed the null AUC with R package 'ENMTools' v1.0.3 and subtracted it from the model AUC to obtain a measure of model performance (Δ AUC).

Niche breadth and similarity

A measure of niche breadth in environmental space (B_{env}) based on Levin's normalized niche breadth (B2) equation was calculated with the 'env.breadth' function in 'ENMTools' (Warren et al., 2019). Two measures of niche similarity in environmental space (D_{env} , I_{env}) based on Schoener's D and Hellinger's I were calculated using the 'identity.test' function (Warren et al., 2019). To evaluate whether these metrics are affected by sample size, we carried out a Monte Carlo simulation. Paired samples ranging in size from 10 to 1000 were drawn without replacement from the *H. titia* ($n=1080$) and *H. americana* ($n=3364$) occurrence records. With each pair of samples, 'ENMTools' was used to generate empirical estimates and 100 null values of the niche similarity metrics. The

empirical and null estimates both increased asymptotically with the number of occurrence records (Figure S1). We explored different ways of adjusting the empirical estimates to reduce their dependence on sample size. Subtracting the null mean was the most effective adjustment (Figure S1), and thus we used this method. The niche breadth metric (B_{env}) was not affected by sample size (Figure S1).

Phylogenetic linear mixed models

To account for phylogenetic nonindependence in comparisons of niche metrics between species pairs (Drury et al., 2018; Tobias et al., 2014), we fit phylogenetic linear mixed models (PLMMs) with the Markov chain Monte Carlo (MCMC) algorithm in ‘MCMCglmm’ 2.3 (Hadfield, 2010). We used the fossil-calibrated phylogeny of Standring et al. (2022), appending *C. diminiata* (Burmeister) and *C. amata* (Hagen) based on the topology of Waller and Svensson (2017) (Figure S2). *C. angustipennis* (Selys) could not be included because no phylogenetic data were available for this species. Each PLMM included a species comparison index, patristic distance, and random-effects terms for species identifiers and most recent common ancestors. We ran each model four times and merged the MCMC chains after verifying convergence (Gelman & Rubin, 1992).

Shared-extent comparisons

ENMs based on different geographic extents are not comparable (Merow et al., 2013), but using large extents for species with small ranges can result in overfitting and AUC inflation (Fourcade et al., 2014; Sillero et al., 2021). To make pairwise species comparisons, we constructed ENMs using the shared extents of each species pair. This allowed species to be compared in niche breadth (B_{env}) and two additional metrics: mean suitability and proportion of the shared extent occupied. If a geographic area has higher mean suitability for species A than species B, it should be easier for species A to become established in the area. If species A occurs in a larger proportion of a geographic area than species B, it follows that species A has some advantage relative to species B. Making such comparisons in the full extent would be uninformative because mean suitability is affected by range size, and the proportion of the full extent occupied by a species is directly proportional to its range size. In the shared extents, however, these comparisons are valid and could be informative.

We made two types of shared-extent comparisons. First, we compared the niche metrics of *H. titia* to those of the other 16 species using Wilcoxon paired tests. Second, we compared *H. americana* and *H. titia* in their respective shared extents with 10 other species by

subtracting the other species' niche metrics from those of *H. americana* and *H. titia* and comparing the two sets of differences with PLMMs.

RESULTS

Principal component space

The first three principal components (PCs) accounted for 68.5% of the variance in the continuous environmental variables (Table S3). The first two PCs separated the species into two clusters, with species with northerly ranges in one cluster and species with southerly ranges in the other; the third PC largely separated species by mean temperature, elevation and tree cover (Figure 4). On the first two PCs, *H. titia* occupied a position between the two clusters but closer to the northern cluster, yet on PC3, *H. titia* occupied a peripheral position, with a mean exceeding that of all species except *H. pilula* (Figure 4). The centroid of *H. titia* was closest to species that occur in the eastern United States and farthest from species restricted to Central America (Table S4).

Maxent models

The full-extent ENMs significantly outperformed chance expectations, with Δ AUC in the 0.2–0.4 range (Table S4); most shared-extent models also had Δ AUC in the 0.2–0.4 range (Tables S5 and S6).

Climatic variables made the largest contributions to the full-extent ENMs, with landcover type perhaps controlling for the relatively recent effects of human activities on damselfly distributions (Table S7). Annual temperature range (BIO7) made the largest contribution to most *Hetaerina* spp. ENMs, while precipitation seasonality (BIO15) or elevation made the largest contribution to *Calopteryx* spp. ENMs. Elevation and precipitation in the coldest quarter (BIO19) made the largest contributions to the ENMs of *H. vulnerata* (Hagen) and *H. fuscoguttata* (Selys), respectively. Landcover type was the second or third most important variable for four *Hetaerina* species and caused the largest decrease in gain when omitted from the *H. miniata* (Selys) model. The vegetation index (NDVI) made the third largest contribution to the *H. americana* spp. ENM. The top three variables for *H. titia* were annual temperature range, elevation, and mean temperature in the warmest quarter (BIO8).

Niche metrics

The hypothesis that *H. titia* overlaps geographically with multiple congeners because it occupies a niche near the

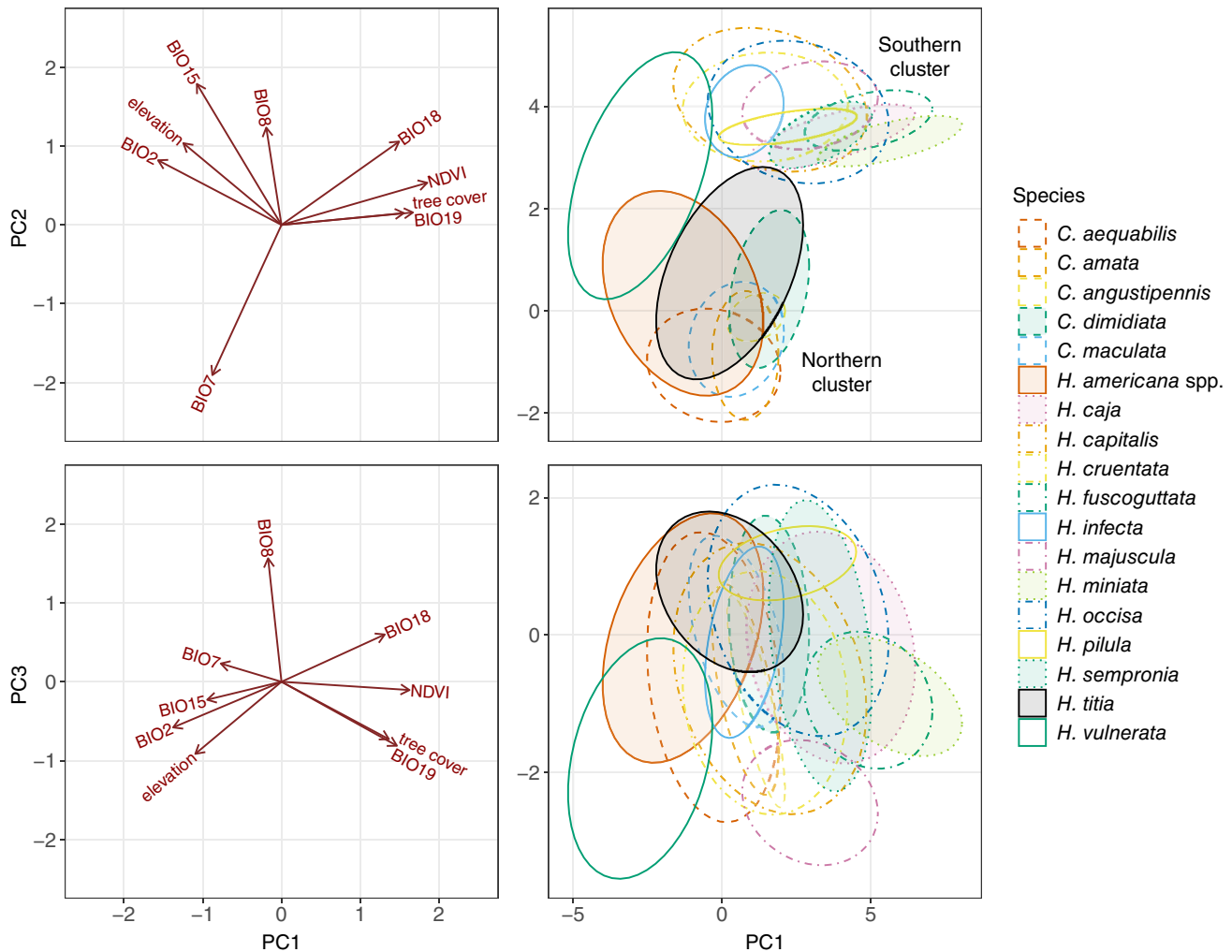


FIGURE 4 Principal component analysis of environmental variation among the occurrence records of 18 species of calopterygid damselflies within the geographic range of *H. titia*. Left panels: arrows representing the principal component loadings of the environmental variables. Right panels: probability ellipses encompassing 68% (1 standard deviation) of the occurrence records. Key to variable names: diurnal temperature range (BIO2), annual temperature range (BIO7), mean temperature of wettest quarter (BIO8), precipitation seasonality (BIO15), precipitation of warmest quarter (BIO18), precipitation of coldest quarter (BIO19), normalized difference vegetation index (NDVI).

genus centroid can be rejected. The ENM of *H. titia* in the full geographic extent was not more similar to that of other calopterygid species than the other species' ENMs were to each other (Figure 5a; Table S4; PLMM, $p \geq 0.24$, 120 species pairs), and the same was true when the analysis was restricted to *Hetaerina* spp. ($p \geq 0.47$, 66 species pairs). The ENM of *H. titia* differed significantly from all other species except *H. sempronia* (Hagen) and *H. fuscoguttata* and was least similar to species with higher latitudinal or elevational ranges (Table S4).

The hypothesis that *H. titia* has an unusually broad niche can also be rejected. *H. titia* had the fourth smallest niche breadth in the full extent (Table S4) and a smaller niche breadth than the other species in the shared extents (Wilcoxon paired test $V = 125$, $n = 16$, $p = 0.002$; Figure 6a). *H. titia* did not differ in mean suitability from the other species ($V = 91$, $n = 16$, $p = 0.25$) and yet occupied a larger proportion of the shared extents ($V = 12$, $n = 16$, $p = 0.002$; Figure 6b).

The congener with the largest longitudinal range, *H. americana* spp., had the largest niche breadth in the full extent (Table S4). In their respective shared extents with other species, *H. americana* spp. and *H. titia* did not differ in the proportion of the shared extents occupied (PLMM, $p = 0.48$, 20 species pairs) or mean suitability ($p = 0.49$), but *H. americana* spp. tended to have greater relative niche breadth ($p = 0.06$).

Sympatric species (67 species pairs) had more similar niches than allopatric species (53 species pairs; PLMM, $p < 0.001$; Figure 5b). Niche similarity was greater between species in the same genus (72 species pairs) than between species in different genera (48 species pairs; Figure 5c), but with only two genera the phylogeny fully accounted for these differences. Across all species, sympatric species were closer in patristic distance than allopatric species (Wilcoxon test, $W = 2926$, $p < 0.0001$, $n_{allo} = 53$, $n_{sym} = 67$), but with the analysis restricted to *Hetaerina* spp., sympatric and allopatric species did not

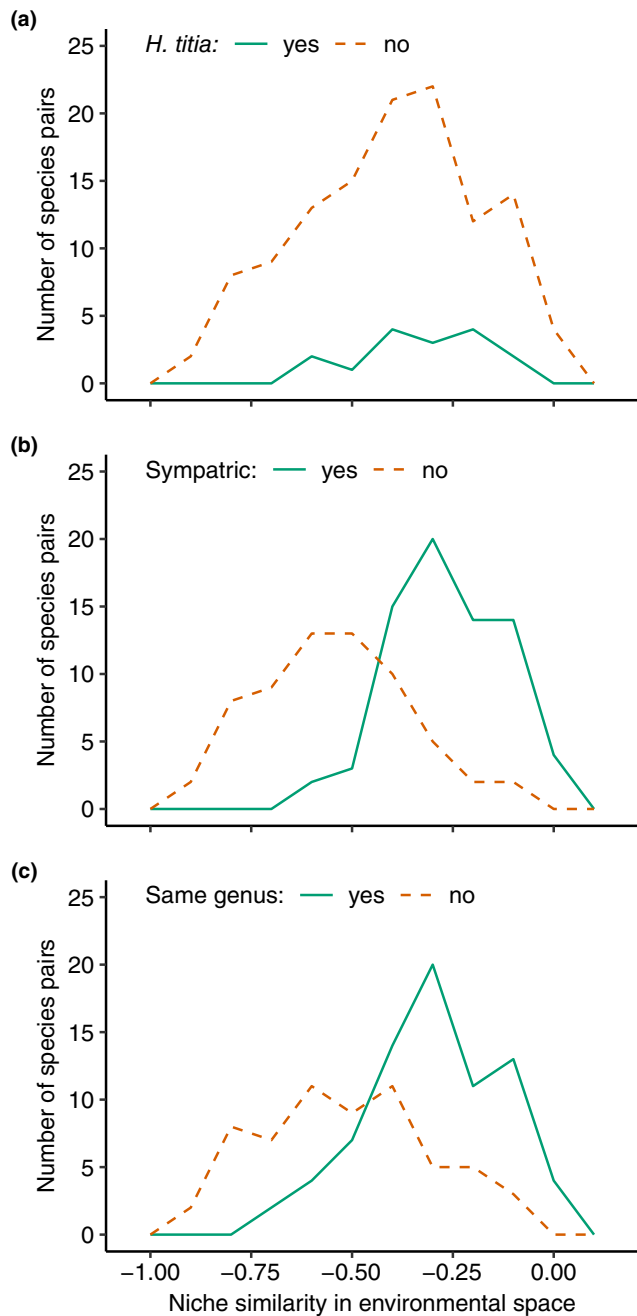


FIGURE 5 Frequency polygons of niche similarity in environmental space for different groupings of calopterygid damselfly species. (a) Species pairs that include *H. titia* versus all other species pairs. (b) Sympatric species pairs versus allopatric species pairs. (c) Species in the same genus versus species in different genera (i.e., *Hetaerina* and *Calopteryx*). Niche similarity was estimated from Maxent ENMs in the full geographic extent using index I_{env} (results for index D_{env} were similar). To remove sample-size effects, the null mean I_{env} was subtracted from the observed I_{env} , and thus values farther below 0 on the horizontal axis represent species pairs with lower niche similarity. Differences less than ~ -0.15 were significant with the identity test (Table S4).

differ in patristic distance ($W=420$, $p=0.21$, $n_{allo}=13$, $n_{sym}=53$). Patristic distance was not a significant predictor of species differences in any niche metric in the full extent (PLMMs, $p \geq 0.21$) or shared extents ($p \geq 0.63$).

DISCUSSION

Modest differences between closely related species in range size might not be predictive of differences in niche breadth, in part simply because some environments are more common than others (Brown, 1984). Yet species with exceptionally large geographic ranges are expected to have relatively broad niches (Brown, 1984; Carscadden et al., 2020; Kambach et al., 2019; Slatyer et al., 2013). Likewise, species that overlap geographically with large numbers of congeners can be predicted to have niches that, if not unusually broad, are close to the genus centroid (Figure 1). We tested these straightforward predictions in calopterygid damselflies, and while we found evidence for niche differentiation within and between genera, the species with the largest latitudinal range and which overlaps with the most congeners (*H. titia*) has a relatively narrow and peripheral climatic niche. By contrast, the species with the largest longitudinal range (*H. americana* spp.) has the largest niche breadth in the genus. Both of these species have largely escaped from reproductive interference, *H. titia* by diverging in female coloration (Drury et al., 2015, 2019; Grether et al., 2020) and *H. americana* spp. by being allopatric to all other congeners across most of its range. Niche differentiation and reproductive interference are not mutually exclusive alternatives—both could have contributed to species differences in range. Nevertheless, finding that the species with the largest latitudinal range, and which encounters the least reproductive interference, has a relatively narrow climatic niche strengthens the hypothesis that reproductive interference has limited the range expansion of other species in this genus.

When a species expands beyond its previous range edge into marginally suitable habitats, ecological and evolutionary responses can broaden the species' niche and facilitate further range expansion (Alexander et al., 2022; Case et al., 2005; Lancaster, 2022). Release from negative species interactions can also lead to niche expansion within a species' existing range (Bolnick et al., 2010; Herrmann et al., 2021; Sjödin et al., 2018). If release from reproductive interference facilitated the range expansion of *H. titia*, why does this species have a relatively narrow climatic niche? In some cases, gene flow from the core of a species' range might prevent adaptation at the edges (Case & Taper, 2000; Kirkpatrick & Barton, 1997), but *Hetaerina* populations in different river drainages are sufficiently isolated genetically to evolve independently (Drury et al., 2019). A better explanation may be that current theory on the effects of species interactions on niche evolution, which is largely based on exploitative resource competition (Alexander et al., 2022), does not apply to reproductive interference.

We propose that, instead of triggering niche expansion, release from reproductive interference enables species to expand in range while specializing on the environments most suitable for breeding. The basic

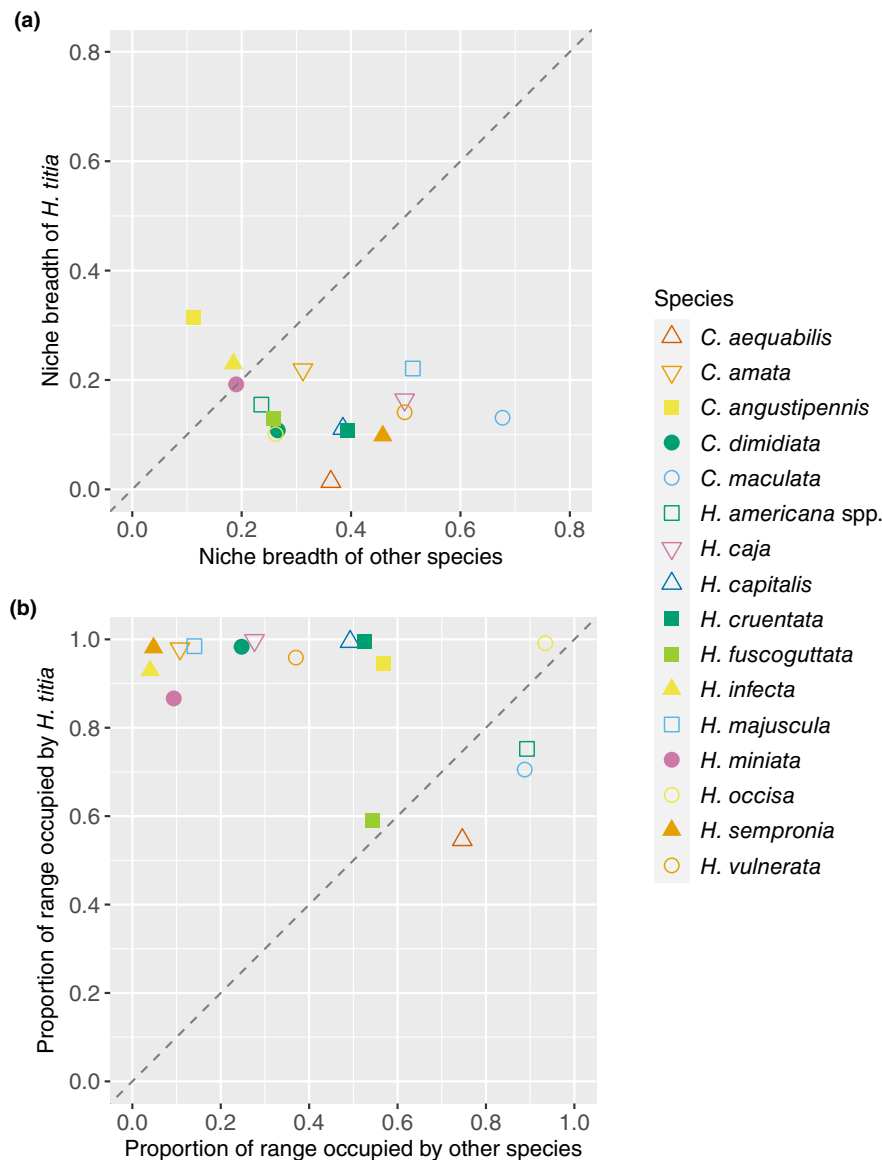


FIGURE 6 Comparisons between *H. titia* and other species of calopterygid damselflies in the shared extents (i.e., rectangular areas of geographic overlap). (a) Niche breadth in environmental space (B_{env}) estimated from separate Maxent models for each shared extent. (b) Proportion of the shared extents occupied (1 km² resolution). The dashed lines have a slope of 1, and thus points above (below) the lines are cases in which *H. titia* had a larger (smaller) value than the other species.

argument for why release from interspecific competition leads to niche expansion is that the absence of the competitor frees up resources and converts unsuitable habitat into suitable habitat (Alexander et al., 2022; Bolnick et al., 2010). Release from reproductive interference would not free up resources, but it would facilitate breeding in areas with high heterospecific densities by eliminating Allee effects. With no barrier to expansion into high-suitability habitats, a species released from reproductive interference would be expected to be concentrated where it reaches the highest population densities, that is, in the core of its fundamental niche. Put simply, release from reproductive interference would enable a species to expand

into habitats that support high population densities, resulting in a distribution of suitability values in environmental space that resembles niche contraction (Warren et al., 2019). Selection might then strengthen the preference for high suitability habitats, resulting in a narrowing of the species' fundamental niche. This hypothesis is similar to the hypothesis that reductions in pollinator-mediated reproductive interference promote coexistence between plants that specialize on the same pollinators (Katsuhara et al., 2021).

We found evidence for niche divergence in the vast majority of congeneric species pairs (Figure 5c), which would seem to differ from what has been found in most other taxa (Peterson, 2011), but *Hetaerina* is an ancient

genus with estimated species divergence times of 10–36 million years (Standring et al., 2022). Our finding that sympatric species have more similar climatic niches than allopatric species is not surprising from a statistical standpoint, because the same result would be expected if species ranges overlapped randomly (Warren et al., 2014). However, several ecological and evolutionary processes can cause the niches of sympatric species to differ more than those of allopatric species (e.g., competitive displacement, species sorting, character displacement; Pfennig & Pfennig, 2012), and negative relationships between niche similarity and geographic overlap have been found in other studies (Novella-Fernandez et al., 2021). Perhaps the factors that mediate coexistence in calopterygid damselflies operate on smaller spatial scales than our niche models (Anderson & Grether, 2011; McEachin et al., 2021). Research on other damselfly clades has shown that predation-mediated trade-offs at the larval stage promote local coexistence in some species assemblages, while in other assemblages the species may be ecologically equivalent and subject to neutral dynamics (Grether et al., 2023; Leibold & McPeck, 2006; Siepielski et al., 2010).

Niche differentiation has often been reported to be decoupled from phylogenetic differentiation (Losos, 2011; Warren et al., 2014), and our study provides another example. Niche similarity was greater within than between genera, but after controlling for tree topology, there was no relationship between niche divergence and patristic distance. This indicates that more closely related congeners are not more similar ecologically. Across all species included in this study, sympatric species were closer in patristic distance than allopatric species, but this appears to be an artefact of the geographic ranges of the genera (*Calopteryx* spp. only occur north of Mexico while most *Hetaerina* spp. occur south of the US-Mexico border; Figures 2 and 3). With the analysis restricted to *Hetaerina*, sympatric and allopatric species did not differ in patristic distance. This might seem to contradict the hypothesis that reproductive interference impedes range expansion, because reproductive interference is generally expected to be stronger between more closely related species (Grether et al., 2017; Gröning & Hochkirch, 2008), but *Hetaerina* does not appear to conform to that expectation (Drury et al., 2015; Grether et al., 2020). In any case, a much larger sample of species would be required to robustly test for phylogenetic effects on geographic overlap between species, especially considering that the null expectation is for sympatric species to be less closely related than allopatric species because speciation usually occurs in allopatry (Tobias et al., 2014; Warren et al., 2014; Weir & Price, 2011).

In principle, joint species distribution models (JSDMs) could be used to statistically distinguish species interactions from niche differentiation (Poggiato

et al., 2021; Wilkinson et al., 2019). We opted not to use this approach because species interactions cannot be distinguished from correlated responses to unmeasured environmental variables (Ovaskainen et al., 2016; Poggiato et al., 2021). With our study system and others, data are not available for all environmental variables that are likely to affect species ranges in similar ways (e.g., dispersal barriers, microhabitat, disturbance). The omission of such variables from a JSDM could result in spurious positive correlations between species that mask negative species interactions (Ovaskainen et al., 2016; Poggiato et al., 2021). Single-species ENMs (and SDMs) are also limited by the data used to construct them, and niche differentiation could be missed if variables that affect species ranges differently are omitted (Peterson & Nakazawa, 2007). We judge this to be a less severe problem than omitting variables that affect species ranges in similar ways from JSDMs, but it is important to consider whether unexplained differences between species ranges could be due to unmeasured variables. We are not aware of any unmeasured variables that might explain continental-scale differences between the species ranges of calopterygid damselflies in North America.

Several other recent studies have drawn inferences about species interactions using ENMs. By comparing the known ranges of Palearctic bat species to those predicted by ENMs, Novella-Fernandez et al. (2021) showed that two pairs of cryptic species with very similar niches exhibit a pattern of geographic avoidance consistent with competitive exclusion. Cavalcante et al. (2022) combined the outputs of ENMs with those of geospatial abundance models and found evidence that the strength of interspecific competition between two endangered Amazonian primates (Atelidae) varies with habitat suitability. Likewise, Braz et al. (2020) found that abundance of a neotropical opossum (*Marmosops incanus*) covaried with the number of potential competitor species at high suitability sites but not at low suitability sites.

In closing we note that niche theory need not be limited to abiotic factors, trophic interactions and resource competition just because Hutchinson (1957) framed it that way. Embracing the full suite of biotic interactions that shape the realized niche would enhance niche theory and its utility for understanding and predicting species distributions. On a methodological note, we encourage niche modellers to pay closer attention to the distinction between geographic and environmental space!

AUTHOR CONTRIBUTIONS

All authors contributed to study design, data analysis, and early manuscript drafts. Ann E. Finneran assembled the data set, wrote and executed the code for running models and most statistical analyses with help from Gregory F. Grether and Jonathan P. Drury.

Gregory F. Grether wrote the final manuscript and created the final figures, tables, and data and code archive.

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DATA AVAILABILITY STATEMENT

The data and code that support the findings of this study are openly available at <https://doi.org/10.6084/m9.figshare.23739582.v2>.

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REFERENCES

- Ahmadi, M., Naderi, M., Kaboli, M., Nazarizadeh, M., Karami, M. & Beitollahi, S.M. (2018) Evolutionary applications of phylogenetically-informed ecological niche modelling (ENM) to explore cryptic diversification over cryptic refugia. *Molecular Phylogenetics and Evolution*, 127, 712–722.
- Aiello-Lammens, M.E., Boria, R.A., Radosavljevic, A., Vilela, B. & Anderson, R.P. (2015) spThin: an R package for spatial thinning of species occurrence records for use in ecological niche models. *Ecography*, 38, 541–545.
- Alahuhta, J., Heino, J. & Luoto, M. (2011) Climate change and the future distributions of aquatic macrophytes across boreal catchments. *Journal of Biogeography*, 38, 383–393.
- Alexander, J.M., Atwater, D.Z., Colautti, R.I. & Hargreaves, A.L. (2022) Effects of species interactions on the potential for evolution at species' range limits. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 377, 20210020.
- Anderson, C.N. & Grether, G.F. (2011) Multiple routes to reduced interspecific territorial fighting in *Hetaerina* damselflies. *Behavioral Ecology*, 22, 527–534.
- Bargielowski, I.E. & Lounibos, L.P. (2016) Satyrization and satyrization-resistance in competitive displacements of invasive mosquito species. *Insect Science*, 23, 162–174.
- Bargielowski, I.E., Lounibos, L.P. & Carrasquilla, M.C. (2013) Evolution of resistance to satyrization through reproductive character displacement in populations of invasive dengue vectors. *Proceedings of the National Academy of Sciences of the United States of America*, 110, 2888–2892.
- Bhowmik, A.K. & Schäfer, R.B. (2015) Large scale relationship between aquatic insect traits and climate. *PLoS ONE*, 10, e0130025.
- Bolnick, D.I., Ingram, T., Stutz, W.E., Snowberg, L.K., Lau, O.L. & Paull, J.S. (2010) Ecological release from interspecific competition leads to decoupled changes in population and individual niche width. *Proceedings of the Royal Society B: Biological Sciences*, 277, 1789–1797.
- Boys, W.A., Siepielski, A.M., Smith, B.D., Patten, M.A. & Bried, J.T. (2021) Predicting the distributions of regional endemic dragonflies using a combined model approach. *Insect Conservation and Diversity*, 14, 52–66.
- Braz, A.G., de Viveiros Grelle, C.E., de Souza Lima Figueiredo, M. & Weber, M.d.M. (2020) Interspecific competition constrains local abundance in highly suitable areas. *Ecography*, 43, 1560–1570.
- Brown, J.H. (1984) On the relationship between abundance and distribution of species. *The American Naturalist*, 124, 255–279.
- Buermann, W., Saatchi, S., Smith, T.B., Zutta, B.R., Chaves, J.A., Milá, B. et al. (2008) Predicting species distributions across the Amazonian and Andean regions using remote sensing data. *Journal of Biogeography*, 35, 1160–1176.
- Burgherr, P. & Ward, J.V. (2001) Longitudinal and seasonal distribution patterns of the benthic fauna of an alpine glacial stream (Val Roseg, Swiss Alps). *Freshwater Biology*, 46, 1705–1721.
- Burner, R.C., Daniel, W.M., Engelstad, P.S., Churchill, C.J. & Erickson, R.A. (2023) BioLake: a first assessment of lake temperature-derived bioclimatic predictors for aquatic invasive species. *Ecosphere*, 14, e4616.
- Busetto, L. & Ranghetti, L. (2016) MODISsp: an R package for automatic preprocessing of MODIS land products time series. *Computational Geosciences*, 97, 40–48.
- Butler, M.J. & Stein, R.A. (1985) An analysis of the mechanisms governing species replacements in crayfish. *Oecologia*, 66, 168–177.
- Cancellario, T., Miranda, R., Baquero, E., Fontaneto, D., Martínez, A. & Mammola, S. (2022) Climate change will redefine taxonomic, functional, and phylogenetic diversity of Odonata in space and time. *npj Biodiversity*, 1, 1.
- Carscadden, K.A., Emery, N.C., Arnillas, C.A., Cadotte, M.W., Afkhami, M.E., Gravel, D. et al. (2020) Niche breadth: causes and consequences for ecology, evolution, and conservation. *The Quarterly Review of Biology*, 95, 179–214.
- Case, T.J., Holt, R.D., Mcpeek, M.A. & Keitt, T.H. (2005) The community context of species' borders: ecological and evolutionary perspectives. *Oikos*, 108, 28–46.
- Case, T.J. & Taper, M.L. (2000) Interspecific competition, environmental gradients, gene flow, and the coevolution of species' borders. *The American Naturalist*, 155, 583–605.
- Cavalcante, T., Weber, M.M. & Barnett, A.A. (2022) Combining geospatial abundance and ecological niche models to identify high-priority areas for conservation: the neglected role of broadscale interspecific competition. *Frontiers in Ecology and Evolution*, 10, 915325.
- Comte, L. & Grenouillet, G. (2015) Distribution shifts of freshwater fish under a variable climate: comparing climatic, bioclimatic and biotic velocities. *Diversity and Distributions*, 21, 1014–1026.
- Corbet, P.S. (1999) *Dragonflies: behavior and ecology of Odonata*. Ithaca, NY: Cornell University Press.
- Córdoba-Aguilar, A. & González-Tokman, D.M. (2014) The behavioral and physiological ecology of adult rubyspot damselflies (*Hetaerina*, Calopterygidae, Odonata). *Advances in the Study of Behaviour*, 46, 311–341.
- Córdoba-Aguilar, A., Leshner-Treviño, A. & Anderson, C. (2007) Sexual selection in *Hetaerina titia* males: a possible key species to understand the evolution of pigmentation in calopterygid damselflies (Odonata: Zygoptera). *Behaviour*, 144, 931–952.
- Delgado, C. (2002) Spatial and temporal distribution patterns of Odonata larvae in the streams of a terra firma forest of the Central Amazon, Brazil. *Journal of Freshwater Ecology*, 17, 555–561.
- Drury, J., Anderson, C., Castillo, M.C., Fisher, J., McEachin, S. & Grether, G.F. (2019) A general explanation for the persistence of reproductive interference. *The American Naturalist*, 194, 268–275.

- Drury, J.P., Grether, G.F., Garland, T. & Morlon, H. (2018) An assessment of phylogenetic tools for analyzing the interplay between interspecific interactions and phenotypic evolution. *Systematic Biology*, 67, 413–427.
- Drury, J.P., Okamoto, K.W., Anderson, C.N. & Grether, G.F. (2015) Reproductive interference explains persistence of aggression between species. *Proceedings of the Royal Society B: Biological Sciences*, 282, 20142256.
- Elith, J. & Leathwick, J.R. (2009) Species distribution models: ecological explanation and prediction across space and time. *Annual Review of Ecology, Evolution, and Systematics*, 40, 677–697.
- Elith, J., Phillips, S.J., Hastie, T., Dudík, M., Chee, Y.E. & Yates, C.J. (2011) A statistical explanation of MaxEnt for ecologists. *Diversity and Distributions*, 17, 43–57.
- Evans, A. & Jacquemyn, H. (2022) Range size and niche breadth as predictors of climate-induced habitat change in *Epipactis* (Orchidaceae). *Frontiers in Ecology and Evolution*, 10, 1203–1214.
- Faria, A.P.J., Paiva, C.K.S., Calvão, L.B., Cruz, G.M. & Juen, L. (2021) Response of aquatic insects to an environmental gradient in Amazonian streams. *Environmental Monitoring and Assessment*, 193, 763.
- Fick, S.E. & Hijmans, R.J. (2017) WorldClim 2: new 1-km spatial resolution climate surfaces for global land areas. *International Journal of Climatology*, 37, 4302–4315.
- Fithian, W., Elith, J., Hastie, T. & Keith, D.A. (2015) Bias correction in species distribution models: pooling survey and collection data for multiple species. *Methods in Ecology and Evolution*, 6, 424–438.
- Fourcade, Y., Engler, J.O., Rödder, D. & Secondi, J. (2014) Mapping species distributions with MAXENT using a geographically biased sample of presence data: a performance assessment of methods for correcting sampling bias. *PLoS ONE*, 9, 1–13.
- Garrison, R.W. (1990) A synopsis of the genus *Hetaerina* with descriptions of four new species (Odonata: Calopterygidae). *Transactions of the American Entomological Society*, 116, 175–259.
- Gelman, A. & Rubin, D.B. (1992) Inference from iterative simulation using multiple sequences. *Statistical Science*, 7, 457–472.
- Gillard, M., Thiébaud, G., Deleu, C. & Leroy, B. (2017) Present and future distribution of three aquatic plants taxa across the world: decrease in native and increase in invasive ranges. *Biological Invasions*, 19, 2159–2170.
- Godsoe, W., Jankowski, J., Holt, R.D. & Gravel, D. (2017) Integrating biogeography with contemporary niche theory. *Trends in Ecology & Evolution*, 32, 488–499.
- Goss, C.W., Sullivan, S.M.P. & Goebel, P.C. (2020) Effects of land-cover transitions on emerging aquatic insects and environmental characteristics of headwater streams in an agricultural catchment. *River Research and Applications*, 36, 1097–1108.
- Grether, G.F., Drury, J.P., Okamoto, K.W., McEachin, S. & Anderson, C.N. (2020) Predicting evolutionary responses to interspecific interference in the wild. *Ecology Letters*, 23, 221–230.
- Grether, G.F. & Okamoto, K.W. (2022) Eco-evolutionary dynamics of interference competition. *Ecology Letters*, 25, 2167–2176.
- Grether, G.F., Peiman, K.S., Tobias, J.A. & Robinson, B.W. (2017) Causes and consequences of behavioral interference between species. *Trends in Ecology & Evolution*, 32, 760–772.
- Grether, G.F., Siepielski, A.M. & Gómez-Llano, M. (2023) Ecological differentiation, interference, and coexistence in odonates. In: Córdoba-Aguilar, A., Beatty, C.D., Bried, J.T. & Suárez-Tovar, C.M. (Eds.) *Dragonflies and damselflies: model organisms for ecological and evolutionary research*. Oxford: Oxford University Press, pp. 203–218.
- Gröning, J. & Hochkirch, A. (2008) Reproductive interference between animal species. *The Quarterly Review of Biology*, 83, 257–282.
- Guisan, A., Tingley, R., Baumgartner, J.B., Naujokaitis-Lewis, I., Sutcliffe, P.R., Tulloch, A.I.T. et al. (2013) Predicting species distributions for conservation decisions. *Ecology Letters*, 16, 1424–1435.
- Hadfield, J.D. (2010) MCMC methods for multi-response generalized linear mixed models: the MCMCglmm R package. *Journal of Statistical Software*, 33, 1–22.
- Hair, J.F., Risher, J.J., Sarstedt, M. & Ringle, C.M. (2019) When to use and how to report the results of PLS-SEM. *European Business Review*, 31, 2–24.
- Herrmann, N.C., Stroud, J.T. & Losos, J.B. (2021) The evolution of ‘ecological release’ into the 21st century. *Trends in Ecology & Evolution*, 36, 206–215.
- Hijmans, A.R.J., Phillips, S., Leathwick, J., Elith, J. & Hijmans, M.R.J. (2021) *dismo: species distribution modeling*.
- Hutchinson, G.E. (1957) Concluding remarks. *Cold Spring Harbor Symposia on Quantitative Biology*, 22, 415–427.
- Inman, R., Fotheringham, A.S., Franklin, J., Esque, T., Edwards, T. & Nussear, K. (2019) Local niche differences predict genotype associations in sister taxa of desert tortoise. *Diversity and Distributions*, 25, 1194–1209.
- Jaime, R., Alcántara, J.M., Bastida, J.M. & Rey, P.J. (2015) Complex patterns of environmental niche evolution in Iberian columbines (genus *Aquilegia*, Ranunculaceae). *Journal of Plant Ecology*, 8, 457–467.
- Johnson, C. (1973) Distributional patterns and their interpretation in *Hetaerina* (Odonata: Calopterygidae). *Florida Entomologist*, 56, 24–42.
- Kambach, S., Lenoir, J., Decocq, G., Welk, E., Seidler, G., Dullinger, S. et al. (2019) Of niches and distributions: range size increases with niche breadth both globally and regionally but regional estimates poorly relate to global estimates. *Ecography*, 42, 467–477.
- Katsuhara, K.R., Tachiki, Y., Iritani, R. & Ushimaru, A. (2021) The eco-evolutionary dynamics of prior selfing rates promote coexistence without niche partitioning under conditions of reproductive interference. *Journal of Ecology*, 109, 3916–3928.
- Kirkpatrick, M. & Barton, N.H. (1997) Evolution of a species' range. *The American Naturalist*, 150, 1–23.
- Kishi, S. & Nakazawa, T. (2013) Analysis of species coexistence mediated by resource competition and reproductive interference. *Population Ecology*, 55, 305–313.
- Kishi, S., Nishida, T. & Tsubaki, Y. (2009) Reproductive interference determines persistence and exclusion in species interactions. *The Journal of Animal Ecology*, 78, 1043–1049.
- Kuno, E. (1992) Competitive exclusion through reproductive interference. *Researches on Population Ecology*, 34, 275–284.
- Kusch, J. (2015) Interacting influences of climate factors and land cover types on the distribution of caddisflies (Trichoptera) in streams of a central European low mountain range. *Insect Conservation and Diversity*, 8, 92–101.
- Kyogoku, D. (2015) Reproductive interference: ecological and evolutionary consequences of interspecific promiscuity. *Population Ecology*, 57, 253–260.
- Kyogoku, D. & Nishida, T. (2012) The presence of heterospecific males causes an Allee effect. *Population Ecology*, 54, 391–395.
- Kyogoku, D. & Wheatcroft, D. (2020) Heterospecific mating interactions as an interface between ecology and evolution. *Journal of Evolutionary Biology*, 33, 1330–1344.
- Lancaster, L.T. (2022) On the macroecological significance of eco-evolutionary dynamics: the range shift-niche breadth hypothesis. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 377, 20210013.
- Leibold, M.A. & McPeck, M.A. (2006) Coexistence of the niche and neutral perspectives in community ecology. *Ecology*, 87, 1399–1410.
- Liu, S.-S., De Barro, P.J., Xu, J., Luan, J.-B., Zang, L.-S., Ruan, Y.-M. et al. (2007) Asymmetric mating interactions drive widespread

- invasion and displacement in a whitefly. *Science*, 1979(318), 1769–1772.
- Losos, J.B. (2011) Seeing the forest for the trees: the limitations of phylogenies in comparative biology. *The American Naturalist*, 177, 709–727.
- Lumbreras, A., Pardo, C. & Molina, J.A. (2013) Bioindicator role of aquatic *Ranunculus* in Mediterranean freshwater habitats. *Aquatic Conservation*, 23, 582–593.
- McEachin, S., Drury, J.P., Anderson, C.N. & Grether, G.F. (2021) Mechanisms of reduced interspecific interference between territorial species. *Behavioral Ecology*, 33, 126–136.
- Megna, Y.S., Pallarés, S. & Sánchez-Fernández, D. (2021) Conservation of aquatic insects in Neotropical regions: a gap analysis using potential distributions of diving beetles in Cuba. *Aquatic Conservation*, 31, 2714–2725.
- Melo-Merino, S.M., Reyes-Bonilla, H. & Lira-Noriega, A. (2020) Ecological niche models and species distribution models in marine environments: a literature review and spatial analysis of evidence. *Ecological Modelling*, 415, 108837.
- Merow, C., Smith, M.J. & Silander, J.A. (2013) A practical guide to MaxEnt for modeling species' distributions: what it does, and why inputs and settings matter. *Ecography*, 36, 1058–1069.
- Morales, N.S., Fernández, I.C. & Baca-González, V. (2017) MaxEnt's parameter configuration and small samples: are we paying attention to recommendations? A systematic review. *PeerJ*, 5, e3093.
- Naimi, B., Hamm, N.A.S., Groen, T.A., Skidmore, A.K. & Toxopeus, A.G. (2014) Where is positional uncertainty a problem for species distribution modelling? *Ecography*, 37, 191–203.
- Namyatova, A.A. (2020) Climatic niche comparison between closely related trans-Palaearctic species of the genus *Orthocephalus* (Insecta: Heteroptera: Miridae: Orthotylinae). *PeerJ*, 8, e10517.
- Noriyuki, S. & Osawa, N. (2016) Reproductive interference and niche partitioning in aphidophagous insects. *Psyche (Stuttg)*, 2016, 4751280.
- Novella-Fernandez, R., Juste, J., Ibáñez, C., Rebelo, H., Russo, D., Alberdi, A. et al. (2021) Broad-scale patterns of geographic avoidance between species emerge in the absence of fine-scale mechanisms of coexistence. *Diversity and Distributions*, 27, 1606–1618.
- Novelo-Gutiérrez, R. (2000) Description of the larva of *Hetaerina infecta* Calvert (Odonata: Calopterygidae). *Proceedings of the Entomological Society of Washington*, 102, 99–104.
- Null, S.E., Viers, J.H., Deas, M.L., Tanaka, S.K. & Mount, J.F. (2013) Stream temperature sensitivity to climate warming in California's Sierra Nevada: impacts to Coldwater habitat. *Climatic Change*, 116, 149–170.
- Ovaskainen, O., Abrego, N., Halme, P. & Dunson, D. (2016) Using latent variable models to identify large networks of species-to-species associations at different spatial scales. *Methods in Ecology and Evolution*, 7, 549–555.
- Pateiro-López, B. & Rodríguez-Casal, A. (2010) Generalizing the convex hull of a sample: the R package alphahull. *Journal of Statistical Software*, 34, 1–28.
- Patterson, C.W. & Drury, J.P. (2023) Interspecific behavioural interference and range dynamics: current insights and future directions. *Biological Reviews*, 98, 2012–2027.
- Pearson, R.G. & Dawson, T.P. (2003) Predicting the impacts of climate change on the distribution of species: are bioclimate envelope models useful? *Global Ecology and Biogeography*, 12, 361–371.
- Peterson, A.T. (2011) Ecological niche conservatism: a time-structured review of evidence. *Journal of Biogeography*, 38, 817–827.
- Peterson, A.T. & Holt, R.D. (2003) Niche differentiation in Mexican birds: using point occurrences to detect ecological innovation. *Ecology Letters*, 6, 774–782.
- Peterson, A.T. & Nakazawa, Y. (2007) Environmental data sets matter in ecological niche modelling: an example with *Solenopsis invicta* and *Solenopsis richteri*. *Global Ecology and Biogeography*, 17, 135–144.
- Pfennig, D.W. & Pfennig, K.S. (2012) *Evolution's wedge. Competition and the origins of diversity*. Berkeley, CA: University of California Press.
- Phillips, S.J., Anderson, R.P., Dudík, M., Schapire, R.E. & Blair, M.E. (2017) Opening the black box: an open-source release of Maxent. *Ecography*, 40, 887–893.
- Phillips, S.J., Anderson, R.P. & Schapire, R.E. (2006) Maximum entropy modeling of species geographic distributions. *Ecological Modelling*, 190, 231–259.
- Phillips, S.J. & Dudík, M. (2008) Modeling of species distributions with Maxent: new extensions and a comprehensive evaluation. *Ecography*, 31, 161–175.
- Poggiato, G., Münkemüller, T., Bystrova, D., Arbel, J., Clark, J.S. & Thuiller, W. (2021) On the interpretations of joint modeling in community ecology. *Trends in Ecology & Evolution*, 36, 391–401.
- Ribeiro, J.M.C. & Spielman, A. (1986) The satyr effect: a model predicting parapatry and species extinction. *The American Naturalist*, 128, 513–528.
- Sexton, J.P., McIntyre, P.J., Angert, A.L. & Rice, K.J. (2009) Evolution and ecology of species range limits. *Annual Review of Ecology, Evolution, and Systematics*, 40, 415–436.
- Shah, D.N., Tonkin, J.D., Haase, P. & Jähnig, S.C. (2015) Latitudinal patterns and large-scale environmental determinants of stream insect richness across Europe. *Limnologia*, 55, 33–43.
- Siepielski, A.M., Hung, K.-L., Bein, E.E.B. & McPeck, M.A. (2010) Experimental evidence for neutral community dynamics governing an insect assemblage. *Ecology*, 91, 847–857.
- Sillero, N., Arenas-Castro, S., Enriquez-Urzelai, U., Vale, C.G., Sousa-Guedes, D., Martínez-Freiria, F. et al. (2021) Want to model a species niche? A step-by-step guideline on correlative ecological niche modelling. *Ecological Modelling*, 456, 109671.
- Sjödin, H., Ripa, J. & Lundberg, P. (2018) Principles of niche expansion. *Proceedings of the Royal Society B: Biological Sciences*, 285, 20182603.
- Slatyer, R.A., Hirst, M. & Sexton, J.P. (2013) Niche breadth predicts geographical range size: a general ecological pattern. *Ecology Letters*, 16, 1104–1114.
- Söderbäck, B. (1995) Replacement of the native crayfish *Astacus astacus* by the introduced species *Pacifastacus leniusculus* in a Swedish lake: possible causes and mechanisms. *Freshwater Biology*, 33, 291–304.
- Strandring, S., Sánchez-Herrera, M., Guillermo-Ferreira, R., Ware, J.L., Vega-Sánchez, Y.M., Clement, R. et al. (2022) Evolution and biogeographic history of rubyspot damselflies (Hetaeriniinae: Calopterygidae: Odonata). *Diversity (Basel)*, 14, 757.
- Suárez-Mota, M.E. & Villaseñor, J.L. (2020) Ecological niche overlap among species of the genus *Zaluzania* (Asteraceae) from the dry regions of Mexico. *Plant Ecology and Evolution*, 153, 337–347.
- Sun, D.-B., Li, J., Liu, Y.-Q., Crowder, D.W. & Liu, S.-S. (2014) Effects of reproductive interference on the competitive displacement between two invasive whiteflies. *Bulletin of Entomological Research*, 104, 334–346.
- Sundar, S., Silva, D.P., de Oliveira Roque, F., Simião-Ferreira, J. & Heino, J. (2021) Predicting climate effects on aquatic true bugs in a tropical biodiversity hotspot. *Journal of Insect Conservation*, 25, 229–241.
- Takahashi, Y., Takakura, K. & Kawata, M. (2016) Spatial distribution of flower color induced by interspecific sexual interaction. *PLoS ONE*, 11, e0164381.
- Ting, J.J. & Cutter, A.D. (2018) Demographic consequences of reproductive interference in multi-species communities. *BMC Ecology*, 18, 46.
- Tobias, J.A., Cornwallis, C.K., Derryberry, E.P., Claramunt, S., Brumfield, R.T. & Seddon, N. (2014) Species coexistence and the dynamics of phenotypic evolution in adaptive radiation. *Nature*, 506, 359–363.
- Tomašových, A., Jablonski, D. & Sandel, B. (2017) Decoupling of latitudinal gradients in species and genus geographic range

- size: a signature of clade range expansion. *Global Ecology and Biogeography*, 26, 288–303.
- Valavi, R., Guillera-Arroita, G., Lahoz-Monfort, J.J. & Elith, J. (2022) Predictive performance of presence-only species distribution models: a benchmark study with reproducible code. *Ecological Monographs*, 92, e01486.
- Vandermeer, J.H. (1972) Niche theory. *Annual Review of Ecology and Systematics*, 3, 107–132.
- Vega-Sánchez, Y.M., Mendoza-Cuenca, L.F. & González-Rodríguez, A. (2019) Complex evolutionary history of the American Rubyspot damselfly, *Hetaerina americana* (Odonata): evidence of cryptic speciation. *Molecular Phylogenetics and Evolution*, 139, 106536.
- Vega-Sánchez, Y.M., Mendoza-Cuenca, L.F. & González-Rodríguez, A. (2020) *Hetaerina calverti* (Odonata: Zygoptera: Calopterygidae) sp. nov., a new cryptic species of the American Rubyspot complex. *Zootaxa*, 4766, 485–497.
- Waller, J.T. & Svensson, E.I. (2017) Body size evolution in an old insect order: No evidence for Cope's Rule in spite of fitness benefits of large size. *Evolution (N Y)*, 71, 2178–2193.
- Warren, D.L., Beaumont, L.J., Dinnage, R. & Baumgartner, J.B. (2019) New methods for measuring ENM breadth and overlap in environmental space. *Ecography*, 42, 444–446.
- Warren, D.L., Cardillo, M., Rosauer, D.F. & Bolnick, D.I. (2014) Mistaking geography for biology: inferring processes from species distributions. *Trends in Ecology & Evolution*, 29, 572–580.
- Warren, D.L., Matzke, N.J., Cardillo, M., Baumgartner, J.B., Beaumont, L.J., Turelli, M. et al. (2021) ENMTools 1.0: an R package for comparative ecological biogeography. *Ecography*, 44, 504–511.
- Weir, J.T. & Price, T.D. (2011) Limits to speciation inferred from times to secondary sympatry and ages of hybridizing species along a latitudinal gradient. *The American Naturalist*, 177, 462–469.
- Wellenreuther, M., Larson, K.W. & Svensson, E.I. (2012) Climatic niche divergence or conservatism? Environmental niches and range limits in ecologically similar damselflies. *Ecology*, 93, 1353–1366.
- Westfall, M.J., Jr. & May, M.L. (1996) *Damselflies of North America*. Gainesville: Scientific Publishers, Inc.
- Wilkinson, D.P., Golding, N., Guillera-Arroita, G., Tingley, R. & McCarthy, M.A. (2019) A comparison of joint species distribution models for presence-absence data. *Methods in Ecology and Evolution*, 10, 198–211.
- Wisz, M.S., Hijmans, R.J., Li, J., Peterson, A.T., Graham, C.H. & Guisan, A. (2008) Effects of sample size on the performance of species distribution models. *Diversity and Distributions*, 14, 763–773.
- Yoshimura, M. (2012) Effects of forest disturbances on aquatic insect assemblages. *Entomological Science*, 15, 145–154.

SUPPORTING INFORMATION

Additional supporting information can be found online in the Supporting Information section at the end of this article.

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