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Substrate-borne Bioacoustics and the Impacts of Anthropogenic Noise

By

Maggie J. Raboin

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requirements for the degree of

Doctor of Philosophy

in

Environmental Science, Policy, and Management

in the

Graduate Division

of the

University of California, Berkeley

Committee in charge:

Professor Damian O. Elias, Chair

Professor Eileen A. Lacey

Professor Justin S. Brashares

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## Abstract

### Substrate-borne Bioacoustics and the Impacts of Anthropogenic Noise

By

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Doctor of Philosophy in Environmental Science, Policy, and Management

University of California, Berkeley

Professor Damian O. Elias, Chair

Across ecosystems, animals face expanding human populations, transportation networks, and urbanization that threaten their survival. Traditional strategies for conservation often hinge on understanding an animal's habitat and population structure, with less attention is paid to the animal's "habit", their way of being in and sensing the world. However, recent appeals to consider the sensory ecology - the way organisms acquire, process, and share information - of species suggest transitioning away from questions like "What do these organisms eat?" and "Who eats them?" toward questions that ask, "In what ways do these animals find food or avoid being eaten?" For many animals, the ubiquitous answer to questions concerning the mechanism for finding food or mates, avoiding predators, choosing suitable habitats, or communicating with conspecifics is "substrate-borne sound".

Substrate-borne bioacoustics is the subset of bioacoustics that concerns the exchange of information via mechanical waves through substrates such as rock, soil, litter, or plant material. It likely long precedes airborne communication via pressure waves in evolutionary time and is used by the vast majority of terrestrial animals - over 95% of all species. Despite this, substrate-borne bioacoustics have been largely ignored in broad acoustic fields of study. For example, animals sensitive to substrate-borne sounds were long excluded from decades of research attempting to explain the diversity of acoustics across taxa (Chapter 2) and mostly left out of research investigating the impacts of anthropogenic noise on animal behavior (Chapter 3). In this dissertation, my colleagues and I do our best to address these exclusions and examine the impacts of substrate-borne anthropogenic noise on a single species; the mason spider (*Castianeira sp.*) (Chapters 4 & 5).

We found that the exclusion of substrate-borne bioacoustics from major fields of study skewed biophysical explanations of acoustic diversity across taxa and may limit understandings of the reach of anthropogenic noise. In addition, when we narrowed our focus to a single invertebrate species, we found that the impact of substrate-borne anthropogenic noise has profound effects on invertebrate behavior. Our research suggests that understanding substrate-borne bioacoustics and substrate-borne anthropogenic noise may be critically important in conserving species, communities, and ultimately, biodiversity

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## CHAPTER 1 INTRODUCTION

### *Invertebrate Bioacoustics*

In natural history museums, you can find millions of individual invertebrates, from ants and anemones to stoneflies and shrimp. All are collected with care, meticulously analyzed by morphology, labeled, and stored on a shelf. Once there, little attention is paid to the sensory life that defined their day-to-day life. At first glance, invertebrates would appear to be too strange and void of emotion to interact with sound in a meaningful way. Moreover, in contrast to mammals, most invertebrates lack protruding folds that are identifiable as ears. They don't delight us with their audible praise of the morning sun the way that birds do. Their anguish isn't conveyed through familiar bellows. But invertebrates have rich sensory lives that are much different from our own and are often dominated by sound.

Recall the cacophony of crickets singing on a warm summer night. Crickets, like humans and vertebrates, rely on airborne pressure waves (far-field sound) to sense and convey acoustic information. However, the audibility of a cricket's song to human ears is an exception among invertebrates. In fact, most of the invertebrate sounds you know, like those produced by crickets, cicadas, katydids, and grasshoppers, are departures from the invertebrate rules of sound production. Few invertebrates are large enough to efficiently move air (Bennet-Clark, 1998) and thus resort to transmitting information through vibrations of solids (substrate-borne sound), a modality of sound out of the sensory reach of humans.

Substrate-borne communication is the subset of acoustic communication that concerns the exchange of information via mechanical waves through substrates such as rock, soil, litter, or plant material. It likely long precedes airborne communication via pressure waves in evolutionary time (Hoch et al., 2006) and is used by the vast majority of terrestrial animals - over 95% of all species (Cocroft and Rodriguez, 2005). Substrate-borne signalers use a variety of mechanisms to produce sounds with frequencies below about 2,000 Hz because low frequencies experience little attenuation in solids (Bennet-Clark, 1998). As a signaling medium, substrates include many differing physical properties, boundaries, and intersecting surfaces (reviewed in Elias & Mason; 2006). This means that in traveling from sender to receiver, substrate-borne sounds can take many forms and be filtered in unexpected or unpredictable ways (Magal et al., 2000). Natural selection is thought to favor those signals that maximize transmission (Endler, 1992; Boughman, 2002), therefore substrate-borne communication is likely more complex than communication through a homogenous medium like air or water.

Invertebrate bioacoustics suggests, with emphasis, the first step to understanding sound in the natural world: human perceptions are limiting. This is reflected in the omission of substrate-borne communication from broad acoustic fields of study. For example, animals sensitive to substrate-borne sounds have largely been left out of decades of research attempting to explain the diversity of acoustics across taxa (Chapter 2) and were long excluded from research into the impacts of anthropogenic noise on animal behavior (Chapter 3; Raboin & Elias, 2019). In this dissertation, my colleagues and I do our best to address these exclusions and examine the impacts of substrate-borne anthropogenic noise on a single species; the mason spider (*Castianeira sp.*) (Chapters 4 & 5; Raboin & Elias, 2021).

### *Anthropogenic Noise*

Anthropogenic noise is widely recognized as an issue of environmental concern (Shannon et al., 2016). Produced by human activities like those associated with urbanization,

economic development, transportation networks, and recreation, anthropogenic noise now penetrates some of the quietest places on Earth (Buxton et al., 2017). In fact, over 60% of US protected lands experience noise levels double those of background noise, despite their distance from major metropolitan areas (Buxton et al., 2017). For vertebrates, the consequences of noise in natural landscapes have been found to be multifaceted, impacting mating, movement, predator-prey dynamics, and physiology (Shannon et al., 2016). However, research has mostly focused on the impacts of pressure waves on vertebrates, with the impact of anthropogenic sound on invertebrates and the acoustic modalities they rely on (mainly particle motion and substrate-borne sound) remaining largely unstudied.

Indeed, when evaluated in 2016, only 4% of the work on the impact of anthropogenic noise on animals had been on invertebrates, despite their comprising 97% of species on Earth (Shannon et al., 2016). However, recent research investigating anthropogenic noise and invertebrates suggests that the impact of noise on invertebrate behavior, physiology, and communities is likely diverse and complicated.

### *Overview of Dissertation*

My dissertation is comprised of four chapters. In Chapter 2, my co-authors and I investigate a biophysical model of the diversity of acoustic signals as it applies to substrate-borne communication. For many decades, scientists have attempted to explain acoustic signal diversity with mathematical models. Recently, patterns of acoustic communication have been attributed to metabolic rate, broadly controlled by body size and temperature. However, it was unclear how well this model applied to substrate-borne communication. We conducted a meta-analysis, including acoustic signals from over 200 species, to examine the relationship between various call characteristics (call frequency and call rate) and body size and temperature. Our results indicate that the diversity of acoustic communication across taxa cannot be attributed solely to metabolic rate, but is likely a result of many factors, including signaling medium and sound-production mechanism.

Chapter 3 provides the framework and impetus for the remaining chapters of this dissertation. Noise created by ever-growing transportation networks and human activities is ubiquitous and has penetrated some of the quietest places on Earth. However, the impact of anthropogenic noise on invertebrates has largely gone unstudied. We aimed to close this knowledge gap by integrating information about invertebrate bioacoustics, characteristics of anthropogenic noise, and the ways in which animals adapt to noise. We propose a general framework and outline predictions for constraints and adaptations specific to invertebrates in responding to anthropogenic noise.

In Chapter 4, my co-authors and I develop a foundation in which to investigate the impacts of anthropogenic substrate-borne noise on invertebrate parental care. We explore spider parental care by quantifying the function and benefits of spider nests built by mason spiders (*Castianeira* sp.), wandering spiders endemic to the Greater Yellowstone Ecosystem (GYE). While most spiders protect their offspring with silk structures (e.g., egg sacs, webs), mason spiders construct mounds made of hundreds of pebbles and leaves. Through field observations, we described this unique behavior among spiders and found that despite the immense effort of mason spiders to construct nests, they only remain for a small portion of time that offspring occupy nest-sites. Through field experiments, we determined that nests greatly reduce the rate of predation and parasitism of egg sacs and offspring mortality. However, these effects are only



present the few days following nest construction. Our study illuminates the idea that durability, the ability to withstand environmental perturbations, is a critical feature to animal structures, expanding how we think about nest evolution.

Chapter 5 builds on Chapter 4 and quantifies the impact of substrate-borne noise on mason spider mound-building behavior. Visitation to the GYE has more than doubled in the last 30 years, and with-it human activities that produce anthropogenic noise. We assessed the consequences of substrate borne noise from roads for mason spider nest building and reproductive success. We compared nest-building behaviors and offspring survival at field sites with high and low substrate-borne noise. We found that in the presence of noise, spiders spent less time building nests and were more likely to “get lost” during building. This study was the first field based empirical study demonstrating that substrate-borne noise from anthropogenic sources has important consequences for parental care behavior and offspring survival.

## CHAPTER 2 IS THERE AN ENERGETIC BASIS TO SUBSTRATE-BORNE ACOUSTIC COMMUNICATION?

Maggie Raboin and Damian O. Elias

Included here with permission from Damian O. Elias

### 2.1 ABSTRACT

Animals communicate with a diversity of sounds that minimally includes the transmission of waves and oscillation of particles in air, water, and solids with an assortment of signal characteristics produced by a range of morphological structures. For many decades, scientists have attempted to explain acoustic signal diversity with mathematical models. Recently, patterns of acoustic communication have been attributed to metabolic rate, broadly determined to be controlled by body size and temperature. However, it was unclear how well this model applies to substrate-borne communication. We conducted a meta-analysis, including acoustic signals from over 300 species, to examine the relationship between various acoustic signal characteristics, like call frequency and call rate, and body size and temperature. Our results indicate that the diversity of acoustic communication across taxa cannot be attributed solely to metabolic rate, but is likely a result of many factors, including signaling medium and sound-production mechanism.

## 2.2 INTRODUCTION

Animals communicate with an incredible diversity of sounds from the head-thumping of termites to the howling of wolves. Acoustic communication includes the transmission of waves and oscillation of particles in air, water, and solids, produced by structures that range from the soft-tissue trachea of humans to the rigid ripples of a spider's exoskeleton (Bradbury & Vehrencamp, 2011). It occurs in oceans and soils, within plant stems and leaves, over grasslands, and through forests. Animals use sound for everything from mating and foraging to caring for young and aiding social relationships (Bradbury & Vehrencamp, 2011). The ubiquity of acoustic communication across taxa exemplifies its importance in carrying out life's basic functions.

For many decades, scientists have sought to explain the diversity of animal acoustics with proximate and ultimate hypotheses (Bradbury & Vehrencamp, 2011; Gerhardt, 1994; James F. Gillooly & Ophir, 2010; Ryan & Brenowitz, 2010; Tonini et al., 2020). Patterns of acoustic communication have been attributed to variation in body size relating to the size of an open mouth or beak (Bradbury & Vehrencamp, 2011; García-Hernández & Machado, 2017; Morton, 1977; Tonini et al., 2020), neurophysiology (Bass & McKibben, 2003), and metabolic rate (James F. Gillooly & Ophir, 2010; Prestwich et al., 1989; Sanborn, 1997). Of particular interest is the hypothesis that metabolic rate, broadly controlled by temperature and body size, is responsible for signal diversity among all animals (Gillooly & Ophir, 2010). Here, the idea is that traits that govern the ability to take up and use energy for survival and reproduction also govern the ability of an animal to produce energetically costly sounds (Oberweger & Goller, 2001; Prestwich, 1994). However, models like this one, aren't often tested with the range of diversity they are trying to explain (Rosenthal et al., 2017), illustrated by the fact that none of these models have been applied to animals that use substrate-borne communication, despite its ubiquity (Cocroft & Rodriguez, 2005).

Substrate-borne communication is the subset of acoustic communication that concerns the exchange of information via mechanical waves through solids such as rock, soil, litter, and plant material. It likely long precedes airborne communication via pressure waves (Hoch et al., 2006) and is by far the most common type of acoustic communication among terrestrial animals, used by over 95% of terrestrial species (Cocroft & Rodriguez, 2005). While the majority of species that communicate with substrate-borne sounds are invertebrates (insects, arachnids, and crustaceans (Cocroft & Rodriguez, 2005; Greenfield, 2016; Hill, 2008), it is also common among mammals, reptiles, and amphibians (Hill, 2008). Recently, birds have been added to the list of potential substrate-borne signalers (Ota, 2020).

Substrates impose different constraints on acoustic communication than air or water. First, solids generally take relatively less force to induce a propagating signal than air and thus, smaller animals can more efficiently produce low frequency substrate-borne sounds than airborne sounds (Bennet-Clark, 1998). Second, solid substrates are inherently more variable than air or water. While air and water tend to be relatively homogeneous mediums, substrates are often made of many solid materials with potentially differing physical properties, architectures, boundaries, and intersecting surfaces (Casas et al., 2007; Markl, 1983). In addition, they have the potential of transmitting vibrations with many different types of waves, including longitudinal, transverse, and/or bending waves (reviewed in Elias & Mason, 2014). This means that in traveling from sender to receiver, substrate-borne sounds can take many forms and be altered, often in unexpected or unpredictable ways (Casas et al., 2007; Magal et al., 2000). Natural selection is thought to favor those signals that maximize transmission and fidelity (Boughman,

2002; Endler, 1992), therefore it is conceivable that the diversity of substrate-borne signals is more constrained by the physics of the medium than airborne or waterborne signals due to the vast forms that solids can take. Finally, animals induce substrate-borne signals in a multitude of ways including vibrating various body parts without contact to the ground or other body parts (tremulation, wing-fanning, tymbal buckling, vocalization), striking the substrate (percussion), or rubbing two body parts together (stridulation) (Hill, 2014). Each mechanism may be differentially influenced by body size, metabolic rate, neurophysiology, or temperature. This lies in stark contrast to airborne acoustic signals, where signal production most often occurs via one mechanism: vibrating membranes driven by airflow (e.g., larynx in mammals and anurans, syrinx in birds) (Bradbury & Vehrencamp, 2011).

Here, we test a general model of acoustic signaling based on principles of animal energetics, outlined by Gillooly and Ophir (2010), with a diversity of substrate-borne signals. The model predictions stem from the presumptive dependence of metabolic rate on body size and temperature for ectotherms and endotherms, alike (J. F. Gillooly et al., 2001). It supposes that a whole organism's metabolic rate per unit mass,  $B/M$ , is related to body mass  $M$  (in grams) and shows exponential temperature dependence ( $e^{-E/kT}$ ) such that

$$(1.1) \left| \frac{B}{M} \right| = b_0 M^{-1/4} e^{-E/kT},$$

where  $\overline{b_0}$  is a normalization constant specific to taxa ( $W g^{-3/4}$ ),  $E$  is the average activation energy of the respiratory complex (approx. 0.6–0.7 eV),  $k$  is Boltzmann's constant ( $8.62 \times 10^{-5} \text{ eV/K}^{-1}$ ), and  $T$  is the absolute temperature in degrees Kelvin (K) (See Gillooly et al., 2001; Allen & Gillooly 2007; and Gillooly & Ophir 2010 for further descriptions of this equation).

Building on equation 1.1 and operating under two main assumptions concerning call frequency and call rate, Gillooly and Ophir (2010) developed model predictions relating to the dependence of acoustic communication on metabolic rate and thus, body mass and temperature. They assumed (1) that call frequency and call rate are driven primarily by the muscular activity of sound producing structures rather than by the morphological features of those structures, and (2) that muscular activity of an individual is proportional to that individual's metabolic rate. Given these assumptions, they predicted that call frequency (i.e., pitch),  $f$  (cycles  $\text{s}^{-1}$ ), and call rate (inverse of call period),  $r$  (call  $\text{s}^{-1}$ ), should show the same body mass and temperature dependence as metabolic rate such that

$$(1.2) \left| f = \frac{f_0 B}{M} = f_0 b_0 M^{-1/4} e^{-E/kT} \right|$$

and

$$(1.3) \left| r = \frac{r_0 B}{M} = r_0 b_0 M^{-1/4} e^{-E/kT}, \right|$$

where  $\overline{f_0}$  and  $\overline{r_0}$  are normalization constants that represent the number of cycles per joule of metabolic energy flux through a gram of tissue (cycle  $\text{J}^{-1} \text{ g}$ ) and calls per joule of metabolic energy flux through a gram of tissue (call  $\text{J}^{-1} \text{ g}$ ), respectively. These equations predict that the relationships between the natural logarithms of temperature-corrected call frequency or call rate (i.e.,  $\ln \left( \left| f e^{E/kT} \right| \right)$  and  $\ln \left( \left| r e^{E/kT} \right| \right)$ ) and the natural logarithm of body mass are linear with slopes of -0.25. Additionally, they predict that the relationships between the natural logarithms of mass-

corrected call frequency or call rate (i.e.,  $\ln(fM^{1/4})$  and  $\ln(rM^{1/4})$ ) and inverse absolute temperature (i.e.,  $1/kT$ ) will be linear with slopes of  $-0.65$  (average of  $0.6$ – $0.7$  range).

Gillooly and Ophir (2010) originally tested this model for animals that signal using airborne and waterborne signals, almost all of which use air or water flow to produce signals (~89% of species in dataset). For this study, we tested the generality of the Gillooly and Ophir (2010) model using animals that communicate via substrate-borne signals and examined whether it is applicable across different sound production mechanisms.

## 2.3 METHODS

We collected from the literature field and laboratory data across a diversity of animals ( $n = 313$  species) that communicate with substrate-borne signals, including insects = 281, arachnids = 15, mammals = 8, amphibians = 4, malacostracans = 4, and Squamata = 1. Species varied in the mechanisms they used to produce substrate-borne sounds including percussion = 128, stridulation = 58, tremulation = 87, tymbalation = 10, wing-fanning = 13, mandible scraping = 1, web-shaking = 1, and vocalizations = 1. Of the species included in our dataset, 25 of them used a mechanism that was unknown. Species varied in size from about  $10^{-4}$ g for the plant lice *Aacanthocnema dobsoni* to  $10^6$ g for the African elephant *Loxodonta africana* and in temperature from  $18^{\circ}\text{C}$  for the plant lice *Cacopsylla picta* to  $37^{\circ}\text{C}$  for various mammals (e.g., African elephant *Loxodonta africana*, Grey seal *Halichoerus grypus*, etc.). Most of the calls considered in our dataset were for the purpose of courtship ( $n = 234$ ), however some calls were for facilitating social relationships = 38, agonistic interactions = 14, distress = 12, alarm calling = 7, foraging = 4, predator avoidance = 2, territoriality = 1, or the context was unknown = 3. We excluded taxa data that was pre-corrected for temperature (following Gillooly & Ophir, 2010).

We defined call frequency in cycles  $\text{s}^{-1}$ , (Hz) in the medium in which communication is likely taking place. For example, call frequency is the vibration of a substrate induced by one strike of a drumming series or one strike of file and plectrum in a stridulatory call. Call frequency was taken from the reported “dominant” frequency or the frequency with the highest amplitude. We defined call rate as the rate at which vibration was induced in a substrate (cycles  $\text{s}^{-1}$ , also defined as the inverse of call period). For example, the rate of strikes to a substrate during drumming or between file and plectrum in a stridulatory call. We did not include measures of call power (dB) in our dataset as this is highly variable in recordings of substrate-borne sounds and often not reported.

Ectotherm body temperatures were estimated based on ambient temperature recordings and were assumed to be  $20^{\circ}\text{C}$  if the study was done in a lab setting and no temperature measure was otherwise recorded. Temperatures of mammals were considered to be  $37^{\circ}\text{C}$  unless otherwise recorded (Gillooly et al., 2001). Mean temperature values were included in the analyses when a range of temperature values were reported. We did not include studies where temperature values ranged greater than  $10^{\circ}\text{C}$ . For invertebrates, body length rather than weight is often recorded. Therefore, we applied body-weight regressions to get an estimate of wet body weight for these taxa (Sage, 1982). We tested model predictions by conducting linear models in RStudio v. 1.3.1056 (R Core Team, 2020).

## 2.4 RESULTS AND DISCUSSION

Overall, our data on substrate-borne acoustic signaling did not support the general model put forth by Gillooly and Ophir (2010) that predicted a negative relationship between signaling frequency and mass/temperature with a slope of -0.25 and signaling rate and mass/temperature with a slope of -0.65. Instead, we found (1) no relationship between the natural logarithm of temperature-corrected call frequency and the natural logarithm of body mass (Figure 2.1a; slope: -0.10;  $r^2 = 0.03$ ; 95% CI: -0.21 to 0.01), (2) no relationship between the natural logarithm of mass-corrected call frequency and inverse absolute temperature (Figure 2.2a; 0.32 eV;  $r^2 = 0.01$ ; 95% CI: -0.46 to 1.11), (3) no relationship between the natural logarithm of temperature-corrected call rate and the natural logarithm of body mass (Figure 2.3a; slope: -0.03;  $r^2 = 0.00$ ; 95% CI: -0.12 to 0.05), and (4) no relationship between the natural logarithm of mass-corrected call rate and inverse absolute temperature (Figure 2.4a; -0.55 eV;  $r^2 = 0.01$ ; 95% CI: -1.24 to 0.13). Note that for all of these relationships, the confidence intervals included a slope of 0. The original predictions were based on the hypothesis that metabolic rate is the primary driver of acoustic signaling diversity among all animals (Gillooly et al., 2001; Gillooly & Ophir, 2010). Our data demonstrates that this is not the case for substrate-borne signals and suggests that other drivers are responsible for signal diversity in substrate-borne communication.

Substrate-borne signaling differs from airborne and waterborne signaling in several key ways including (1) the energetics of wave propagation and (2) the variation in mechanisms used to produce substrate-borne signals. One of the key differences between air- and substrate-borne signaling is the energy required to create a propagating wave. For small animals, producing low frequency signals in air is much less efficient than coupling them with substrates (Bennet-Clark, 1998). For example, an Asian elephant (*Elephas maximus*) and a jumping spider (*Habronattus doessenus*) that is less than 1/3000000 its size, produce airborne and substrate-borne signals of the same frequency (65 Hz), respectively (Elias et al., 2003; O'Connell-Rodwell et al., 2000). Comparing our data with Gillooly and Ophir (2010), we see that after controlling for size and temperature, animals that signal at low frequencies (<1000 Hz) were rare in the Gillooly and Ophir (2010) dataset but common in ours. We suggest that higher signaling efficiency at lower frequencies in substrate-borne signalers, which may decouple muscular activity from an individual's metabolic rate in a way that is not the case with airborne signalers, is one of the primary reasons that our data fail to meet model predictions (Gillooly & Ophir, 2010). This pattern also holds broadly for call rate suggesting that the relaxation of the constraints governing the production of low frequencies also affects low-rate callers. Comparing our call rate data with Gillooly and Ophir (2010), we find that our data set includes a much wider range of body mass and calling rates (Figure 2.3a).

Substrate-borne signalers have evolved a variety of sound production mechanism relative to airborne signalers (Hill, 2008) giving us the opportunity to test the predictions of Gillooly and Ophir (2010) as they relate to sound production mechanism. Gillooly and Ophir (2010) based their model on a main assumption that across acoustic signals, signal frequency and rate are more strongly dependent on muscle activity and therefore, metabolic rate, than on morphological features. However, the muscle activity necessary to produce substrate-borne signals of a given spectral frequency differ substantially if signals are being produced via mechanisms with frequency multipliers (e.g., stridulation, tymbalation) or those without (e.g., tremulation, percussion) (Bennet-Clark, 1999; Hill, 2008). Because our data seems to violate the main assumption of the Gillooly & Ophir (2010) model, we predicted that different sound production mechanisms would show different patterns in the dependence of signal characteristics on metabolic rate (temperature and mass).

We found that when grouped by sound-production mechanism, the relationship between the natural logarithm of temperature-corrected call frequency and the natural logarithm of body mass came close to the predicted negative slope of -0.25 for sounds produced by tremulation (slope: -0.24;  $r^2= 0.26$ ; 95% CI: -0.37 to -0.11) (Figure 2.1b) and that this relationship for sounds produced by tymbalation had a significantly negative slope (slope: -0.54;  $r^2= 0.68$ ; 95% CI: -0.96 to -0.12). However, we did not find a significant linear relationship for any of the other sound production mechanisms examined (Figure 2.1b). We did not find a linear relationship between the natural logarithm of mass-corrected call frequency and inverse absolute temperature for any sound production mechanism (Figure 2.2b). When grouped by sound-production mechanism, we found a negative linear relationship between the natural logarithm of temperature-corrected call rate and the natural logarithm of body mass for calls produced by wing-fanning (-1.50 eV;  $r^2$ : 0.93; 95% CI: -2.28 to -0.71) that was different from the predicted slope of -0.25 (Gillooly & Ophir, 2010). For all other sound-production mechanisms, this relationship was not statistically different from zero (Figure 2.3b). Finally, when grouped by mechanism, we found a negative linear relationship between the natural logarithm of mass-corrected call rate and inverse absolute temperature for percussive signals (-1.42 eV;  $r^2$ : 0.16; 95% CI: -2.13 to -0.71) although this was far from the predicted value of -0.65 (Gillooly & Ophir, 2010). All other sound-production mechanisms were not statistically different from zero (Figure 2.4b).

For animals that produce calls via tremulation, we found that the relationship between mass and call frequency was consistent with the relationship predicted by Gillooly and Ophir (2010). Tremulation produces an acoustic signal when an animal rapidly vibrates their muscles of a body part at a given frequency that is directly transmitted to the substrate. Therefore, it is not a surprise that signals produced by tremulation are influenced by metabolic rate. In contrast, stridulatory mechanisms have accessory morphological structures that determine the frequencies of calls (i.e., frequency multipliers) (Bennet-Clark, 1999) and thus it is not surprising that we did not find a relationship between call frequency and body mass and temperature for these species. While call frequencies produced via tymbalation failed to meet model predictions, we did find a significant negative relationship between call frequency and body mass. In this case, we predict that surface of the tymbal (which is correlated with size) constrains call frequency such that larger animals are able to produce lower frequency calls. Interestingly, the slope for tymbalation is much greater than that of other mechanisms (including the predicted slope from Gillooly and Ophir (2010)) suggesting that signals produced via tymbalation experience more extreme size constraints on call frequency. Much more work is needed to examine these patterns.

Our findings included additional findings that were striking. First, we found that the relationship between calling frequency and body mass and calling rate and body mass is not equivalent, as predicted by the model (Gillooly & Ophir, 2010). This suggests that the energetics behind calling at low frequencies and calling at low rates are different for substrate-borne signalers than airborne signalers. Second, we found negative relationships between signal characteristics and body mass/temperature for some signal production mechanisms and not for others. Negative relationships between body mass/temperature and signal characteristics are thought to reflect intrinsic trade-offs that are used by choosers to select high quality mates (Bradbury & Vehrencamp, 2011; Gerhardt & Huber, 2002; McLean et al., 2012). A classic example is the oft-cited negative relationship between body mass and call frequency - considered to be a major “rule” in animal communication (Bradbury & Vehrencamp, 2011; Gerhardt & Huber, 2002; McLean et al., 2012; but see Rodríguez et al., 2015). Our results suggest that this is

not a broad rule across substrate-borne acoustic communication but is only relevant for signals produced via tremulation and tymbalation.

We would be remiss to ignore the fact that biotremology, the study of substrate-borne communication, is a relatively new field of study with little consistency in the way that terms associated with call characteristics are applied across taxa and sound production mechanisms (Hill & Wessel, 2016). It is a great challenge to identify analogous components of calls produced with the same mechanism but by different species, let alone calls that are produced by different sound-production mechanisms. What distinguishes call frequency and call rate from one another, but also from other characteristics of calls (i.e., syllables, transients, trains, bouts, etc.)? For example, species that produce calls with stridulation may produce a single coherent sound or a series of sounds based on how fast or slow their file teeth impact the plectrum (Bennet-Clark, 1998). In this scenario, without knowing the specific motions that produce each sound component, call frequency and call rate can easily be confused. Additionally, new substrate-borne signalers are being found each day (Benediktov et al., 2020; Quiroga et al., 2020), many of which the mechanism of sound production is unknown, and best-practices for recording and analyzing substrate-borne communication among them is constantly evolving (Raboin & Elias, 2019; Stritih-Peljhan & Virant-Doberlet, 2021). All of which makes for a field of study that is exciting, progressing, and full of discovery, but difficult to generalize across. Moving forward, it will be important to standardize terminology to better understand the broad principles governing substrate-borne communication.

Our findings suggest that the diversity of acoustic communication across taxa is likely a result of many important factors, of which metabolism is one. Substrate-borne signals are less dependent on energetics than airborne signals and the degree to which call frequency or call rate is influenced by metabolic rate depends on the mechanism of sound production. In considering a broad theoretical framework that attempts to explain the diversity of all acoustic communication, including substrate-borne communication, we recommend a theoretical framework contrary to Gillooly and Ophir (2010), one that is more complex and incorporates medium, mechanism, and metabolism. However, standardized terminology of signal characteristics across acoustic communication and a better understanding of substrate-borne communication, in general, is necessary. Over 200,000 described insect species are estimated to use substrate-borne signals (Cocroft & Rodriguez, 2005; Hill & Wessel, 2016), not to mention the many arachnid, crustacean, amphibian, and mammalian species that do the same. However, only a vanishingly small fraction of these species has been studied. Only with these data can we begin to understand general principles governing acoustic communication.

#### *Acknowledgements*

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## 2.5 FIGURES

Figure 2.1a

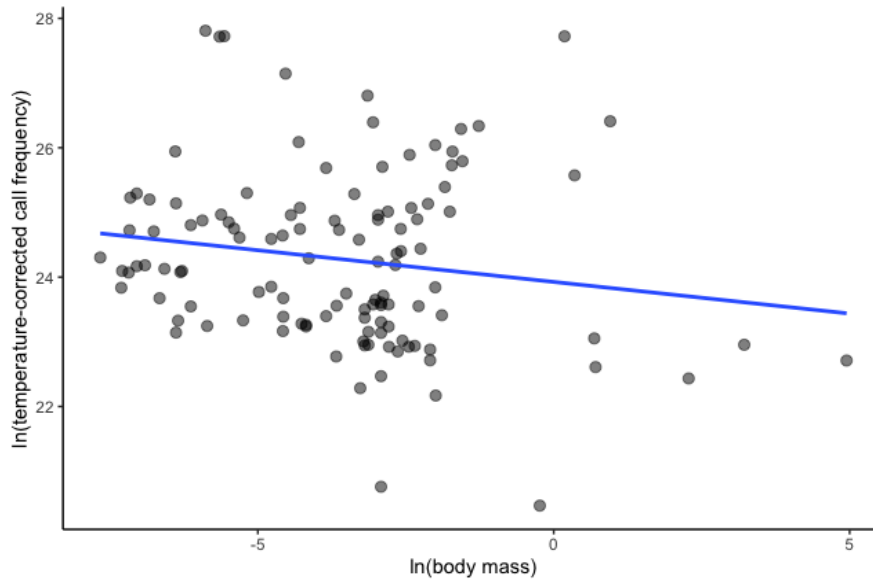
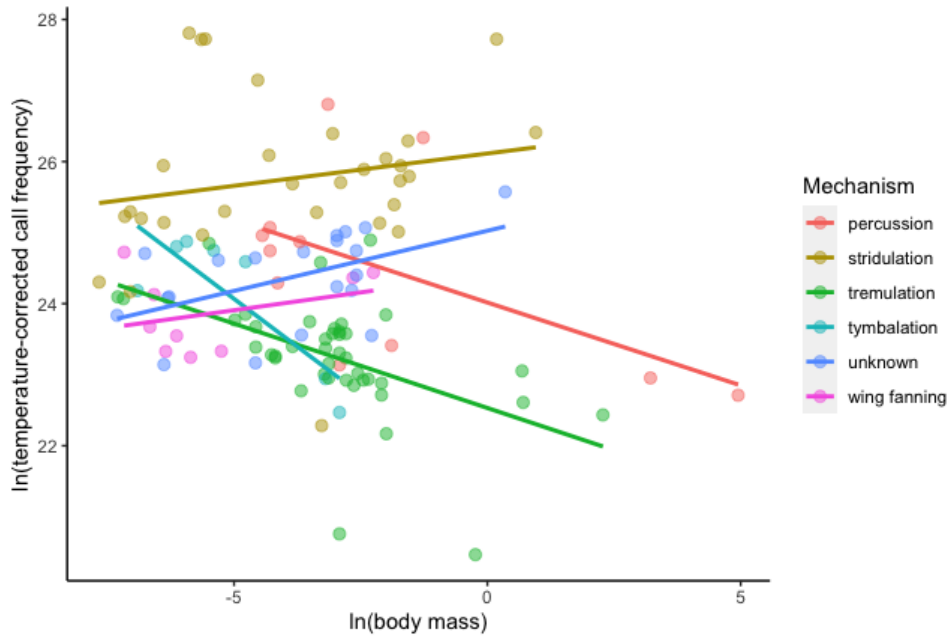


Figure 2.1b



**Figure 2.1.** Relationship between the natural logarithm of temperature-corrected call frequency (cycles  $s^{-1} \cdot e^{E/kT}$ ; equation (1.2)) and the natural logarithm of body mass (in grams) (a) across all substrate-borne signals ( $y = -0.10x + 23.93$ ;  $r^2 = 0.03$ ;  $n = 119$ ) and (b) when grouped by sound production mechanism. Note that equation (1.2) predicts a slope of  $-0.25$ . Percussion,  $y = -0.23x + 24.02$ ;  $r^2 = 0.31$ ;  $n = 11$ ; stridulation,  $y = 0.09x + 26.13$ ;  $r^2 = 0.03$ ;  $n = 30$ ; tremulation,  $y = -0.24x + 22.53$ ;  $r^2 = 0.26$ ;  $n = 42$ ; tymbalation,  $y = -0.54x + 21.38$ ;  $r^2 = 0.68$ ;  $n = 7$ ; unknown,  $y = 0.17x + 25.03$ ;  $r^2 = 0.25$ ;  $n = 20$ ; wing fanning,  $y = 0.10x + 24.41$ ;  $r^2 = 0.10$ ;  $n = 10$ .

Figure 2.2a

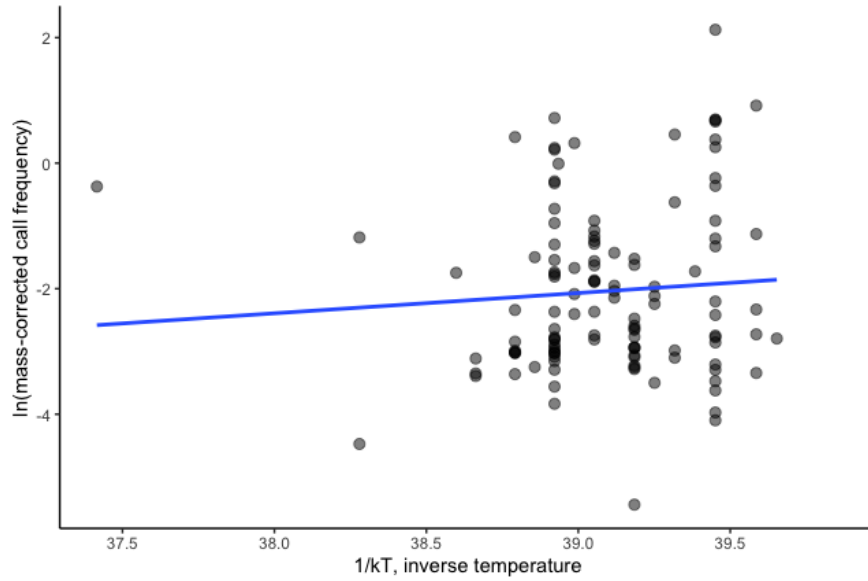
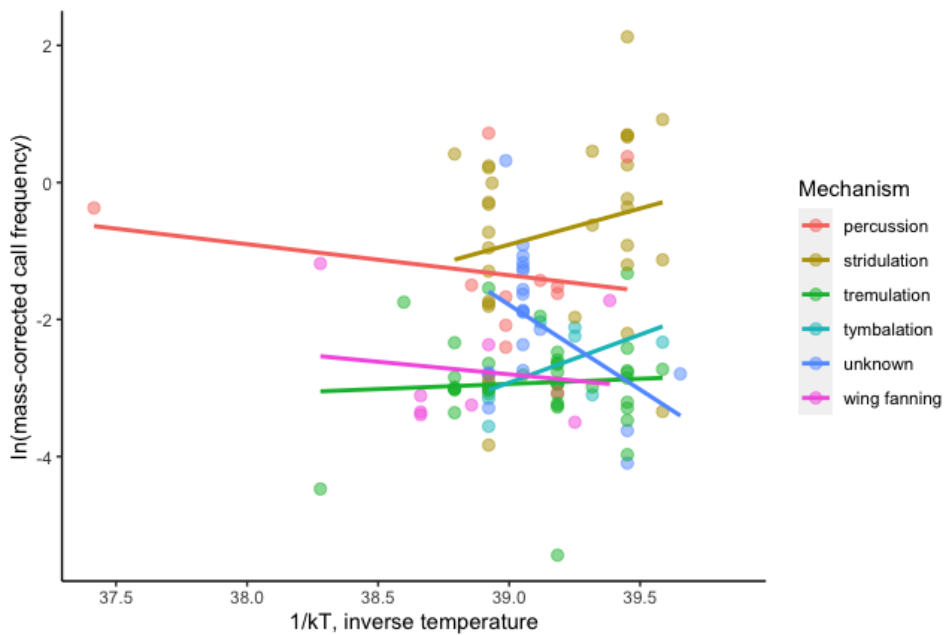


Figure 2.2b



**Figure 2.2.** Relationship between the natural logarithm of mass-corrected call frequency (cycles  $s^{-1} \cdot M^{1/4}$ ; equation (1.2)) and inverse absolute temperature,  $1/kT$ , where  $k$  is Boltzmann's constant and  $T$  is the absolute temperature in degrees Kelvin, (a) across all substrate-borne signals ( $y = 0.32x - 14.68$ ;  $r^2 = 0.01$ ;  $n = 119$ ) and (b) when grouped by sound production mechanism. Note that equation (1.2) predicts a slope of  $-0.65$ . Percussion,  $y = -0.45x + 16.32$ ;  $r^2 = 0.04$ ;  $n = 11$ ; stridulation,  $y = 1.05x - 42.05$ ;  $r^2 = 0.05$ ;  $n = 30$ ; tremulation,  $y = 0.15x - 8.81$ ;  $r^2 = 0.00$ ;  $n = 42$ ; tymbalation,  $y = 1.40x - 57.58$ ;  $r^2 = 0.29$ ;  $n = 7$ ; unknown,  $y = -2.48x + 94.79$ ;  $r^2 = 0.20$ ;  $n = 20$ ; wing fanning,  $y = -0.37x + 11.50$ ;  $r^2 = 0.02$ ;  $n = 9$ .

Figure 2.3a

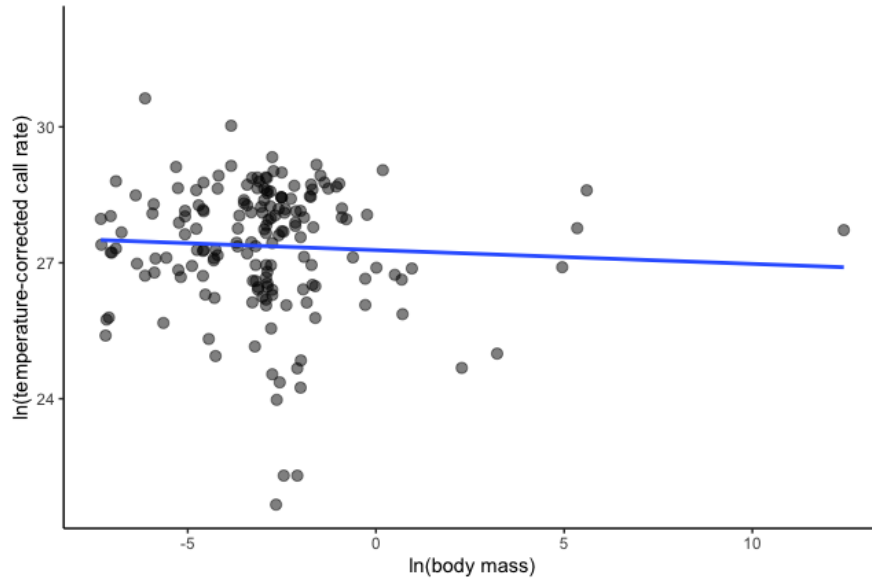
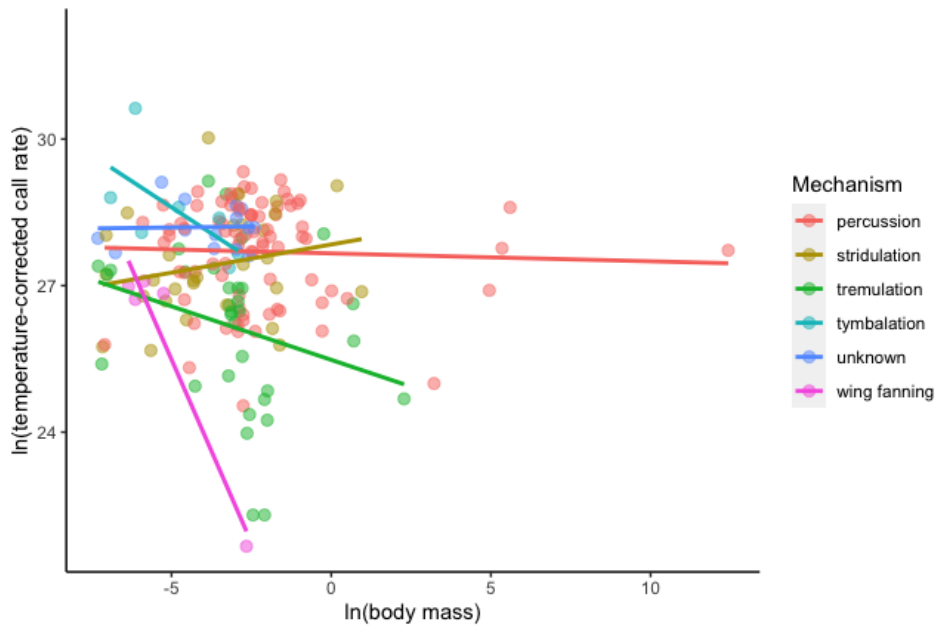


Figure 2.3b



**Figure 2.3.** Relationship between the natural logarithm of temperature-corrected call rate (calls  $s^{-1} \cdot e^{E/kT}$ ; equation (1.3)) and the natural logarithm of body mass (in grams) (a) across all substrate-borne signals ( $y = -0.03x + 27.28$ ;  $r^2 = 0.00$ ;  $n = 171$ ) and (b) when grouped by sound production mechanism. Note that equation (1.3) predicts a slope of -0.25. Percussion,  $y = -0.02x + 27.66$ ;  $r^2 = 0.00$ ;  $n = 83$ ; stridulation,  $y = 0.12x + 27.84$ ;  $r^2 = 0.05$ ;  $n = 29$ ; tremulation,  $y = -0.22x + 25.48$ ;  $r^2 = 0.08$ ;  $n = 33$ ; tymbalation,  $y = -0.43x + 26.46$ ;  $r^2 = 0.41$ ;  $n = 7$ ; unknown,  $y = 0.01x + 28.23$ ;  $r^2 = 0.00$ ;  $n = 14$ ; wing fanning,  $y = -1.50x + 17.99$ ;  $r^2 = 0.93$ ;  $n = 5$ .

Figure 2.4a

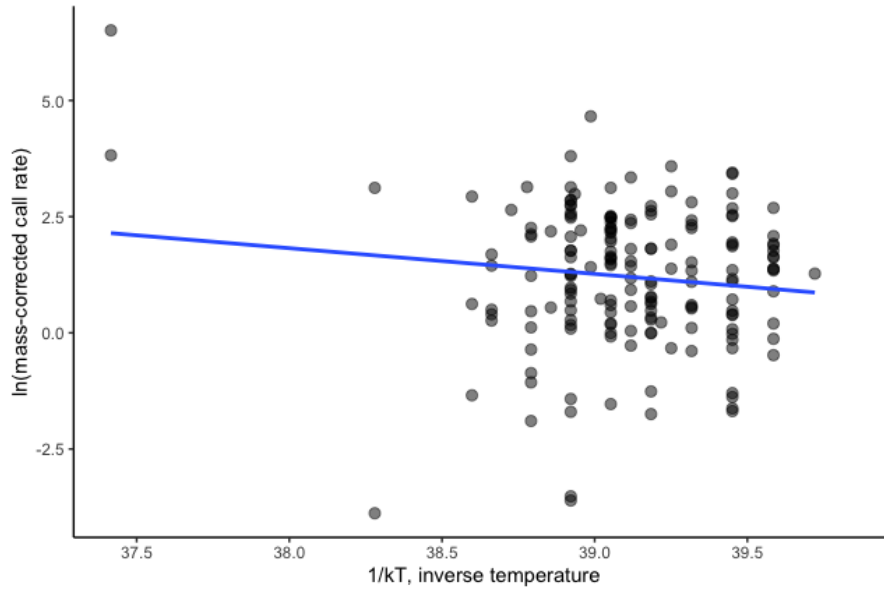
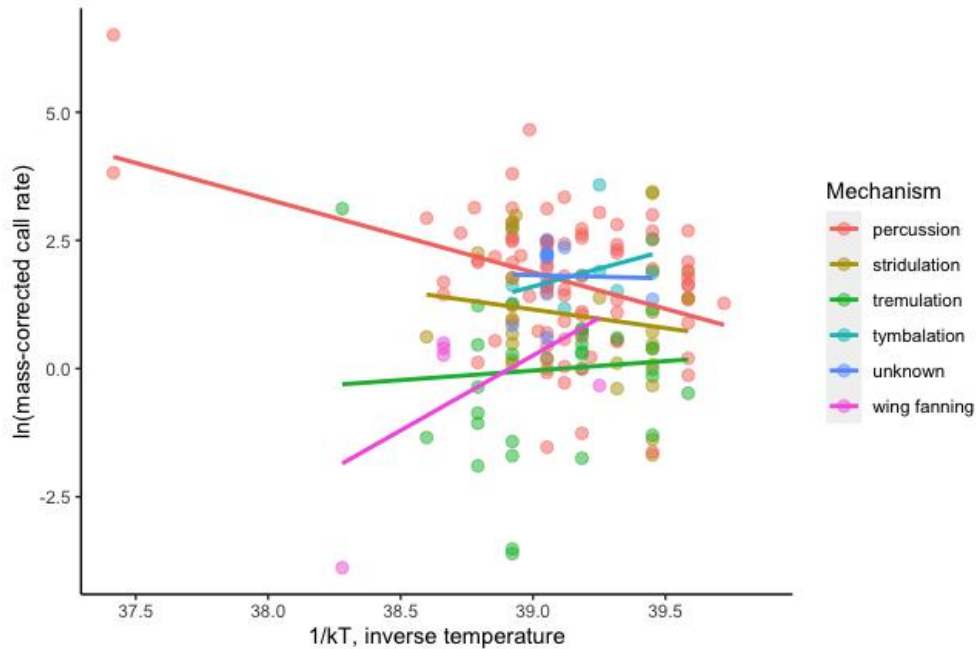


Figure 2.4b



**Figure 2.4.** Relationship between the natural logarithm of mass-corrected call rate (calls  $s^{-1} \cdot M^{1/4}$ ; equation (1.3)) and inverse absolute temperature,  $1/kT$ , where  $k$  is Boltzmann's constant and  $T$  is the absolute temperature in degrees Kelvin, (a) across all substrate-borne signals ( $y = -0.55x + 22.86$ ;  $r^2 = 0.01$ ;  $n = 171$ ) and (b) when grouped by sound production mechanism. Note that equation (1.3) predicts a slope of  $-0.65$ . Percussion,  $y = -1.42x + 57.34$ ;  $r^2 = 0.16$ ;  $n = 83$ ; stridulation,  $y = -0.72x + 29.40$ ;  $r^2 = 0.03$ ;  $n = 29$ ; tremulation,  $y = 0.37x - 14.48$ ;  $r^2 = 0.00$ ;  $n = 33$ ; tymbalation,  $y = 1.39x - 52.74$ ;  $r^2 = 0.12$ ;  $n = 7$ ; unknown,  $y = -0.13x + 6.81$ ;  $r^2 = 0.00$ ;  $n = 14$ ; wing fanning,  $y = 2.93x - 113.85$ ;  $r^2 = 0.30$ ;  $n = 5$ .

## CHAPTER 3 ANTHROPOGENIC NOISE AND THE BIOACOUSTICS OF TERRESTRIAL INVERTEBRATES

Maggie Raboin and Damian O. Elias

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### 3.1 ABSTRACT

Anthropogenic noise is an important issue of environmental concern owing to its wide-ranging effects on the physiology, behavior and ecology of animals. To date, research has focused on the impacts of far-field airborne noise (i.e., pressure waves) on vertebrates, with few exceptions. However, invertebrates and the other acoustic modalities they rely on, primarily near-field airborne and substrate-borne sound (i.e., particle motion and vibrations, respectively) have received little attention. Here, we review the literature on the impacts of different types of anthropogenic noise (airborne far-field, airborne near-field, substrate-borne) on terrestrial invertebrates. Using literature on invertebrate bioacoustics, we propose a framework for understanding the potential impact of anthropogenic noise on invertebrates and outline predictions of possible constraints and adaptations for invertebrates in responding to anthropogenic noise. We argue that understanding the impacts of anthropogenic noise requires us to consider multiple modalities of sound and to cultivate a broader understanding of invertebrate bioacoustics.

## 3.2 INTRODUCTION

Anthropogenic acoustic noise (see Glossary, Table 3.1.) is widely recognized as an issue of environmental concern (Barber et al., 2011b; Shannon et al., 2016). Noise created by ever growing transportation networks and human activities associated with economic development has become so prevalent that it has penetrated some of the quietest places on Earth. For example, 63% of US protected public lands have background noise levels double that of environmental background noise levels (Buxton, McKenna, et al., 2017). Importantly, the acoustic characteristics of anthropogenic noise differ from those of environmental noise (e.g., sound produced by wind, rain, conspecifics, heterospecifics); thus, it represents a novel challenge for animals. For vertebrates, the impact of noise is multifaceted. It has been found to have negative consequences for mating and courtship behavior, predator–prey dynamics, movement, habitat selection and physiology (Shannon et al., 2016).

However, the impact of anthropogenic noise on invertebrates has largely gone unstudied (Morley et al., 2014; Shannon et al., 2016). Although invertebrates comprise 97% of animal species on Earth, and are important in most ecological processes, only 4% of the work on noise and wildlife has been on invertebrates (Shannon et al., 2016). Viewed with an evolutionary history perspective, this skew becomes even more dramatic. The ability to hear, which requires a specialized organ or organs, evolved from a single ancestor in vertebrates (Manley, 2012), whereas the ability to hear far-field airborne sounds (see Glossary, Table 3.1.) evolved independently at least 24 times in insects alone (Greenfield, 2016), and likely many more times in invertebrates. Invertebrates send and receive sound in highly diverse ways; thus, the ways in which anthropogenic noise affects invertebrate behavior are likely vast and complicated.

In this Review, we focus on the effects of anthropogenic noise on acoustically mediated behaviors in terrestrial invertebrates. We define acoustics in the broad sense, as any exchange of information that occurs via mechanical waves propagating in a medium (Fig. 3.1.). Acoustics in terrestrial environments are extensive and their simplest configuration includes airborne far-field sound (pressure waves), as well as airborne near-field sound (particle motion; see Glossary) and substrate-borne sound (see Glossary, Table 3.1.).

For sounds transmitted through air, the power produced via waves is a product of pressure and particle velocity (Kinsler et al., 1999). Close to the sound source (or in the near-field), particle velocity dominates, whereas further from the sound source (or far-field), pressure waves dominate (Kinsler et al., 1999). This physical phenomenon is partially driven by the fact that pressure attenuates less with distance ( $1/r$ , where  $r$  is the distance from the source) than do particle movements ( $1/r^2$ ) (Jacobsen, 2007; Kinsler et al., 1999). In a general sense, the near-field only occurs at a distance of approximately 0.5–1 wavelength from the source (Jacobsen, 2007; Kinsler et al., 1999), whereas far-field sound waves (pressure waves) can travel many meters, thus dominating long-range airborne communication in animals.

Substrate-borne sounds are waves transmitted through or on the surface of a solid substrate. Solid substrates are an inherently more complex medium for transmission than air or water. Contrary to airborne or waterborne signals (see Glossary, Table 3.1.) that travel through a single medium, substrate-borne signals often travel through multiple media with differing properties and many articulating surfaces and boundaries (Elias & Mason, 2014). Additionally, substrate-borne waves can take multiple forms beyond longitudinal waves, depending on the material, geometry and/or size of the medium (Elias & Mason, 2014). The nature of a substrate

(e.g., plants, rocks, soil, litter, wood) has major implications for the types of waves transmitted (e.g., longitudinal, transverse, bending, Rayleigh), the distance sound travels, the speed at which different frequencies travel and the optimal frequencies for transmission (Aicher & Tautz, 1990; Brownell, 1977; Hill, 2008b; Michelsen et al., 1982). The diversity of possible substrates introduces many possibilities regarding distortion in the spectral and temporal domain of signals.

The study of the interplay between the acoustics of terrestrial vertebrates and anthropogenic noise has been dominated by animals using pressure waves in air (airborne far-field) with few, but notable, exceptions (Mortimer et al., 2018; Narins, 1990, 2001; Shier et al., 2012). However, terrestrial invertebrates routinely use all three types of sound. At the same time, anthropogenic activities are known to produce noise in each modality. Here, we attempt to bridge the gap between two fields of study: invertebrate bioacoustics and anthropogenic noise. By mining known information (or current understandings) about the mechanisms of invertebrate bioacoustics, the characteristics of anthropogenic noise and the ways in which animals adapt to noise, we have come up with a framework for investigating and understanding the potential impact of anthropogenic noise on invertebrates.

We have organized our discussion into three main categories that correspond to three interrelated types of sound: far-field pressure waves, near-field particle motion and substrate-borne waves. For each category, we present what little information is known about the impact of anthropogenic noise on invertebrates. In addition, we make predictions about the ways in which noise can affect communication, including detection, by focusing on the mechanisms invertebrates use for sending and receiving acoustic information. We also discuss how invertebrates might adapt to anthropogenic noise in each acoustic modality and the various constraints inherent to each sound type. In Fig. 3.2, we lay out hypotheses on how invertebrates may adjust signal characteristics in order to avoid the impacts of anthropogenic noise if there is substantial overlap between noise and communication signals (masking; see Glossary, Table 3.1.). Finally, with the goal of more broadly understanding anthropogenic noise as an environmental issue, we outline our thoughts on the most pressing lines of inquiry for research in each modality.

### 3.3 AIRBORNE SOUND

#### **Far-field**

Of the modalities of sound that invertebrates use, airborne sound has by far received the most attention, despite the fact that it is the least common modality for acoustic communication. Seven orders of insects (Coleoptera, Diptera, Hemiptera, Lepidoptera, Mantodea, Neuroptera and Orthoptera; Greenfield, 2016) are known to use far-field sound (600 to >130,000 Hz; Schmidt & Balakrishnan, 2015). Their calls can contain information about species identity (Hoy et al., 1982), the caller's location (Cade, 1975; Doherty, 1985) and different aspects of mate quality (Tuckerman et al., 1993; Wagner & Reiser, 2000), as well as fighting ability (Brown et al., 2006). In addition, some species of flies eavesdrop on calling Orthoptera to locate and parasitize them (Cade, 1975). Finally, many species use far-field sounds to detect, avoid or deter predators. For example, some species rely on far-field ultrasound to detect and avoid echolocating bat predators when they are being pursued, whereas others produce defensive acoustic signals to ward off predators (ter Hofstede & Ratcliffe, 2016).

Thanks to a great effort to understand the impact of specific noise sources on vertebrates that use far-field sounds (Shannon et al., 2016), we have a good understanding of the noise

sources likely to impact invertebrates that communicate in the far-field. Anthropogenic noise sources such as traffic on roads and railways (10–10,000 Hz up to 50,000 Hz; Hayek, 1990; Talotte et al., 2003), oil, gas and wind development (compressor noise <20–5000 Hz, wind turbines <50,000 Hz; Barber et al., 2011; Bunkley et al., 2017; Long et al., 2011), military activities (Larkin et al., 1996) and general urban environments (0–22,000 Hz; Wood & Yezerinac, 2006) create noise that overlaps with the frequency ranges of signals and cues (see Glossary, Table 3.1.) used by invertebrates. Anthropogenic noise sources either produce intermittent noise (traffic on roads and railways, military activities, oil and gas development) or continuous noise (wind turbines, general urban environments); the constancy of the noise can have implications for masking, as well as other impacts (i.e., distraction; Table 3.2.), and likely has implications for the ability of invertebrates to habituate to particular sources of anthropogenic noise (Barber et al., 2009).

### *Impacts of airborne noise*

One of the most common and likely consequences of noise from anthropogenic activities for far-field communication is masking. Masking happens when noise co-occurs with an acoustic signal or cue, thus making it difficult or impossible for animals to assess the information encoded therein. For invertebrates where males use far-field sounds to call to females, such as field crickets and katydids, masking by road noise renders females unable to detect or locate males for mating (Bailey & Morris, 1986; Bent et al., 2018; R. Schmidt et al., 2014). Similarly, male grasshoppers that locate females by their calls are less responsive to calling females in the presence of noise but are nonetheless accurate when turning in their direction (Reichert, 2015). Masking noise also decreases the ability of the parasitoid fly *Ormea ochracea* to localize calls of their host cricket (N. Lee & Mason, 2017). For some animals, masking of courtship songs may change the preference functions of choosers. For example, in the grasshopper *Chorthippus biguttulus*, masking noise significantly increases female preference for some traits and decreases it for others, leading to a decreased overall responsiveness of females to male signals (Reichert & Ronacher, 2015). However, masking noise does not always change a chooser's preference. For example, Einhäupl et al., (2011) found that individual male songs preferred by female grasshoppers (*C. biguttulus*) without noise present also elicited responses by females at higher amplitudes of white noise than songs that were not preferred. Finally, rather than suffering the consequences of their signals being masked in the presence of noise, some invertebrates put less energy into calling by shortening calls (which likely diminishes their effectiveness) (Orci et al., 2016) or by stopping calls altogether (Costello & Symes, 2014).

Masking by anthropogenic noise might impact invertebrate predator–prey dynamics, especially between some invertebrate species and their bat predators. Recent studies have found that some bat species alter their foraging patterns (Bunkley et al., 2015; Schaub et al., 2008) or are less efficient foragers in the presence of anthropogenic noise (Siemers & Schaub, 2011a). Although this evidence suggests invertebrates experience predation relief when noise is present, the signals and cues used by invertebrates to hear and flee approaching predators may also be masked. For example, most flying insects use acoustic cues to detect predators, and some produce defensive signals to avoid or deter predators (ter Hofstede & Ratcliffe, 2016). Masking of these signals and cues might increase predation pressure on invertebrates where noise is present; however, these dynamics have yet to be tested.

### *Strategies to avoid masking*



Invertebrates that transmit far-field sounds in noisy environments have multiple ways of overcoming the challenges associated with anthropogenic noise. First, they can avoid noise spatially by moving away from areas heavily impacted by human activities. Although there is no direct evidence of this in invertebrates, Bunkley et al. (2017) found that grasshoppers and camel crickets were less abundant at gas drilling sites with noise than those without. Second, animals may avoid anthropogenic noise temporally, by calling at times when noise is absent or reduced. However, one source of noise, road traffic, is predictably most intense during dawn and dusk (rush hour), times when many invertebrates concentrate their calling activity (Luther & Gentry, 2013). Researchers have suggested that it is unlikely that invertebrates will be able to temporally shift calling because there are increased costs to calling at other times of day, owing to suboptimal atmospheric conditions, particularly at sunset (Van Staaden & Römer, 1997). In addition, short-term temporal activities for some invertebrates might be constrained by phylogenetic history (Bao-sen Shieh et al., 2015). Third, like some vertebrates, invertebrates could increase the amplitude of their calls (Lombard effect) (Nemeth & Brumm, 2010). However, to date, there is no evidence that this occurs. Finally, animals can alter their signals to avoid calling in masked frequencies or, alternatively, increase the intensity of masked frequencies. This has been found to be the case for some invertebrate species. For example, male grasshoppers and cicadas shift their songs to higher frequencies in the presence of noise (Lampe et al., 2012, 2014; Shieh et al., 2012).

On the receiving end, listening animals may be able to overcome masking by anthropogenic noise using a variety of neurological mechanisms, including frequency tuning, where receivers are most sensitive to the call frequencies that contain the most energy (Schmidt & Römer, 2011). Receivers might also use spatial release from masking (see Glossary, Table 3.1.) and/or active amplification mechanisms (Mhatre & Robert, 2013; Morley & Mason, 2015; Romer, 2013; Schmidt & Balakrishnan, 2015). All of these strategies have been demonstrated in invertebrates that contend with environmental noise, but they have yet to be found in response to anthropogenic noise. It is critical that future work seeks to understand neurological strategies for avoiding or compensating for anthropogenic noise.

### *Constraints to plasticity*

Most invertebrates produce far-field sound with hardened sclerotized structures. For example, many invertebrates produce calls by rubbing together structures on their forewings, or tegminal stridulation (see Glossary, Table 3.1.), coupled with a resonating structure. This sound-production mechanism requires a hardened file of teeth and a scraper that is dragged across the teeth to produce vibrations (see Glossary, Table 3.1.; Bennet-Clark, 1999; Koch et al., 1988; Montealegre-Z & Mason, 2005). For species that use this mechanism, the frequencies they can produce are driven by the shape of the hardened structures as well as the speed of the scraper movement (Bennet-Clark, 1999; Chivers et al., 2017; Koch et al., 1988). Song frequency is also highly dependent on resonator geometry, with some tree crickets able to produce different frequencies using multiple resonant modes of their wings (Mhatre et al., 2012). Regardless, it is likely that most invertebrate species are constrained in their ability to plastically adjust the spectral properties of signals. Interestingly, Lampe et al., (2014) found developmental plasticity in response to anthropogenic noise, whereby grasshoppers raised under noisy conditions produce higher-frequency songs as adults. Similar predictions can also be made for invertebrates that use other sound-producing mechanisms such as tymbalation, percussion, fluid compression and aerodynamic sound, as these also use sclerotized structures to produce sound (Chapman, 1998).

Across and within invertebrate species that produce far-field sound, active plasticity in response to noise may be partly limited by constraints resulting from their small body sizes. The frequency and intensity of far-field calls that insects can produce is highly correlated with size (Bennet-Clark, 1998; Simmons, 1995). In general, larger individuals can produce louder, lower-frequency sounds and receive lower-frequency sounds than smaller invertebrates (Bennet-Clark, 1998); thus, call frequency is often an honest signal of size in insects (Simmons, 1995). These dynamics will likely limit the use of frequency-shift mechanisms in response to anthropogenic noise by some insects.

Finally, it is important to consider the environment and ecological community in which invertebrates are calling when evaluating the potential impacts of, and adaptations to, masking by anthropogenic noise. Invertebrates face constraints to the frequencies and timing of their calling brought on by habitat structure (attenuation and distortion), masking by other calling animals and predation (Romer, 2013; Schmidt & Balakrishnan, 2015). For example, many crickets and katydids live in assemblages where they are already partitioning the spectral or temporal aspects of their calls so that they do not overlap with one another (Gogala & Riede, 1995; Jain et al., 2014; Schmidt et al., 2011). In these communities, it is unlikely that all species will be able to avoid masking by anthropogenic noise while maintaining their spectral or temporal partitioning.

### **Near-field**

Near-field sound attenuates quickly relative to far-field sound, and conventional thought has been that its detection by terrestrial invertebrates is physiologically limited to less than a wavelength from the source (Kinsler et al., 1999). Thus, research on near-field communication among terrestrial invertebrates has primarily focused on a limited number of animals that communicate at close range and detect near-field sound via the Johnston's organ (see Glossary, Table 3.1.; flies, mosquitoes, bees; Gibson & Russell, 2006; Göpfert & Robert, 2002; Tsujiuchi et al., 2007) or filiform hairs (e.g., trichobothria) sensitive to particle motion (crickets, spiders; Barth, 2000; Kämper & Kleindienst, 1990). However, many terrestrial invertebrates are covered in innervated hairs, and particle motion is a rich source of information; therefore, many invertebrates are likely to have some capacity for detecting particle motion that has been underappreciated. This hypothesis, however, needs to be assessed further.

Near-field receptors are exquisitely sensitive, with deflections as small as 1 Å triggering neural responses (Humphrey & Barth, 2008; Shimosawa & Kanou, 1984). These receptors are generally tuned to low frequencies (<500 Hz) (Shamble et al., 2016; Wang et al., 2000). The tuning of each receptor depends primarily on its length, diameter and mass (Barth et al., 1993). Invertebrates use near-field receptors in mating interactions (Lapshin & Vorontsov, 2017; Tauber & Eberl, 2003), to forage (Barth & Höller, 1999; Kirchner, 1994), to detect incoming predators (Tautz & Markl, 1978) and during antagonistic interactions (Santer & Hebets, 2008).

Near-field anthropogenic noise has been overlooked in the literature thus far. However, all sources of airborne anthropogenic noise produce particle motion, as it is inherent to the production of airborne sound (Kinsler et al., 1999). Thus, in the presence of noise, invertebrates that communicate with near-field sound may be subject to a host of consequences, from masking to injury. Noise sources that have a large amount of energy in frequencies below 500 Hz, such as noise from roads (Hayek, 1990), railways (Talotte et al., 2003), and oil and gas development (Barber et al., 2011b), are the most likely candidates to produce near-field noise relevant to invertebrates.

At present, no direct evidence exists to support the idea that anthropogenic noise impacts near-field communication. However, one study by Samarra et al., (2009) found that near-field white noise hinders the ability of female *Drosophila montana* to detect and recognize male courtship song when it falls within the same frequency bands. In addition, a host of studies contain new information that expands our understanding of near-field communication in ways that suggest invertebrates may be less robust to the effects of anthropogenic near-field noise than previously thought.

First, near-field communication may take place at much longer distances than once suspected. Previous physical modeling and measurements suggested that near-field communication could only occur across small distances (1–70 cm), such as when animals are flying next to each other in a swarm (Aldersley et al., 2017) or interacting at close distances (Santer & Hebets, 2008; Tauber & Eberl, 2003). However, Shamble et al. (2016) demonstrated with behavioral and physiological data that jumping spiders can detect acoustic energy at distances of at least 3 m using near-field receptors. Similarly, Menda et al., (2019) demonstrated that mosquitoes can detect sound up to 10 m away using their antennae. Furthermore, Zhou & Miles (2017) presented models showing that small fibers (>1  $\mu\text{m}$ ), such as those used as near-field receptors, move with the surrounding medium. This scenario suggests that thin hairs will move in response to a large range of stimuli, even those produced at long distances. Additionally, some invertebrates, such as mosquitoes and flies, employ near-field receptor organs that actively amplify quiet signals and provide directional sensitivity (Gopfert et al., 2005; Göpfert & Robert, 2001; Morley et al., 2018). Together, these pieces of evidence extend the effective range of anthropogenic near-field noise to at least 10 m and potentially much further.

Second, invertebrate near-field receptors are likely able to detect a much wider range of frequencies than previously thought. In the past, research mainly focused on the detection of low frequency particle motion, because it was thought that biologically relevant cues such as wind and predator/prey movements are also dominated by low frequencies (Barth & Höller, 1999). More recent work, however, has demonstrated that near-field receptors can respond to a wider range of frequencies than previously thought (Bathellier et al., 2012; Zhou & Miles, 2017), and that biologically relevant stimuli also likely contain a broad range of frequencies (Casas et al., 2008). Zhou and Miles (2017), for example, suggested that thin fibers move in response to a large range of frequencies. For example, spider silk, an important near-field detector for many spiders, could measurably capture airflow over frequency ranges spanning infrasound to ultrasound (1–50,000 Hz) (Zhou and Miles, 2017). In another study, Bathellier et al. (2012) used particle image velocimetry (PIV) to demonstrate that spider and cricket filiform hairs are extremely sensitive at much higher frequencies than previously suspected.

### *Strategies to avoid masking*

Although more research is needed, new understandings of near-field sound and receptors suggest that terrestrial invertebrates are likely to be vulnerable to near-field noise in very fundamental ways. Nevertheless, invertebrates that experience negative impacts of anthropogenic noise have a few important tools for avoiding or mitigating these impacts. First, relative to avoiding other types of acoustic noise, it could be especially effective for invertebrates to move a few meters away from the sound source when possible. Second, some species have shown the physical ability to adjust the frequencies of their signals. For example, male and female mosquitoes modulate near-field flight frequencies during courtship (Cator et al., 2009; Gibson & Russell, 2006). Third, some invertebrates use behavioral strategies, such as clustering

flight tones in swarms during mating, to reduce acoustic interference to near-field signals (Aldersley et al., 2017). And fourth, near-field receptors of some invertebrates may have the capacity to avoid harmful effects of noise. For example, some invertebrates have nearfield receptors with active, non-linear tuning, where the tuning of receptors is amplitude dependent (Albert & Kozlov, 2016; Göpfert & Robert, 2002). In these cases, animals may use non-linear tuning to avoid injury to their receptors when noise is loud. Additionally, there is some evidence that this mechanism may be useful for spatial release from masking (see Glossary, Table 3.1.; Morley et al., 2018). These examples likely represent just a small sample of the potential strategies available to invertebrates for avoiding near-field noise. Much more research is needed to understand the prevalence of near-field communication and how it is affected by anthropogenic noise.

### 3.4 SUBSTRATE-BORNE SOUND

Substrate-borne acoustics have largely been left out of the study of anthropogenic noise and its impacts on animals. This is despite the fact that anthropogenic sources create substrate-borne noise (Dowding, 1996; Forman, 2000; Heckl et al., 1996), and over 90% of all animals use some type of substrate-borne sound (R B Cocroft & Rodriguez, 2005). Invertebrates rely on substrate-borne sounds for many important aspects of their lives, including as a way to collect information about their environment (T. A. Evans et al., 2005), to communicate with conspecifics during courtship (Elias et al., 2003; D. Ota & Čokl, 1991), in competition (De Souza et al., 2011; Elias et al., 2008; Yack et al., 2001) and cooperation (Baroni-Urbani et al., 1988; Endo et al., 2019; Michelsen et al., 1986), to detect prey during foraging (Fertin & Casas, 2007; Klärner & Barth, 1982; Pfannenstiel et al., 1995), to avoid predators (Castellanos & Barbosa, 2006; Rohrig et al., 1999) and to facilitate symbiotic relationships (DeVries, 1990).

Invertebrates that communicate with substrate-borne sound most often use frequencies <1000 Hz, because low frequencies experience little attenuation in substrates (Bennet-Clark, 1998; Čokl & Virant-Doberlet, 2003). At the same time, human activities such as airport traffic (Fidell et al., 2002), construction (Dowding, 1996) and use of railroads (Heckl et al., 1996), are known to produce low-frequency vibrations (<1000 Hz). Previous work also suggests that because the majority of spectral energy of road noise is in low frequencies (<2000 Hz), roads are likely to represent significant sources of substrate-borne noise (Forman, 2000). Substrate-borne vibrations do not attenuate quickly and can be detected up to 3000 m from the source (Mortimer et al., 2018). Given that many animals have such sensitive receptors of substrate-borne sound, they are likely to detect anthropogenic sources from even longer distances (Barth & Geethabali, 1982; Mortimer et al., 2018; Shaw, 1994). Anthropogenic sources can produce substrate-borne noise in two ways. First, sources can directly vibrate the earth, producing waves that travel through or on its surface. Second, sources may produce airborne noise that secondarily induces vibrations in substrates (sympathetic vibrations; see Glossary, Table 3.1.). In general, the former produces vibrations that are louder and will travel further than the latter, because sound loses energy at any substrate boundary (Caldwell, 2014). However, invertebrate receptors are likely to be sensitive to both types of substrate-borne sound.

Variation in the types and structures of substrates is likely to influence how, or whether, anthropogenic noise affects invertebrates that communicate with substrate-borne sound. First, the material properties and dimensions of a substrate determine the likelihood that airborne noise will produce sympathetic vibrations (Press & Ewing, 1951), the propagation of vibratory noise

(Elias & Mason, 2011) and its intensity (Hill, 2008b). For instance, substrate-borne vibrations do not propagate as well through materials such as sand as they do through plant material (Elias & Mason, 2011), so invertebrates that live on plants or leaf litter might contend with a noisier signaling environment in the presence of human activities than those that live on sand. Second, material properties and dimensions determine the resonance, attenuation and filtering of a given substrate (Kinsler et al., 1999). Invertebrates often take advantage of substrate properties to enhance the efficacy of their signaling by preferentially signaling with frequencies that transmit well through the substrate (Cocroft et al., 2010; Čokl et al., 2005; Elias et al., 2004, 2010). In some cases, vibratory noise from anthropogenic sources could stimulate the resonant characteristics of substrates and mask the frequencies used by invertebrates for communication. Complicating matters further, some substrates reflect sound waves (i.e., echoes and reverberations). For example, noise propagated through rod-like substrates (e.g., stems) could drive the production of standing and/or reflected waves that would distort the properties of signals (Michelsen et al., 1982; Miklas et al., 2001). Third, the boundary conditions of the substrates in question (i.e., substrate shape, articulating surfaces, heterogeneities in surface) can be complex and affect transmission in important ways (Magal et al., 2000). Finally, an individual's signaling environment can include a diversity of substrates, including combinations of natural or human-made materials, all of which differently influence how sound propagates (C. Wu & Elias, 2014). Currently, it is difficult to make predictions about the ways in which invertebrates will be affected by anthropogenic noise because little is known about the details of substrate-borne sound propagation in natural signaling environments (Elias & Mason, 2014).

#### *Impact of substrate-borne noise*

For the most part, researchers have investigated the impact of general substrate-borne noise disturbance (rather than anthropogenic noise) on communication by experimentally inducing white noise. Substrate-borne noise has been found to disrupt mating in a variety of contexts. For example, female stink bugs adjust the frequency of their signal in response to noise of similar frequencies but do not adjust temporal aspects of their signal when this is masked by noise (Polajnar & Čokl, 2008). In the presence of temporally disruptive substrate-borne noise, male stink bugs respond less frequently to female signals (Polajnar & Čokl, 2008). Wolf spiders that rely on substrate-borne signals for courtship are less likely to mate in the presence of white noise (Gordon & Uetz, 2012). Finally, male black-faced leafhoppers (which produce substrate-borne courtship signals with three distinct sections) cease signaling when white noise is played during the initial phase but not latter phases (Hunt & Morton, 2001). Interestingly, a handful of studies on substrate-borne noise and mating were motivated by the potential to use noise for pest management in agriculture (Polajnar et al., 2015). For example, researchers found that playing substrate-borne noise to mask the mating calls of leafhoppers results in reduced mating (Eriksson et al., 2012; Mazzoni et al., 2009; Nieri & Mazzoni, 2018). In addition, Hofstetter et al., (2014) reported reduced reproductive success, movement and survival of pine bark beetles in the presence of substrate-borne noise that spectrally overlapped with beetle signals.

Masking by anthropogenic noise is also likely to impact foraging and predator-prey relationships of invertebrates, as many animals use substrate-borne sound for finding prey or avoiding predators (Castellanos & Barbosa, 2006; Fertin & Casas, 2007). To date, no evidence exists that this is the case and only one study has investigated this dynamic. Wu & Elias (2014) experimentally induced vibratory white noise of different intensities in webs of European garden spiders and did not find a consistent decrease in prey-detection sensitivities. By contrast, studies

investigating substrate-borne noise induced by wind indicate that some invertebrates exploit substrate-borne noise when hunting prey. The salticid spider (*Portia*) and the assassin bug (*Stenolemus bituberus*) are more likely to successfully hunt web-spiders in the presence of wind (Wignall et al., 2011; Wilcox et al., 1996). In these cases, substrate-borne noise is likely to mask vibrations that the predators produce during their approach.

### *Strategies to avoid masking*

Substrate-borne noise may be especially difficult for invertebrates to spatially avoid owing to the fact that it does not attenuate quickly (Kinsler et al., 1999). However, evidence exists that some species may be able to exploit gaps in noise and temporally avoid masking. Wind produces intermittent substrate-borne noise similar to anthropogenic noise sources such as traffic. Researchers found that male treehoppers (which produce substrate-borne courtship signals) preferentially signal during wind-free gaps rather than during experimentally induced wind gusts; females are less likely to respond to male signals during induced wind gusts, and natural treehopper populations are more likely to signal during times of day with less wind (McNett et al., 2010).

Invertebrates produce substrate-borne sound with a diverse array of mechanisms (e.g., percussion, stridulation; see Glossary), and the mechanism an animal uses is likely to affect the extent to which they will be able to shift their signaling frequency in the presence of noise. For example, Bunkley et al. (2017) found differential impacts of noise on the abundances of arthropod families that communicate with substrate-borne sound, potentially owing to the fact that a range of signaling mechanisms is used across families. Terrestrial arthropods use four major mechanisms to produce substrate-borne sound: percussion, stridulation, tymbalation and tremulation (see Glossary). Percussive sound is produced by collisions between parts of the body or between a part of the body and the substrate. These sounds are broadband at the source and the spectral information that arrives at the receiver is solely a result of filtering properties of the substrate (Elias & Mason, 2011). Thus, it is hypothesized that receivers are more likely to use information contained in the timing and/or amplitude of percussive signals than frequency (Elias & Mason, 2011). In the presence of substrate-borne noise, invertebrates that communicate with percussion are unlikely to be able to use frequency-shift mechanisms, such as narrowing the spectrum of signals, to overcome masking.

By contrast, tremulation mechanisms use simple muscular movements of the body and appendages to produce narrowband signals. The ability to shift the spectral properties of tremulation signals depends solely on muscle properties, suggesting that invertebrates that use tremulation will be able to modify the spectral content of their signals in the presence of noise. For example, female southern green stink bugs change the frequency of their tremulations in the presence of frequency-overlapping noise (Polajnar & Čokl, 2008). As mentioned above, invertebrates can also produce substrate-borne signals with stridulation and tymbalation. These mechanisms require specialized exoskeletal structures; when the individual produces vibrations, these structures concentrate acoustic energy to specific bandwidths. In this way, animals can maximize their signal efficacy in particular substrates. The ability of invertebrates to shift the frequency characteristics of signals when using stridulation and tymbalation is likely constrained by muscular physiology (how fast can muscles twitch) and the particular details of their sclerotized sound-producing structures. We do not yet know the extent to which these animals are able to shift the frequencies of their signals in the presence of noise. However, for all invertebrates that communicate with substrate-borne sound, spectral or temporal components of

signals are often species- and plant-host specific (Cocroft et al., 2006; Cokl et al., 2005; McNett & Cocroft, 2008). Thus, even if senders can adjust these components to avoid masking by noise, receivers may not respond to the new signals.

### 3.5 FUTURE DIRECTIONS

Across the different communication modalities discussed in this Review, some common research needs have emerged that will need to be addressed to allow us to understand the potential impacts of anthropogenic noise on invertebrate communication. First, it is vital that noise from anthropogenic sources is adequately quantified. In the past, recording equipment developed to record sound that is audible to humans has been sufficient for recording sound relevant to most vertebrates. In order to record sound that is relevant to a broader taxonomic range, many have argued for recording far-field sound without the use of recording filters designed for human hearing (Francis & Barber, 2013a; Morley et al., 2014; Shannon et al., 2016). We echo that suggestion here. In addition, there is presently no straightforward method for measuring near-field anthropogenic noise. Although technologies such as PIV and hot wire anemometers have potential applications in this context (Bomphrey et al., 2005; Sane & Jacobson, 2006), PIV is expensive and has not been easily adapted for field applications, and hot wire anemometers do not have fine enough resolution to record particle motion relevant to invertebrates. Similarly, the recording of substrate-borne sound currently requires expensive laser vibrometer technology. This technology has only been employed once for measuring noise from anthropogenic activities (C. Wu & Elias, 2014), and it is not accessible to many researchers. Accelerometers are potentially low-cost alternatives to measure substrate-borne noise, although they are unsuitable for recording sound through many substrates given their weight (Cocroft et al., 2014). All in all, recording and quantifying near-field and substrate-borne anthropogenic noise is a major challenge, but with increased attention and focused effort, existing technologies could be adapted to fit this purpose.

Second, future research should seek to understand the variety of impacts that anthropogenic noise in each modality has on a diversity of invertebrate species. Studies should make sure to report the noise source, characteristics of the noise, the signaling environment (e.g., substrate type), signaling mechanisms and type of acoustic receivers of the animals of interest. Additionally, research should establish patterns of constraints and adaptations to noise in each modality and should address whether adaptations result in differential reproductive success.

Finally, it is pivotal that research seeks to understand how noise affects invertebrates under natural field conditions, particularly for invertebrates that use near-field and substrate-borne modalities. While laboratory studies are important, especially for isolating and manipulating sounds, it is difficult to extract meaningful information about the ecological implications of noise from their results. So far, the best studies in this regard combine data from field and laboratory settings (Lampe et al., 2012, 2014), or take advantage of heterogeneous noise in the landscape (Bunkley et al., 2017). For example, Bunkley et al. (2017) compared arthropod communities at sites developed for gas extraction that had gas compressors (noise) with those that did not have compressors (no noise). Results from these types of studies are critically important for understanding the impact of anthropogenic noise on invertebrates.

### 3.6 CONCLUSIONS

Anthropogenic noise is an issue of critical environmental concern, predicted to become an even greater problem with increasing population growth and land-use change. At the same time, invertebrates are experiencing major declines across ecosystems (Hallmann et al., 2017; Potts et al., 2010). The evidence presented in this Review, although limited, suggests that anthropogenic noise is likely to impact invertebrate communication in significant ways. Over the past decades, researchers have conducted important work to understand how anthropogenic noise affects vertebrate species (for excellent reviews, see Barber et al., 2011a; Brumm & Slabbekoorn, 2005; Francis & Barber, 2013; Ortega, 2012; Patricelli & Blickley, 2006; Radford et al., 2014; Shannon et al., 2016; Slabbekoorn & Ripmeester, 2008), but investigation into the ways it affects invertebrate life is incomplete. Vertebrates and invertebrates differ in substantive ways, including in the mechanisms they use for producing and receiving sound; critically, this informs how they can respond to noise and the constraints in doing so.

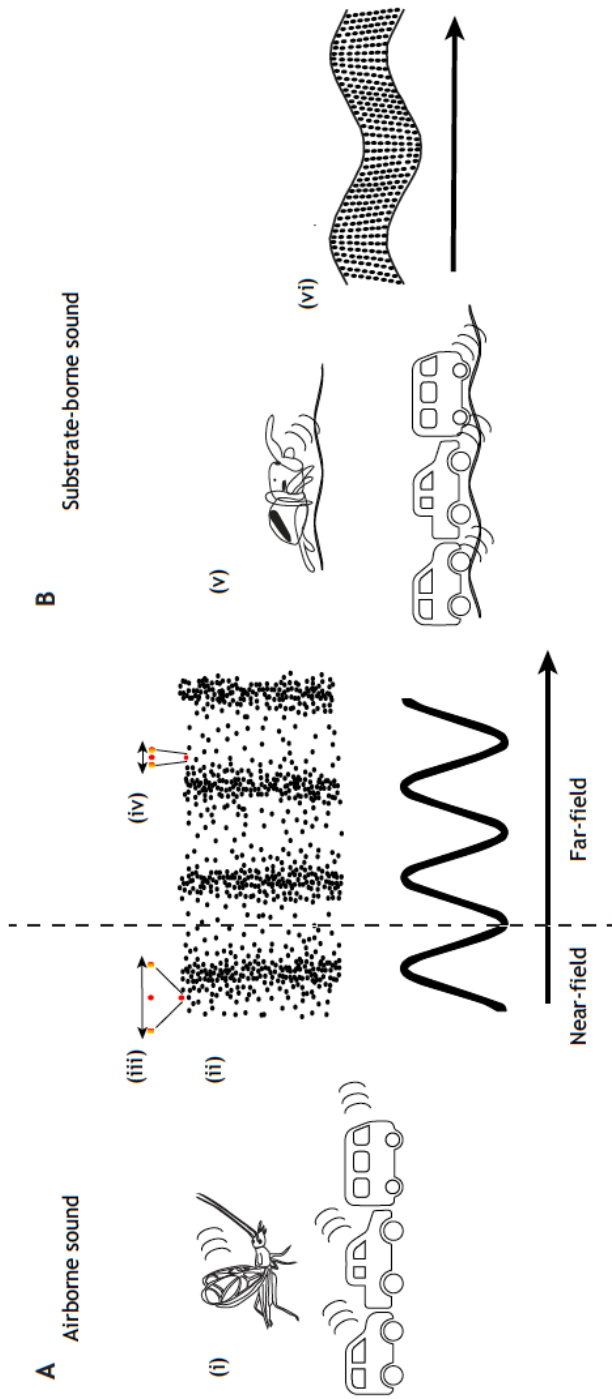
Throughout this Review, we have discussed the sound-producing mechanisms used by invertebrates and their respective vulnerabilities to anthropogenic noise for three types of sound – far-field, near-field and substrate-borne. For the most part, we have focused on how anthropogenic noise might affect the individual interactions of animals, but these dynamics are also important to understand for their bearing on communities and ecosystems (Francis et al., 2012). The majority of species on Earth are invertebrates; they are critical parts of ecosystems (Mulder et al., 1999; Yang & Gratton, 2014) and are food for many species (Morse, 1971). They also provide ecosystem services important for human life, such as pollination, nutrient cycling and waste removal (Losey & Vaughan, 2006; Noriega et al., 2018). The ways in which invertebrates are affected by and adapt to anthropogenic noise could have great implications for ecosystems and ecosystem services. We argue for the expansion of the focus of anthropogenic noise to include noise in near-field and substrate-borne modalities. Expanding our definition of anthropogenic noise and our focus of research will create a more holistic understanding of the potential reach of anthropogenic noise as a pollutant and will potentially lead to effective and efficient mitigation strategies.

### *Acknowledgements*






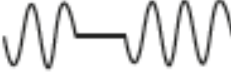

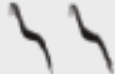



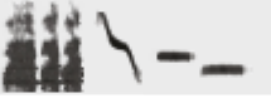


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3.7 FIGURES



**Figure 3.1.** Acoustic properties of airborne and substrate-borne sound. (A) (i) Both biotic and anthropogenic sources can produce airborne sound. (ii) Airborne sound waves are produced by air particle movements (movement of one air particle highlighted in red with double-headed arrows denoting amplitude of movement), and the resultant changes in air pressure (air pressure amplitude denoted by spacing of air particles shown in black) based on those particle movements are shown. Airborne sound waves can be categorized as either near-field or far-field depending on the distance from the sound source and the relative differences between particle motion or air pressure amplitude. Arrow denotes the direction of a propagating airborne sound wave (iii). In the near-field, air particle movement dominates relative to air pressure differences. Near-field sound occurs within one wavelength of the sound source. (iv) In the far-field, air pressure differences dominate over air particle movements. Far-field sound occurs at a distance of greater than one wavelength from the sound source. (B) (v) Both biotic and anthropogenic sources can produce substrate-borne sound. (vi) Substrate-borne waves are complex and can occur within or on the surface of a solid substrate (solid lines denote outline of solid surface). Arrow denotes the direction of a propagating substrate-borne sound wave.

Signal characteristic	Potential for conveying information in noise	
	Weak	Strong
Amplitude	Quiet 	Loud 
Bandwidth	Broadband 	Narrowband 
Temporal structure	Continuous 	Intermittent 
Modulation	Constant 	Modulated 
Repetition	No repetition 	Repetition 
Dimensionality	Unidimensional 	Multidimensional 
Multimodality	Unimodal 	Multimodal 

**Figure 3.2.** Signal characteristics and their hypothesized robustness to masking noise. Column 1: signal information can be conveyed across multiple acoustic characteristics. Column 2: certain variations of each characteristic are predicted to be better at conveying information in the presence of masking noise than others (right to left, respectively). Information can be conveyed using temporal, amplitudinal (white box– wave form) and spectral (grey box– spectrogram) properties.

### 3.8 TABLES AND GLOSSARY

**Table 3.1.** Glossary

**Aerodynamic sound:** Sound production mechanism where the flow of fluid over a structure excites resonance properties of the structure.

**Airborne far-field sound:** Vibration propagating in air at a distance where sound pressure dominates, and sound particle velocity and sound pressure are in phase.

**Airborne near-field sound:** Vibration propagating in air at a distance where air particle velocity dominates and sound particle velocity and sound pressure are not in phase.

**Cue:** Act or structure produced by a sender that has information but has not evolved to elicit a response in a receiver.

**Fluid compression:** Sound-production mechanism where animals produce rapid changes in the local pressure of the medium (e.g. cavitation, ‘sonic boom’).

**Johnston’s organ:** Sensory organ found in the antennae of insects that can detect vibrations in the air.

**Masking:** Situation where a signal/cue co-occurs with noise thereby increasing the threshold for detection by the receiver.

**Noise:** Mechanical waves uncorrelated with any acoustic feature of interest to a receiver. May be anthropogenic (e.g. traffic, industry) or environmental (e.g. heterospecifics, rain, water, wind).

**Percussion:** Sound-production mechanism whereby animals produce vibrations using transient impacts of an appendage against another appendage or against the substrate.

**Signal:** Act or structure produced by a sender that has evolved to elicit a response in a receiver.

**Spatial release from masking:** A phenomenon whereby a signal/cue is more easily detected when spatially separated from noise.

**Substrate-borne sound:** Vibration propagating in a solid.

**Stridulation:** Sound-production mechanism whereby animals produce vibrations using two rigid structures that are rubbed against each other. At least one of the structures (the file) is ridged.

**Sympathetic vibrations:** Phenomenon whereby an airborne sound causes vibrations in a solid that was previously not moving.

**Tremulation:** Sound-production mechanism whereby animals produce vibrations using oscillations of a body part.

**Tymbalation:** Sound-production mechanism whereby animals produce vibrations using a tymbal, a corrugated structure on the exoskeleton.

**Vibration:** Mechanical waves propagating in a solid or fluid medium.

Consequence	Invertebrate	Sound modality	Source
Distraction	Caribbean hermit crab ( <i>Coenobita clypeatus</i> )	Airborne far-field	Chan et al., 2010; Walsh et al., 2017
Processing errors	Earthworm	Substrate-borne	Darwin, 1892; Mitra et al., 2009
Increased stress	Monarch butterfly larvae ( <i>Danaus plexippus</i> )	Unclear, likely substrate-borne	Davis et al., 2018
Developmental changes	Indian meal moth larvae ( <i>Plodia interpunctella</i> ) Field cricket ( <i>Teleogryllus oceanicus</i> )	Airborne far-field	Huang et al., 2003; Gurule-Small and Tingnitella, 2018; 2019
Decreased life span	Field cricket ( <i>Teleogryllus oceanicus</i> )	Airborne far-field	Gurule-Small and Tingnitella, 2019

**Table 3.2.** Summary of literature examining the consequences of anthropogenic noise beyond masking for invertebrates

## CHAPTER 4 BUILT TO LAST A DAY: THE FUNCTION AND BENEFITS OF SPIDER MOUND NESTS

Maggie Raboin and Damian O. Elias

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### 4.1 ABSTRACT

Nests are crucial to the survival of offspring and reproductive success of the animals that build them. These benefits are subject to change over time due to fluctuating conditions inside and outside of nests. For many species, nests are assumed to benefit offspring until they disperse and therefore, nest destruction prior to offspring dispersal results in reduced reproductive success for parents. However, the consequences of nest destruction to reproductive success, or lack thereof, remain largely unstudied across diverse taxa. Here, we experimentally investigate the function and benefits of nests of a mound-building spider. Mason spiders (*Castianeira* sp.) are wandering spiders that build intricate nests (mounds) on top of their egg sacs. Their offspring inhabit egg sacs at nest-sites for up to 7 months, including through winter. We find that despite requiring hundreds of collecting trips and many hours to construct, mason spider nests remain for a small portion of time that offspring occupy nest-sites. Our study finds that nest benefits change over time, likely explaining this dynamic. Nests greatly reduce the rate of predation and parasitism of egg sacs by 19.1% and offspring mortality within egg sacs due to abiotic factors by 19.9%. These effects are only present the few days following nest construction. Our study illuminates the idea that nest destruction does not always result in reduced reproductive success for nest builders. We suggest that nest durability, the ability of a nest to withstand environmental conditions, may be subject to natural selection and a critical, yet understudied, aspect of parental care.

## 4.2 INTRODUCTION

For many animals, early life stages are at higher risk of mortality from sources like predation, disease, and environmental fluctuations than other life stages (Clutton-Brock, 2001). To increase their reproductive success, some parents invest time and energy into protecting offspring by building protective structures. Benefits of protective structures like nests have been found across species. For example, nests increase reproductive success for many mammals (Reichman & Smith, 1990), birds (Mainwaring et al., 2014), reptiles (Angilletta et al., 2009), fish (Ishimatsu et al., 2007; Takegaki & Nakazono, 2000), and arthropods (Hieber, 1992b; Wyatt, 1986). However, the benefits of nests are unlikely to be static through time. From the day they are constructed, conditions outside of the nest like environmental conditions and predator dynamics, are constantly fluctuating and threatening to damage the nest and/or harm offspring.

The dynamic nature of nest benefits through time has great implications for how much parents invest in building and/or maintaining nests. Across species, this varies greatly. For example, some nests are actively maintained for years like those belonging to social animals that use nests for generations, including some termite mounds that are occupied for centuries (Erens et al., 2015). In contrast, many songbirds build nests that are made to last one breeding season (Mainwaring et al., 2014), and still others, like some amphibians (Giaretta & Menin, 2004) and army ants (Schneirla et al., 1954), construct nests that only need to remain intact for hours to days.

Outside nests, the intensity, frequency, and predictability of environmental conditions may impact the length of time that nests benefit offspring. For example, rainfall has been implicated in nest destruction of vertebrate and invertebrate nests, sometimes zeroing the benefits of nests and leading to their abandonment by parents (Thompson & Furness, 1991; Rojas *et al.*, 2019). In other well-known cases, the presence of predators induces accelerated hatching or fledging of offspring. In these situations, the benefits of nests for aggregated, stationary life stages rapidly decrease as the risk of predation increases (Martin et al., 2018; Warkentin, 1995).

Inside nests, life history traits of species, like how much parental care early life stages require and the length of time that care is necessary, impact the rate that nest benefits decrease over time. For example, scientists have long recognized that the young of most species can be placed along an altricial to precocial spectrum (Augustine et al., 2019). Nests of altricial animals (species where young are born at earlier stages of development) need to stay in nests longer and the benefits of nests decay slowly, whereas many precocial animals do not need nests at all (Starck & Ricklefs, 1998). Over time and as offspring develop, the benefits of nests decrease until they are no longer necessary.

For most species, the length of time that nests benefit offspring and the amount of time that offspring occupy nest-sites are assumed to be inextricably linked. In fact, examples of nest-destruction prior to offspring dispersal are almost always reported to result in reduced reproductive success for nest-builders (Thompson & Furness, 1991; Rojas *et al.*, 2019). However, the consequences of nest destruction to reproductive success, or lack thereof, remains largely unstudied across diverse taxa. We investigate the consequences of nest destruction on reproductive success by untangling the length of time that offspring occupy nest-sites, the benefits of nest presence, and how those benefits change over time.

Our study focuses on the mason spider, *Castianeira* sp., wandering spiders of the northern Rocky Mountains, USA that cover their eggs with elaborate mound nests, protective structures

that include both an egg sac and elaborate mound. While most spiders protect their eggs and young with protective structures that minimally involve investment in silk for egg sacs or webs (Hieber, 1992; Toyama, 1999), mason spiders construct mounds made of hundreds of pebbles, leaves and sticks, held together with silk. Our preliminary observations indicated that despite a large investment in parental care and the fact that offspring persisted at nest sites for many months, mason spider mounds are regularly destroyed. Why do mason spiders go through such effort to build a mound nest that only remains intact for a small portion of the time that offspring inhabit nest sites? We sought to answer this question by describing the details of parental care and life history of a mound-building spider for the first time. We then experimentally investigated the function of mason spider mound nests and how the benefits of mounds change over time. Finally, we discuss why nest duration is an important, yet understudied aspect of parental care.

#### 4.3 MATERIALS AND METHODS

##### *Study Site and Data Collection*

The mason spider studied here is an undescribed species in the *descripta* group of the genus *Castianeira* (Corrinidae) (Reiskind, 1969). The necessary taxonomic work for species determination of mason spiders is outside the scope of this study. Mason spider specimens have been deposited in UC Berkeley's Essig Museum (EMEC 1199520 – EMEC 1199527).

We studied mason spiders along the Snake River in Bridger-Teton National Forest, WY, USA, unceded lands of the Cheyenne, Eastern Shoshone, and Shoshone-Bannock nations (43°18'10.3"N 110°46'27.6"W; altitude 1800 m) in open habitats dominated by sage, alfalfa, and various grasses. These habitats have gravel soil with large rocks scattered throughout. Mason spiders lay egg sacs in the crevices of rocks and build their mounds on top (see Figure 4.1A, B; *Behavioral description*). We observed 1,368 mounds (June – September 2015-2018 and in the winter of 2017/18) and filmed the construction of 14 mounds using GoPro (HERO 4) cameras across 4 mound-building seasons. Each day, we deployed 3-4 GoPro cameras that were mounted on tripods and positioned one meter from nest sites so as not to disrupt mason spiders during mound construction. We describe mason spider mound-building behavior using observational data collected at field sites in combination with videos of mound construction. We analyzed the videos of mound construction using BORIS (Behavioral Observation Research Interactive Software) (Friard & Gamba, 2016) and ImageJ (MtrackJ) software (Schneider et al., 2012) to collect data on the number of collecting trips per hour and length of collecting trips.

##### *Mound removal experiment*

We experimentally tested the impact of mason spider mounds on offspring survival over time. In 2017, we identified 188 egg sacs under construction and randomly assigned them to one of four treatment groups. For each treatment, we removed mounds (1) immediately following their completion (N=50), (2) 24 hours post-construction (N=51), and (3) 10 days post-construction (N=42). Control mounds (4) were never removed (N=45). All egg sacs remained in place until they were collected (range 24-109 days). Mound construction was determined to be completed when mason spiders did not return to the mound for 1 hour. We put a small dab of super glue on top of all mounds to ensure that they remained in place for duration of the treatment. The super glue was superficial; structurally reinforcing a few pieces of the mound and



did not touch the egg sac. Treatment time intervals were chosen haphazardly based on field observations.

Egg sacs were randomly chosen to be collected in August (N=84) or October (N=104) of 2017. Egg sacs collected in August remained in place for 24 - 45 days ( $31.23 \pm 5.08$ , mean  $\pm$  SD) and egg sacs collected in October remained in place for 84 - 109 days ( $94.44 \pm 5.00$ ) before they were collected. Once collected, egg sacs were transported to UC Berkeley and dissected to determine offspring mortality in each egg sac (presented as a proportion - number of dead spiderlings and/or eggs divided by total offspring in egg sac), life stage (egg or spiderling), presence of mold or desiccation, and evidence of parasitism or predation (Figure 4.1C-E). Eggs were determined to be dead if they were desiccated or covered in mold. Egg sacs were determined to be parasitized if wasps or wasp larva were found in egg sacs and predated if egg sac had a large hole and was empty. Dissections were done blind to any information about treatment.

### *Statistical analyses*

All statistical tests were performed in R v.3.6.2 (R Core Team, 2020). We performed generalized linear models to assess the effect of mounds on mortality in mason spider egg sacs overtime. Egg sacs that were predated/parasitized resulted in 100% mortality of eggs. Therefore, we analyzed mortality due to predation or parasitism (biotic mortality) and desiccation or mold (abiotic mortality) separately. We first assessed the effect of mounds on predation/parasitism of mason spider egg sacs over time by using a generalized linear model with a binomial distribution and assigned presence/absence of biotic mortality as our response variable and treatment as our predictor variable.

Next, we assessed the effect of mounds on abiotic mortality in mason spider egg sacs over time. In our model, we assigned dead/live offspring as our response variable with the *cbind* function in R and treatment as our predictor variable. We accounted for overdispersion in our data by using a quasibinomial distribution.

## 4.4 RESULTS

### *Behavioral description: Mound-building*

Female mason spiders lay eggs and construct mound nests at field sites mid-July through the beginning of September (Figure 4.2A). To begin, a female constructs an egg sac in a small indentation in a rock by laying silk flush with the rock. The female then lays 8 - 36 orange eggs ( $18 \pm 5.1$ , mean  $\pm$  SD, N=151) and covers them with another layer of silk (Figure 4.1A). This layer of silk is tough and papery, with an opaque pearl color, much like those described in other *Castianeira* species (Montgomery, 1909; Reiskind, 1969). Often, this is the final layer of silk in egg sac construction, however some females will construct a third, thin layer of silk with an empty space (~2-4 mm) between the second and third layers of silk.

Following completion of the egg sac, the female constructs a mound (Figure 4.2B). To build mounds, she gathers items (e.g. pebbles, dried leaves, seeds, small sticks, arthropod parts, bird feathers) in individual collecting trips traveling between 1.25 - 80.66 cm ( $15.34 \pm 11.75$ , N=462 observed in 14 females), and returning to assemble the items into a mound held together by silk (Figure 4.1A,B). Shorter collecting trips were often due to females stealing material from nearby mounds rather than collecting their own. The number of collecting trips conducted by individual females to construct a mound varies. Females conduct 36 - 174 collecting trips per

hour ( $102.43 \pm 44.31$ ,  $N=14$ ) until the mound is finished, which can take anywhere from 6 to 13 hours (Figure 4.2B). Based on this, female mason spiders likely conduct ~ 500 - 2,000 collecting trips to complete a mound. During all observations, we never witnessed a female spider returning to her mound following its completion. Eggs remain in the egg sac for ~ 2 - 4 weeks before they hatch into spiderlings and overwinter as 1<sup>st</sup> or 2<sup>nd</sup> instar spiderlings in the egg sac for ~ 7 months. In May, mason spiders hatch out of egg sacs as 2<sup>nd</sup> instar spiderlings, leaving a molt behind in the egg sac (Figure 4.2A).

In our observations, mounds were regularly destroyed likely by weather events. Egg sacs visited one month or more following their completion were never observed with a mound. However, egg sacs remained in place through winter until spiderlings dispersed in spring (Figure 4.2A).

#### *Mound removal experiment*

Mason spider egg sacs were most often predated on by field crickets (*Gryllus sp.*) and parasitized by parasitoid wasps (*Gelis spp.*). We found that mounds significantly decreased predation/parasitism. Overall, 20.2% of egg sacs ( $N = 188$ ) were predated/parasitized and these rates differed across the four treatments. Egg sacs whose mounds were removed immediately following completion and 1 day later were the most heavily predated and parasitized (28% ( $N=50$ ) and 29.4% ( $N=51$ ), respectively) as compared to egg sacs whose mounds were removed at 10 days and >24 days (11.9% ( $N=42$ ) and 8.9% ( $N=45$ ), respectively). We found significant differences between predation/parasitism of egg sacs with mounds removed immediately and those that were never removed (control) (Table 4.1, Figure 4.3). In addition, we found that the difference between egg sacs where mounds were removed immediately and at 10 days approached significance (Table 4.1).

We also found that mounds significantly decrease mortality due to abiotic factors. The proportion of offspring mortality in egg sacs showed a negative relationship with the number of days mounds were present (Table 4.2, Figure 4.4). The proportion of offspring mortality in egg sacs where mounds were removed at 1 day ( $0.12 \pm 0.23$  mean  $\pm$  SD:) and never removed (control) ( $0.11 \pm 0.25$ ) were significantly lower than mortality in egg sacs where mounds were removed immediately,  $0.30 \pm 0.35$  (Table 4.2, Figure 4.4).

## 4.5 DISCUSSION

Mason spiders construct elaborate mound nests by building mounds on top of their egg sacs, a parental care behavior that involves hundreds of collecting trips over many hours. We found that mounds are instrumental in reducing offspring mortality only for a short period of time following nest construction. This effect decreases throughout the first 10 days following construction, at which point egg sacs with and without mounds experience the same amount of offspring mortality. Additionally, we found that mounds protect offspring from mortality due to abiotic and biotic factors and that these benefits change differently over time.

Despite requiring a large investment of energy, we observed that mounds only remain on egg sacs for a small portion of the time that offspring inhabit nest-sites. This finding contrasts with other animals that similarly invest in building protective structures but whose nests are durable enough to protect offspring throughout development and/or until they leave the nest-site (Mainwaring *et al.*, 2014). Our study suggests that mason spider mound destruction after a short

critical period (~ 10 days) does not result in reduced fitness for mason spiders due to the rapidly decreasing benefits of mounds over time.

#### *Benefits of mounds against abiotic factors*

The benefits of mason spider mounds are large early on, reducing the rate of predation/parasitism by 19.1% and abiotic mortality within egg sacs by 19.9%. Our experiment suggests the existence of a critical period early in offspring development where mason spider eggs are not able to survive without a mound. Across species, early development is often the period when offspring are most vulnerable to a variety of factors, including fluctuating environmental conditions, parasites and predation (Clutton-Brock, 2001). We found that for the most part, mounds protect offspring from mortality due to abiotic factors in the first 24 hours following construction. Abiotic mortality in mason spider egg sacs may be due to a variety of factors including fluctuating temperatures, humidity, or exposure to UV radiation (UVR) when mounds are not present. In other spider species, eggs have been found to be especially sensitive to temperature and humidity (Hieber, 1992b; Li & Jackson, 1996; Pike et al., 2012). Mounds may provide insulation from fluctuating temperatures or humidity, which can be dramatic in alpine habitats during summer months. Additionally, mason spider eggs are laid on the tops of rocks in exposed alpine habitats with direct UVR exposure. UVR exposure can cause mortality and sub-lethal damage in the early life stages of arthropods and amphibians (Blaustein & Belden, 2003; Bothwel et al., 1994). Mason spider mounds may reduce UVR exposure of eggs early in development and could be one driver of mound-building behavior.

Interestingly, mounds do not protect mason spider offspring during arguably the most dramatic environmental conditions that they encounter, winter. In the eight months that follow mound construction, mason spider offspring remain inside egg sacs at nest-sites as spiderlings and without mounds. They experience dramatic fluctuations in environmental conditions, including rain, snow, freezing temperatures, and snow melt. For arthropods, cold tolerance has been described as a critical component of their biology (Brandt et al., 2020; R. E. Lee, 2010). Future research should seek to understand aspects influencing mason spiderling overwintering, including their cold tolerance strategy, the insulating role of silk and egg sac construction, the conditions that cue diapause and the resumption of development in the spring, and the metabolic characteristics necessary for overwintering.

#### *Benefits of mounds against biotic factors*

In addition to their abiotic benefits, mounds also protected mason spider egg sacs from predation and parasitism, however this benefit was only present in the first 10 days following construction. Our observations suggest that egg sacs were most often predated on by crickets, although one predation event by a grouse was captured on film. Crickets have been found to be the primary predators of eggs of other arachnid species and may drive the evolution of active defense strategies of a cave-dwelling species of harvestmen (García-Hernández & Machado, 2017). The parasitoids of mason spider egg sacs were wasps in the genus *Gelis* — ichneumonids commonly known to parasitize a variety of silk covered masses including spider egg sacs and cocoons (Harvey, 2008). One hypothesis as to the mechanisms of mound protection is that mounds camouflage egg sacs from predators. To human eyes, mason spider mounds are visually conspicuous. However, *Gelis* parasitoid wasps and crickets likely detect their hosts/prey via olfactory cues (Matsumoto & Mizunami, 2000; Van Baarlen et al., 1996). In the initial days following construction, mounds might camouflage the scent of silk or eggs rendering them

undetectable or increase the amount of time parasitoid wasps and other predators need to find eggs. Similarly, some birds are known to include green plant material in their nest construction, and it is thought that compounds in these materials serve as olfactory camouflage (Mainwaring et al., 2014).

### *Costs of mound-building*

Although this study did not explicitly evaluate costs, our observations of mound-building point to a variety of costs that should be assessed in the future. First, building a mound is likely an energetically expensive activity. During mound-building, mason spiders are regularly observed stealing material from nearby mounds rather than collecting their own. Similar behaviors have been witnessed among a variety of bird species and used as indicators that nest-building is energetically costly (Mainwaring & Hartley, 2013). Additionally, direct measurements indicate that birds use vast amounts of energy when transporting nesting material (Mainwaring & Hartley, 2013). Second, increased predation risk is a common cost of parental care and likely applies in this case (Ghalambor & Martin, 2002; Magnhagen, 1992; Reguera & Gomendio, 1999). Mason spiders build mounds and conduct collecting trips in mid-day in exposed areas making them and their nest location conspicuous to visual predators. One hypothesis as to the reason that mason spiders lay eggs at night is that by performing the most vulnerable and exposed parental behaviors in the dark, they reduce their risk of predation by visually oriented predators. Finally, energy spent on the current reproductive event, laying eggs and building mounds, reduces the time and energy that could be spent on future reproductive events (Alonso-Alvarez & Velando, 2012).

### *Nest Durability*

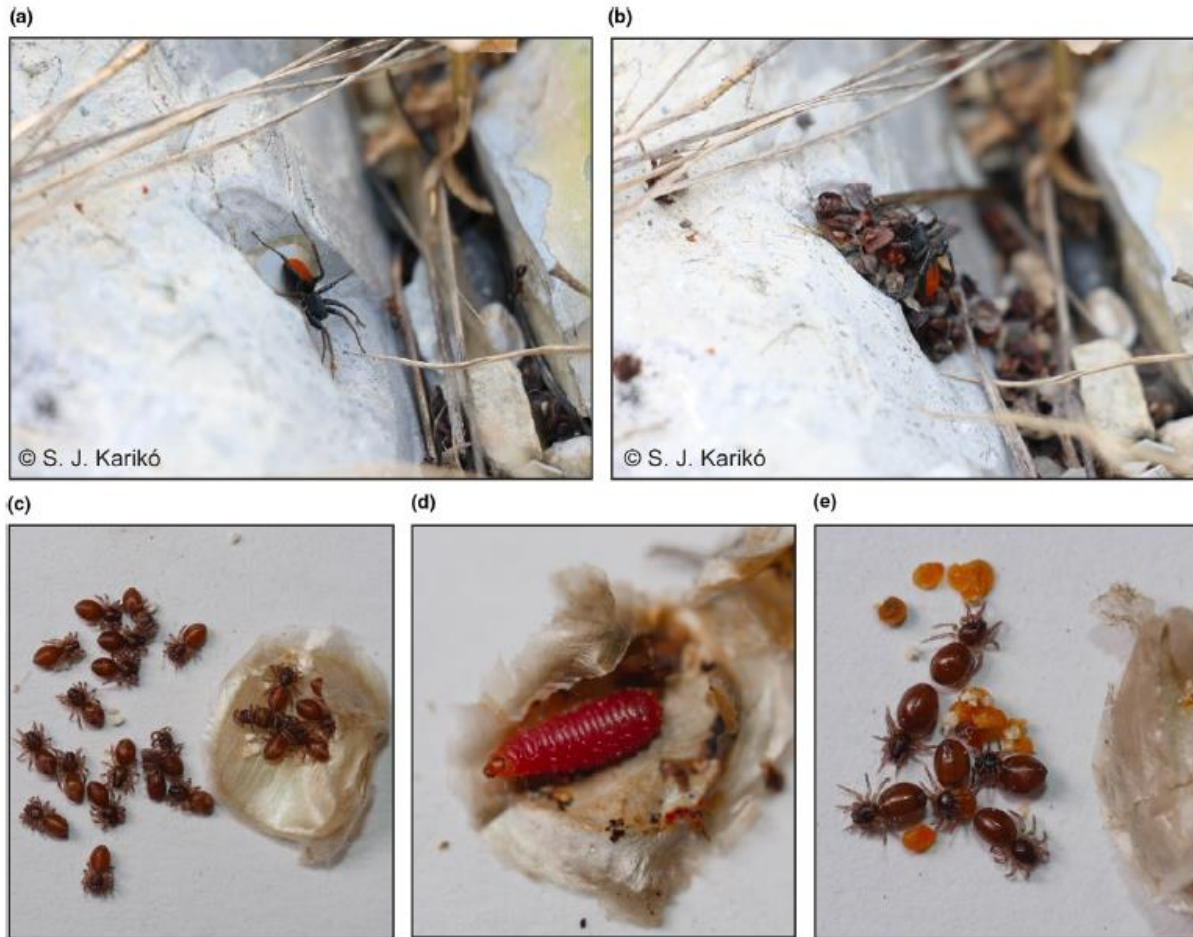
Our study points to multiple proximate reasons as to why mound destruction does not result in reduced reproductive success for mound builders. However, our findings also hint toward ultimate reasons for this dynamic that should be investigated further. Previous research suggests that nests that last longer are more costly to build and maintain (Abé et al., 2017). Therefore, mason spider nest durability, defined as the amount of time that a nest is able to withstand environmental conditions, may be a trait adapted to predictable environmental conditions that cause nest destruction. Similar traits of extended phenotypes, like the materials of bird bowers constructed for signaling or the shape of spider webs constructed for prey capture, have been shown to evolve or plastically respond to changing environments (Blamires, 2010; Schaedelin & Taborsky, 2009). We suggest that nest durability is also subject to selection. Nest durability may be driven by (1) the ability of young to survive independent of nests, (2) the frequency, intensity, and predictability of environmental conditions, (3) nest-site selection, and (4) nest shape and structure, including the costs and benefits of its construction and maintenance. Future work should examine these hypotheses in mason spiders but also across taxa in order to understand the different factors influencing the evolution of parental care strategies and animal architecture.

### *Acknowledgements*

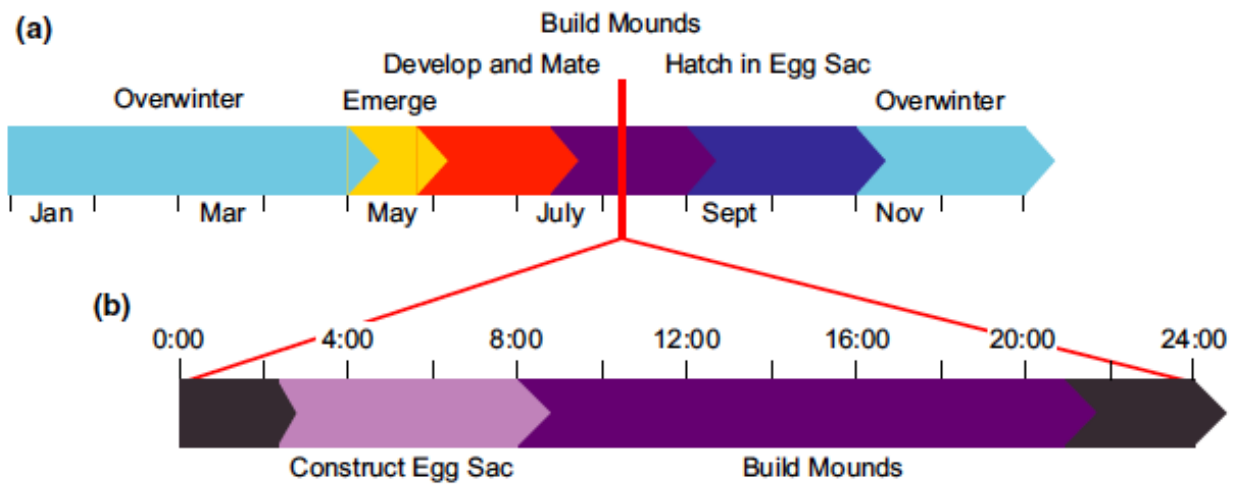
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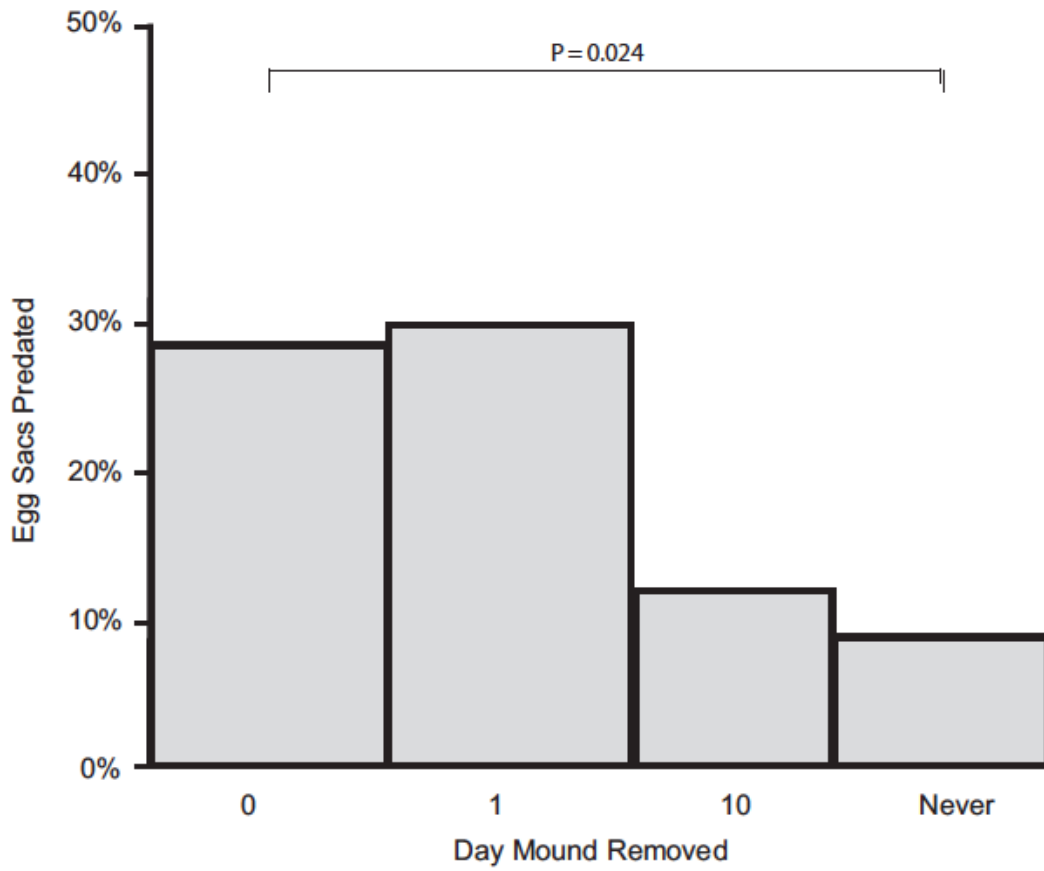
## 4.6 FIGURES



**Figure 4.1.** Mason spider constructing an egg sac and mound in the field and examples of egg sac contents. Mason spider constructing (a) egg sac and (b) mound at a single nest site. (c) Egg sac with healthy spiderlings. (d) Egg sac with wasp larva and total mortality of mason spider offspring. (d) Egg sac with healthy spiderlings and desiccated eggs.

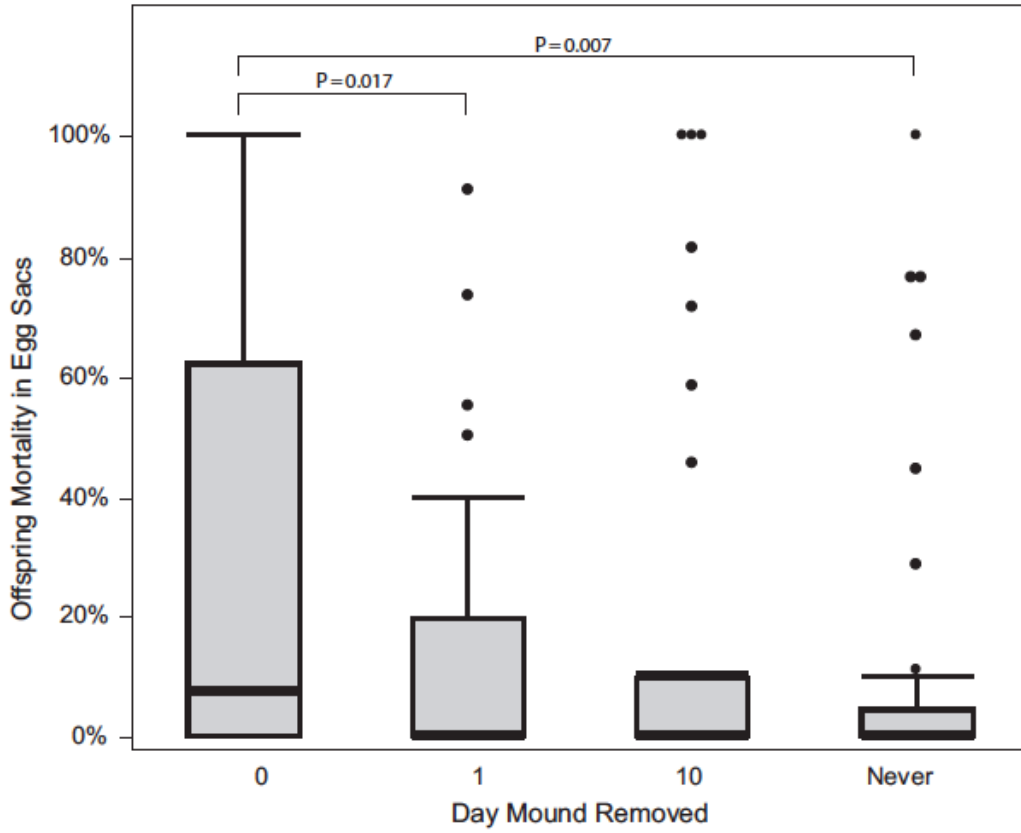


**Figure 4.2.** Timelines of mason spider life history and mound-building. (a) Mason spider stages of development throughout one year. Mason spiders lay eggs and build mounds in July and August. Eggs hatch into spiderlings and remain in egg sac throughout winter. In May, spiderlings emerge from egg sacs and develop into adults in early July. (b) Timeline of the building of one mason spider mound over 24 hr. Information included in timelines is generalizations based on hundreds of observations of individual spiders and egg sacs. However, variation exists in and around each stage of development.



**Figure 4.3.** Predation/parasitism rate of egg sacs when mounds were removed from egg sacs at different time intervals following construction.





**Figure 4.4.** Box plot representing the percentage of abiotic mortality in egg sacs when mounds were removed at different time intervals following construction.

#### 4.7 TABLES

Parameter	Estimate	Std. Error	<i>t</i>	<i>p</i>
Intercept	-0.944	0.315	-2.999	<b>0.003</b>
Day 1	0.069	0.440	0.157	0.875
Day 10	-1.057	0.571	-1.851	0.064
Day >24-100	-1.383	0.611	-2.263	<b>0.024</b>

**Table 4.1.** Summary of results for generalized linear model assessing the effects of mound removal over time on the presence of predation/parasitism in egg sacs. Bolded values indicate statistical significance.

Parameter	Estimate	Std. Error	<i>t</i>	<i>p</i>
Intercept	-0.770	0.305	-2.520	<b>0.013</b>
Day 1	-1.264	0.524	-2.411	<b>0.017</b>
Day 10	-0.769	0.467	-1.646	0.102
Day >24-100	-1.384	0.505	-2.739	<b>0.007</b>

**Table 4.2.** Summary of results for generalized linear model assessing the effects of mound removal over time on the proportion of abiotic mortality in mason spider egg sacs. Bolded values indicate statistical significance.

## CHAPTER 5 SUBSTRATE-BORNE NOISE DISTRACTS FROM PARENTAL CARE

Maggie Raboin and Damian O. Elias

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### 5.1 ABSTRACT

Caring for young often requires sustained attention that when disrupted can have consequences for offspring survival. Through distraction, extraneous stimuli occupy the limited attention of animals and divert their attention away from important tasks, like parental care. One potential distracting stimulus, anthropogenic noise, has become widespread across ecosystems. Anthropogenic substrate-borne noise, although often overlooked, is thought to be as prevalent as airborne noise and could greatly impact the many animals sensitive to substrate-borne sounds. Here, we record anthropogenic substrate-borne noise at sites with differing road traffic in the Greater Yellowstone Ecosystem – an area with some of the most protected land in the world. We investigate the potential of substrate-borne noise from roads to alter parental care behavior and offspring survival of mason spiders (*Castianeira* sp.), a species that builds elaborate nests to protect offspring from predators and fluctuating environmental conditions. We found that mason spiders spent significantly less time building nests, made fewer nest-building collecting trips per hour, and were more likely to get lost while navigating to and from their nests in the presence of greater substrate-borne noise. However, we did not find an effect of noise on offspring survival. These results suggest that substrate-borne noise distracts mason spiders from providing parental care. Our study provides field-based empirical evidence that substrate-borne noise from an anthropogenic source is present in ecosystems and can impact the parental care of animals.

## 5.2 INTRODUCTION

Many animal parents must perform behaviors that are critical for offspring success but that often require their sustained attention. Attention occurs through neural processes that organize and filter many environmental sensory stimuli to just a few and, in doing so, allows animals to coordinate an appropriate behavioral response to a given stimulus (Bushnell, 1998; Chan et al., 2010). Parental care behaviors like nest-building, brooding, and feeding require attention to the extent that parents will forgo foraging (Sabat, 1994; Weimerskirch, 1995) or investment in future reproduction (Golet et al., 2004; Hanssen et al., 2005) while caring for young. In some cases, extraneous stimuli may occupy the limited attention of animals, thereby involuntarily diverting their attention away from essential tasks – a process called distraction (Chan et al., 2010). Through distraction, stimuli have been found to interrupt behavioral and cognitive processes of animals, like risk assessment (Chan et al., 2010; Morris-Drake et al., 2016), mate assessment (Evans et al., 2006), and food localization and discrimination (Purser & Radford, 2011; Riffell et al., 2014). However, little is known about how distraction can impact parental care behaviors and ultimately, offspring success including survival.

Across environments, animals contend with extraneous stimuli introduced by human activities associated with urbanization, economic development, and transportation networks (Halfwerk & Slabbekoorn, 2015). One such stimulus, anthropogenic noise, is widespread and has been demonstrated to impact animal behavior in a multitude of ways (reviewed in Shannon et al., 2016). For example, anthropogenic noise can alter parental care behaviors like nest maintenance (Picciulin et al., 2010), feeding (Injaian, Taff, & Patricelli, 2018; Nedelec et al., 2017; Ng et al., 2019), and offspring defense (Injaian, Poon, et al., 2018; Nedelec et al., 2017), sometimes impacting offspring growth and survival. Altered behavioral responses are driven by the tendency of anthropogenic noise to mask important sounds, introduce misleading information, or distract animals from processing information appropriately (Dominoni et al., 2020). Through distraction, noise has been shown to affect the cognitive processing of diverse taxa leading to altered perceived predation risk (Chan et al., 2010; Morris-Drake et al., 2016) and decreased foraging efficiency (Halfwerk & Van Oers, 2020; Purser & Radford, 2011).

To date, most research on the impact of anthropogenic noise on animal behavior has focused on airborne and waterborne sound (Hawkins et al., 2021; Raboin & Elias, 2019). However, most terrestrial animals lack the ability to directly detect airborne sound and instead have highly sensitive organs for detecting substrate-borne sound (Cocroft & Rodriguez, 2005; Raboin & Elias, 2019). Spiders, in particular, are well-known for their use of substrate-borne sound in courtship and mating (Elias et al., 2003; Rosenthal & Hebets, 2012), foraging (Klärner & Barth, 1982; Wignall & Taylor, 2011) and avoiding predators (Lohrey et al., 2009; Riechert & Hedrick, 1990). Human activities are thought to produce substrate-borne noise as readily as airborne noise (Lecocq et al., 2020; Riahi & Gerstoft, 2015) and emerging evidence suggests that it has the ability to alter the behavior of animals sensitive to substrate-borne sounds. Substrate-borne noise is known to impact earthworm density (Velilla, Collinson, et al., 2021) and disrupt courtship and mating in stinkbugs, leafhoppers, beetles, and spiders (reviewed in Raboin & Elias, 2019). In addition, Phillips et al. (2020) showed that substrate-borne noise significantly impacted burying beetle brood size, likely through altered parental care behaviors.

In this study we examined mason spiders (Araneae, Corinnidae, *Castianeira* sp.) and the impact of substrate-borne noise on their parental care behavior. The behavioral ecology of mason spiders suggests that anthropogenic substrate-borne noise may impact parental care and the

survival of offspring in keyways. Mason spiders are wandering spiders that often live along roadsides in the Greater Yellowstone Ecosystem - an area with some of the most protected land in the world (Lynch et al., 2008) that draws increasing amounts of visitors and human activity every year (NPS, 2021). Mason spiders provide care for young by constructing mound nests to protect their egg sacs from parasitoid wasps and fluctuating environmental conditions (Raboin & Elias, 2021). Mound-building involves focused attention over at least 12 hours to collect nest material and is instrumental in reducing offspring mortality (Raboin & Elias, 2021). Like the majority of spiders, species in the family Corinnidae lack tympanal ears to detect airborne sound (Haddad, 2012; 2013). Instead, they have well developed lyriform organs suggesting that they rely on substrate-borne vibrations for collecting information about the environment (Haddad, 2012; 2013). If human activity produces substrate-borne noise that mason spiders' sense, then it may occupy their limited attention and alter parental care behaviors critical for offspring survival.

Here, we investigate anthropogenic substrate-borne noise and the parental care of two populations of mason spiders in areas of high and low road traffic in the GYE. We first record substrate-borne noise produced by road traffic and test the hypothesis that areas of high traffic will have higher levels of substrate borne noise than areas of low traffic. Next, we investigate the behavioral responses of mason spiders to anthropogenic noise while providing parental care. We measured substrate-borne noise and parental care behaviors associated with building nests including the amount of time spent building nests, the ability of female mason spiders to navigate while building nests, and the directness of collecting trips to and from nests. Finally, we examine if altered parental care behaviors result in reduced offspring survival by measuring the mortality of mason spider offspring at sites with high and low road traffic. We predict that parents distracted by substrate-borne noise spend less time constructing nests resulting in a reduction of resources allocated to offspring protection, and ultimately, in increased offspring mortality.

### 5.3 MATERIALS AND METHODS

#### *Study sites and substrate-borne noise*

We studied mason spiders at two sites in Bridger-Teton National Forest, WY, USA in open habitats dominated by sage, alfalfa, and various grasses. Sites were chosen because of their abundance of mason spider nests and proximity to roads. Site A was adjacent to a high traffic paved road and Site B was adjacent to a low traffic dirt road. Sites were 13 miles apart and similar in habitat and elevation. At each site, we identified mason spiders building nests within 25 meters of the road and measured the amount of substrate-borne noise with a portable laser Doppler vibrometer (Polytec PDV 100) connected to a digital recorder (Olympus VN-721PC). To measure substrate-borne noise, we focused the laser on a piece of reflective tape adhered to nylon that was stretched across a metal ring and placed on the ground 25 meters from each road. We made two recordings at each site that were five minutes in length. We analyzed recordings for vehicles passing per minute using Audacity 3.0.0 ® recording and editing software (Audacity Team, 2021). Next, we took three different pseudo-random samples from each recording of five seconds in duration and measured maximum and RMS amplitude using MATLAB (The Mathworks, Natick, MA, USA). Samples were evaluated to ensure stability within the five second period. We report differences in overall maximum and RMS amplitude in dB with the softest measurement across all recordings as the reference.

### *Behavioral responses to substrate-borne noise*

Adult female mason spiders were identified at field sites during egg sac construction. Once identified, a female was monitored every 15 minutes until it began collecting items to build a nest at which point, we began filming. We filmed the construction of 52 nests using GoPro (HERO 4) cameras across 30 days (July 17th - August 16th, 2018). GoPro cameras were mounted on tripods and positioned one meter from each egg sacs so as not to disrupt mason spiders during nest construction. Each nest-building bout was filmed for a total of 3-4 hours. Egg sacs were collected November 4th, 2018. Once collected, egg sacs were transported to UC Berkeley and dissected to determine offspring mortality in each egg sac, life stage (egg or spiderling), presence of mold or desiccation, and evidence of parasitism or predation. Eggs were determined to be dead if they were desiccated or covered in mold. Egg sacs were determined to be parasitized if wasps or wasp larva were present and predated if the egg sac had a large hole and was empty. Dissections were done blind to any information about treatment.

We analyzed videos of mason spider nest-building behavior using BORIS (Behavioral Observation Research Interactive Software) (Friard & Gamba, 2016). We analyzed the second hour of nest-building of each female for the following behaviors: amount of time spent on the nest vs. away from the nest, number of collecting trips per hour, and miss rate (Figure 5.1). The amount of time spent on the nest was analyzed as a proportion of the total time analyzed. Mason spiders were considered to “miss” the nest if they overshot it and had to loop back to find it or if they did at least one 360 degree turn prior to arriving at the nest (Figure 5.1B). We then calculated the miss rate by dividing the number of times a spider missed the nest by the total number of collecting trips made in an hour. Additionally, we used ImageJ (MtrackJ) software (Schneider et al., 2012) to find an average length of collecting trips for each mason spider by measuring the length of 15 consecutive outgoing trips for each spider, beginning with the first collecting trip of the second hour. We also analyzed the average “straightness” of collecting trips by calculating the ratio of the Euclidean distance to a collected item and the actual path length the spider used to get there.

### *Statistical analyses*

All statistical tests were performed in R v. 4.0.2 (R Core Team, 2020). We performed generalized linear models to assess the effect of site and treatment on nest-building behavior and offspring mortality. First, we analyzed the effect of site (predictor variable) on collecting trip length, proportion of time spent building, number of collecting trips per hour, miss rate, and average straightness (response variables) using generalized linear models with normal distributions. We found that the length of collecting trips was significantly different between sites, with spiders at Site A conducting longer collecting trips than those at Site B. To test the hypothesis that distraction rather than differences in collecting trip length were responsible for the site differences we found in our predictor variables (proportion of time spent building, number of collecting trips per hour, miss rate), we re-ran our models with collecting trip length as a random effect.

Next, we assessed the effect of mounds on offspring mortality in mason spider egg sacs between sites. In our model, we assigned dead/live offspring as our response variable with the `cbind` function in R and site as our predictor variable. We accounted for overdispersion in our data by using a quasibinomial distribution.

## 5.4 RESULTS

### *Substrate-borne noise*

Our study sites differed in amount of traffic and amplitude of substrate-borne sounds. We found that vehicles passed Site A at a rate of 5.6 vehicles per minute with an average RMS amplitude of  $21.49 \pm 0.92$  dB and a maximum RMS amplitude of  $25.87 \pm 1.09$  dB. No vehicles passed Site B during our recordings. Site B had an average RMS amplitude of  $3.97 \pm 5.18$  dB and a maximum RMS amplitude of  $5.43 \pm 4.51$  dB.

### *Behavioral responses to substrate-borne noise*

Mason spiders at Site A made significantly longer collecting trips on average than spiders at Site B ( $22.13 \pm 6.95$  cm (N=22) and  $16.89 \pm 8.02$  cm (N=22), respectively) (Figure 5.2A). However, there was no difference in the straightness of collecting trips. We found that substrate-borne noise significantly impacted nest-building behavior in multiple ways with or without controlling for collecting trip length (Table 5.1). Spiders at Site A spent 46% of their time actively building nests and made 74.67 collecting trips per hour (N = 28) while spiders at Site B spent 56% of their time actively building nests and made 122.49 collecting trips per hour (N = 24) (Figure 5.2B, C). Additionally, spiders were more likely to miss their nests in the presence of more substrate-borne noise. Spiders at Site A missed their nest in 13% (N = 28) of collecting trips while spiders at Site B missed their nest in 8% of collecting trips (N = 24) (Figure 5.2D).

We recovered 45 of 52 egg sacs. Of those, three (Site A = 2, Site B = 1) showed signs of predation or parasitism and 16 (Site A = 9, Site B = 7) showed signs of mortality due to desiccation or mold. We did not find any effect of substrate-borne noise on offspring mortality (Figure 5.3).

## 5.5 DISCUSSION

We found that substrate-borne noise levels differed across the landscape. Noise recorded in close proximity to a paved road with traffic (Site A) was about 20 dB higher than noise recorded in close proximity to a dirt road with no traffic (Site B). Despite there being little traffic on the dirt road, we recorded substrate-borne noise levels at 5.4 dB presumably from distant roads with traffic. These results suggest anthropogenic substrate-borne noise can be transmitted relatively long distances. In urban areas, human activity is responsible for up to 50 percent of substrate-borne noise (Lecocq et al., 2020), but little is known about substrate-borne noise in protected areas and the way it impacts animals sensitive to substrate-borne sounds, including invertebrates, rodents, and amphibians, among others (Cocroft & Rodriguez, 2005; Hill, 2008). Anthropogenic airborne noise can disrupt the ability of many animals to communicate and perceive natural sounds important for reproduction and survival (Shannon et al., 2016). Our results suggest that anthropogenic substrate-borne noise is widespread in protected areas and could impact many animals in ways similar to airborne noise (Buxton et al., 2017).

We found that anthropogenic substrate-borne noise significantly alters the parental care behavior of mason spiders. When road noise was relatively loud, mason spiders spent less time constructing nests and made fewer collecting trips per hour than when road noise was relatively quiet, suggesting that road traffic distracted mason spiders from parental care. Intermittent anthropogenic noise, like that from road traffic, is hypothesized to increase antipredator and hiding behavior in some animals (Francis & Barber, 2013). Animals distracted by intermittent



noise may experience an involuntary shift in their attention from an important task, like providing parental care, to an unimportant one, like hiding from a non-existent predator (Chan et al., 2010; Simpson et al., 2016). Substrate-borne noise, in particular, can resemble foraging predators and illicit antipredator behaviors (Catania, 2008). Mason spider nests are conspicuously built atop large rocks in fields that are otherwise covered in vegetation (Raboin & Elias, 2021). As with other animals engaged in nest-building behavior, mason spider nest construction likely leaves animals vulnerable to predators and therefore, vigilant (Lima, 2009). During our study, we often observed mason spiders startle, run, and freeze in the vegetation surrounding their nests, which we interpreted as hiding. Our results indicate that although spiders at our loud site conducted longer collecting trips, this alone does not account for our finding that they spent less time actively building nests and fewer collecting trips per hour. These results could be due to mason spiders hiding more often in the presence of noise.

When mason spiders did venture to collect nesting material, spiders conducted longer collecting trips at the loud site and became lost more often in returning to their nest than those at the quiet site. Mason spiders search for, identify, and assess nesting items in a way that resembles foraging. Anthropogenic noise has been shown to impair foraging behaviors of animals by not only masking the sounds of prey (Mason et al., 2016; Siemers & Schaub, 2011), but also by distracting or misleading foragers (Dominoni et al., 2020; Luo et al., 2015). For example, some animals have been found to be less efficient foragers even when anthropogenic noise does not spectrally overlap with the sounds of prey, suggesting that noise can act as a distraction or aversive stimulus (Luo et al., 2015; Purser & Radford, 2011). In the presence of distracting noise, mason spiders may cover more ground in searching for nesting material because they are less efficient in identifying and assessing items for their nest. However, differences in collecting trip length might also be accounted for by subtle differences in habitat at our two sites.

Most spiders use path integration to navigate through environments (Gaffin & Curry, 2020). Navigation, via a mechanism like path integration, requires cognitive processes that are subject to error in the presence of overwhelming or distracting stimuli (Heinze et al., 2018). We hypothesize that substrate-borne noise from road traffic disrupted cognitive processes in mason spiders, resulting in more navigational errors for spiders at our loud site. Some have suggested that anthropogenic noise may impact navigation by some whales and bats by masking echolocation ticks (McGregor et al., 2013). However, we do not know of any examples where noise has been shown to disrupt navigation through distraction. Here, we present one of the first examples of the kind.

Mason spider nests are instrumental in protecting offspring from predation, parasitism, and abiotic factors that contribute to offspring mortality, but only for a short amount of time (Raboin & Elias, 2021). We found a significant effect of substrate-borne noise on nest construction but not offspring mortality. Similarly, Pandit et al. (2021) found that noise altered the parental care behavior of bluebirds (*Sialia sialis*), but that it did not result in altered fledgling body condition. In these cases, parents maybe compensating for inadequate conditions by providing more parental care. Behavioral compensation for noise is common in communication contexts, for example many bird and frog species compensate for the presence of masking noise by adjusting frequency and amplitude of their song, however these studies often fail to include measures of reproductive success (Read et al., 2014; Roca et al., 2016). In parental care contexts, birds have been found to increase their feeding rate of nestlings to compensate for noise with varying success (Injaian, Taff, Pearson, et al., 2018; Pandit et al., 2021). Mason spiders might

have compensated for the behavioral impacts of noise (e.g., slower rate of building, getting lost more often) by taking more time overall to complete their nests. Future work should examine differences in the size of nests and the overall time building nests at quiet and loud sites.

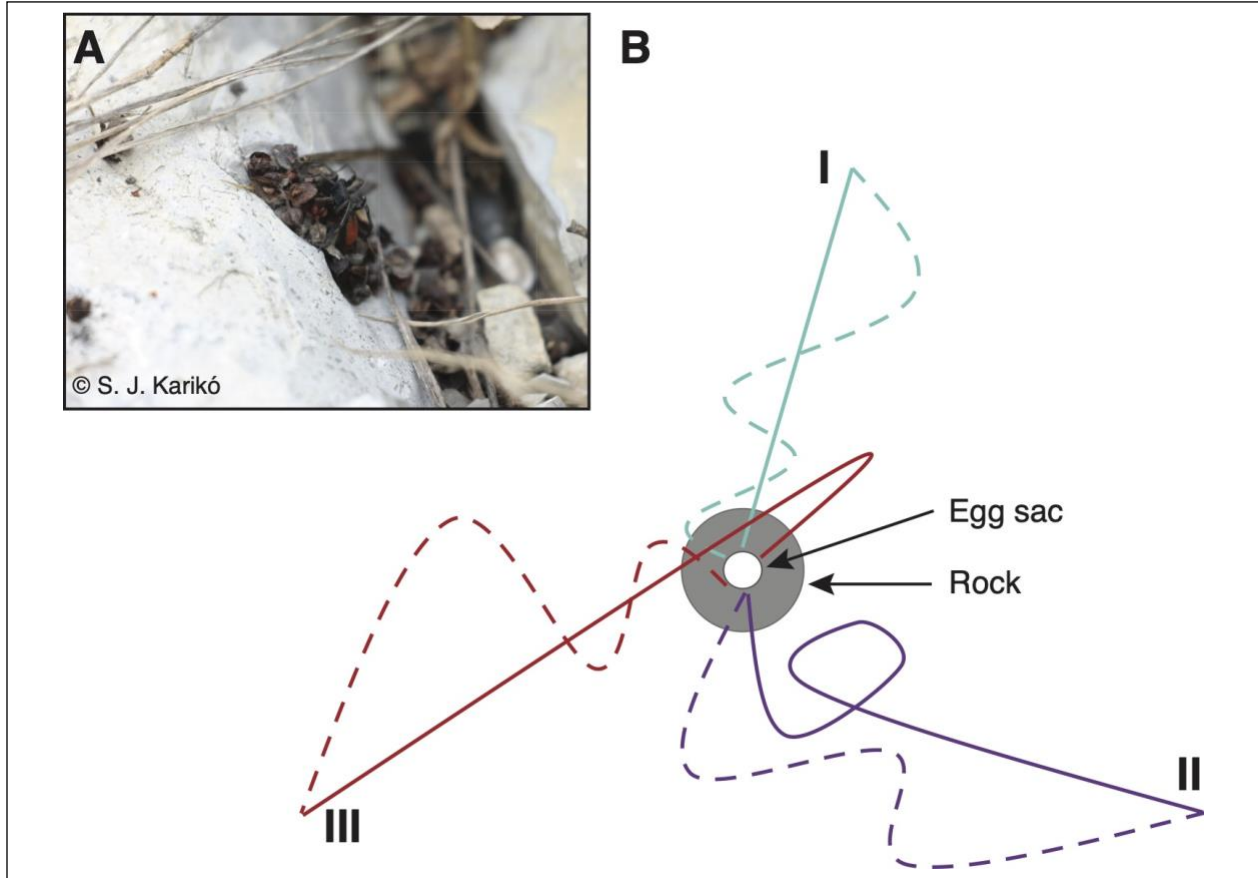
In this study, we made behavioral and mortality measurements for a single reproductive event; however, it is likely that mason spiders lay multiple egg sacs over a lifetime as seen in other spiders (Foelix, 2011). We also did not examine other measures of reproduction like the number of eggs or spiderlings in each egg sac or brood mass, which have been shown to be negatively impacted by substrate-borne noise in burying beetles (*Nicrophorus marginatus*) (Phillips et al., 2020). Under one scenario, we predict that spiders might compensate for noise at loud sites by taking longer to build a single nest. In this scenario, they may lay less egg sacs over their lifetime than spiders at quiet sites. Future studies should seek to examine mason spider behavior with more replicates of loud and quiet sites, include substrate-borne noise playbacks, and measure indirect and sublethal effects of noise on mason spider reproductive success.

Our study indicates that anthropogenic substrate-borne noise in protected areas has the potential to impact animals in profound ways and adds to a growing body of evidence that substrate-borne noise can alter animal behaviors (Caorsi et al., 2019; Phillips et al., 2020; Roberts et al., 2016; Velilla et al., 2021; Wu & Elias, 2014), including through distraction. Most terrestrial invertebrates, like mason spiders, are highly sensitive to substrate-borne sounds (Cocroft & Rodriguez, 2005). They contribute immensely to biodiversity and are experiencing declines in abundances and species across ecosystems (Hallmann et al., 2017; Lister & Garcia, 2018). While these historic changes have been attributed to several factors (Goulson, 2019) most models of invertebrate decline fail to account for the magnitude of loss (Goulson, 2019; Hladik et al., 2018; Lister & Garcia, 2018). Future research into substrate-borne noise and its impact on invertebrate species, populations, and communities will inform conservation as anthropogenic noise continues to be a widespread environmental issue.

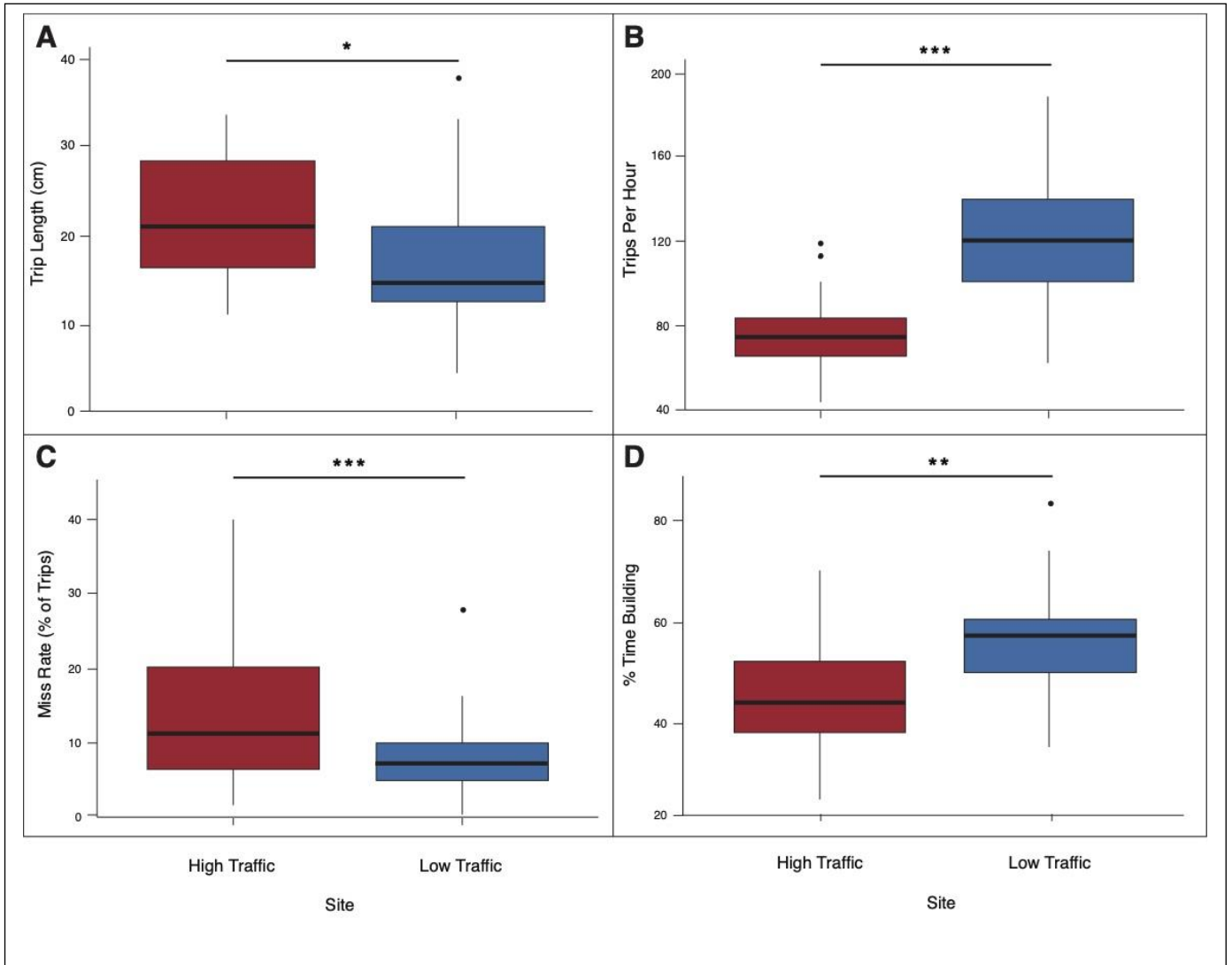
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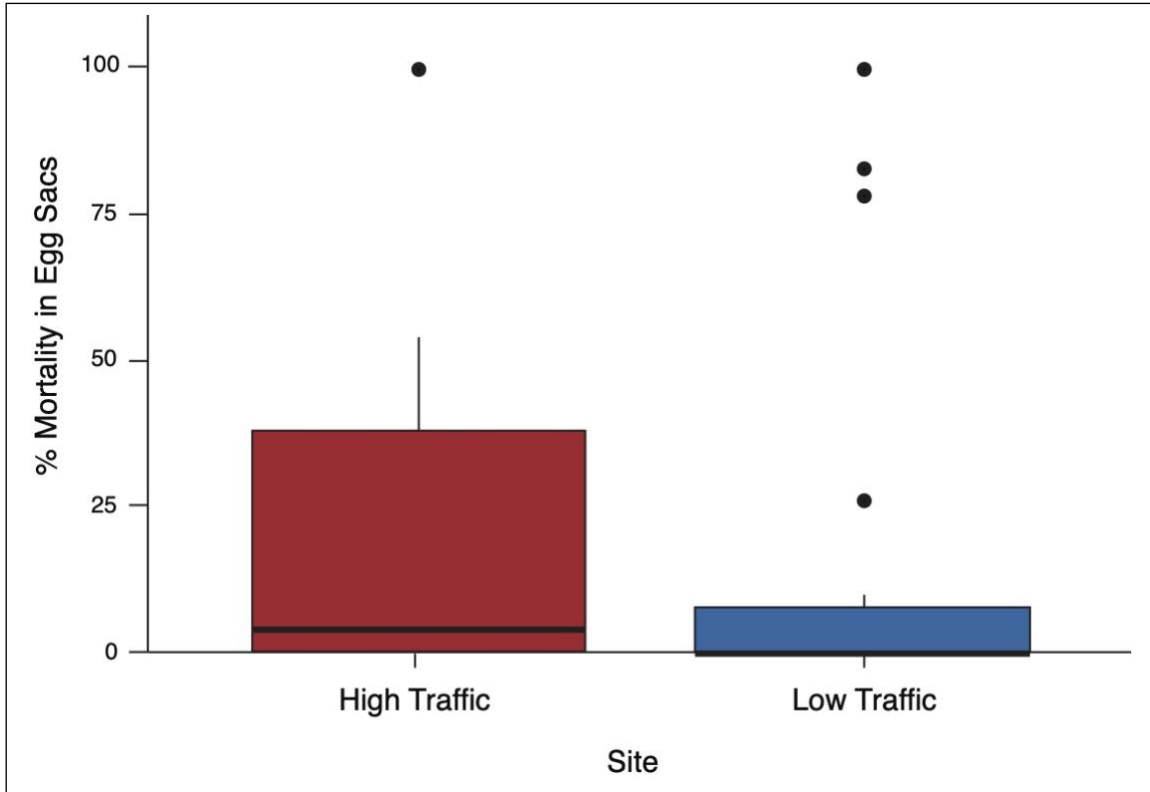
## 5.6 FIGURES



**Figure 5.1.** Mason spider constructing a nest (A) and examples of collecting trips taken to (solid line) and from (dotted line) a nest by a mason spider during nest construction (B). During most collecting trips, mason spiders returned directly to the nest (BI). However, in some instances mason spiders did at least one 360 degree turn (BII) or overshot the nest (BIII). In these cases, the collecting trip was categorized as a “miss”.



**Figure 5.2.** Box plots representing mason spider nest-building behaviors at high and low traffic sites. Nest-building behaviors include collecting trip length (A), number of collecting trips per hour (B), miss rate (C), and percentage of time spent actively nest-building (D). The line within each box denotes the median value and the lower and upper borders of each box represent the first and third quartiles. Whiskers above and below the boxes mark the range of maximally 1.5 times the interquartile range. Outliers are indicated with circles. \*  $P \leq 0.05$ , \*\*  $P \leq 0.01$ , \*\*\*  $P \leq 0.001$ .



**Figure 5.3.** Box plots representing the percentage of mason spider offspring mortality in egg sacs at high and low traffic sites. The line within each box denotes the median value and the lower and upper borders of each box represent the first and third quartiles. Whiskers above and below the boxes mark the range of maximally 1.5 times the interquartile range. Outliers are indicated with circles.

## 5.7. TABLES

Variable	Parameter	Quiet	Loud	Test Statistic	P
Collecting trip length (cm)	Mean $\pm$ SD	16.89 $\pm$ 8.02 (N=22)	22.13 $\pm$ 6.95 (N=22)	ANOVA F (1, 1) = 5.37,	0.025
Time spent building (%)	Mean $\pm$ SD	56% $\pm$ 12% (N=24)	46% $\pm$ 11% (N=28)	ANOVA F (1, 1) = 10.69	0.002
				ANCOVA F (1, 1) = 10.66	0.002
Collecting trips/hour	Mean $\pm$ SD	122.49 $\pm$ 31.95 (N=24)	74.67 $\pm$ 18.44 (N=28)	ANOVA F (1, 1) = 45.25	< 0.001
				ANCOVA F (1, 1) = 36.40	< 0.001
Miss Rate (%)	Mean $\pm$ SD	8% $\pm$ 6% (N=24)	13% $\pm$ 9% (N=28)	ANOVA F (1, 1) = 6.01	0.018
				ANCOVA F (1, 1) = 12.15	0.001
Collecting trip straightness	Mean $\pm$ SD	0.79 $\pm$ 0.09 (N=22)	0.78 $\pm$ 0.07 (N=22)	ANOVA F (1, 1) = 0.01	0.940
				ANCOVA F (1, 1) = <0.01	0.947
Mortality (%)	Mean $\pm$ SD	16% $\pm$ 33% (N=18)	21% $\pm$ 33% (N=27)	GLM (quasibinomial)	0.654

**Table 5.1.** Summary of results for linear regressions, including those that control for collecting trip length, and generalized linear model assessing the effects of low and high road traffic on mason spider nest-building behaviors and mortality in egg sacs.

## CHAPTER 6 CONCLUDING REMARKS

Invertebrates were once so abundant that their swarms darkened skies, fields, and intertidal zones. Perhaps most telling, invertebrates were once so numerous that their services to ecosystems were once incalculable and their disappearance inconceivable. However, in the last decade, the scientists that collect and identify invertebrates began to notice a concerning trend, a precipitate drop-off in the number of invertebrates visiting their traps. In fact, in parts of the world such as Germany, Puerto Rico, and the United States mainland, scientists have reported up to a 90% loss in the abundance of some invertebrate populations (Goulson, 2019).

It just so happens that over the past century, while invertebrate populations have been in free fall, sound from increased human activity has been rising (Buxton et al., 2017; Goulson, 2019). For vertebrates, anthropogenic sound has been found to impact behavior and physiology, leading, in some cases, to reductions in vertebrate species abundances (Shannon et al., 2016). However, much less is known about the impact of anthropogenic sound on invertebrates.

This dissertation attempts to bridge the gap between two fields of study, invertebrate bioacoustics and anthropogenic sound. Chapter 2 emphasizes how our scientific understandings of evolution and ecology are limited when invertebrates and their ways of sensing the world, are ignored. By including these animals in previously accepted models, we suggested a shift in the current broad theoretical framework that attempts to explain the diversity of all acoustic communication and recommended one that is more complex. In reviewing the available literature related to invertebrate bioacoustics and anthropogenic noise, we found that invertebrates might be impacted by anthropogenic noise in vastly different ways than vertebrates. Therefore, they should be studied in their own right. We presented a set of hypotheses and predictions to guide this research in the future and underlined a need to investigate substrate-borne anthropogenic noise. By focusing our efforts on a single species of invertebrate, the mason spider, in Chapters 4 and 5, we were able to test our own predictions. We found that substrate-borne anthropogenic noise can have major implications on invertebrate behavior.

In total, this collection of studies suggests that in searching for the major contributors to invertebrate decline, we might look first to anthropogenic noise. While much more research is needed, I hope this dissertation contributes to a better understanding of how noise may impact invertebrates and their conservation.

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Appendix 1. Data shown in Figures 2.1-2.4. Body mass (grams wet), temperature ( $^{\circ}\text{C}$ ), frequency ( $f$ ) in kHz, and ( $r$ ) in calls per second.



Class	Order	Family	Genus species	body mass	temp	f	r	habitat	mech.	context	Ref.
Amphibia											
Anura											
Aromobatidae											
			Anomaloglossus beebei		25.2	5.200		plants	tremulation	mating	1
Leptodactylidae											
			Leptodactylus albilabris			0.045	4	plants	percussion	mating	2
Phyllomedusidae											
			Agalychnis callidryas				13.28	plants	tremulation	agonistic interactions	3
			Agalychnis moreletii			0.007		plants	tremulation	agonistic interactions	4
Arachnida											
Araneae											
Lycosidae											
			Gladicosa gulosa	0.200	24		3.00	litter	percussion	mating	5
			Gladicosa gulosa	0.200	24		1.49	litter	stridulation	mating	5
			Hygrolycosa rubrofasciata	0.015	22		29.00	plants	percussion	mating	6
			Lycosa tarentula	1.422	24.5	1.260		litter	unknown	mating	7
			Schizocosa stridulans	0.021	21	1.045	80.00	plants	stridulation	mating	8
			Schizocosa stridulans	0.021	21	0.106	33.00	plants	tremulation	mating	8
Salticidae											
			Cosmophasis umbratica	0.014	25		7.30	plants	percussion	mating	9
			Cosmophasis umbratica	0.014	25	0.133	0.70	plants	tremulation	mating	9
			Habronattus dossenus		21	0.230	5.70	litter	stridulation	mating	10
			Habronattus dossenus		21	0.065		litter	tremulation	mating	10
			Habronattus pugillis		21	1.082		litter	unknown	mating	11
			Maratus volans	0.010	21	0.084	12.10	litter	unknown	mating	12



Dendroctonus ponderosae	0.011	22	4.900	2.10	plants	stridulation	mating	26
Ips pini	0.006	21		7.28	plants	stridulation	distress	27
Platypus quercivorus	0.008	27.5		6.31	plants	stridulation	mating	28
Polygraphus nigrielytris	0.004	21	8.018	4.34	plants	stridulation	mating	29
Polygraphus proximus	0.004	21	7.961	1.03	plants	stridulation	mating	29
Polygraphus subopacus	0.003	21	8.715	3.13	plants	stridulation	mating	29
Geotrupidae								
Trypocorpris alpinus	0.088	21	1.280		soil	stridulation	distress	30
Trypocorpris pyrenaicus	0.214	21	1.160		soil	stridulation	distress	30
Trypocorpris vernalis	0.209	21	1.910		soil	stridulation	distress	30
Ptinidae								
Xestobium rufovillosum	0.009	19		11.60	plants	percussion	mating	31
Scarabaeidae								
Trypoxylus dichotoma	25.000	25.5	0.100	0.77	soil	percussion	agonistic interactions	32
Tenebrionidae								
Eupsophulus castaneus		25		5.35	plants	percussion	mating	33
Eusattus convexus	0.054	21		25.60	litter	percussion	mating	34
Eusattus muricatus	0.081	25		23.40	litter	percussion	mating	35
Eusattus reticulatus	0.114	25		30.00	litter	percussion	mating	35
Eusattus robustus	0.189	25		12.00	litter	percussion	mating	35
Diptera								
Agromyzidae								
Liriomyza huidobrensis	0.001	25	0.231	14.44	plants	unknown	mating	36
Chloropidae								
Lipara lucens		25	0.202		plants	tremulation	mating	37
Drosophilidae								
Drosophila biarmipes		25		14.49	plants	tremulation	mating	38
Drosophila biarmipes		24.5	0.304		plants	wing fanning	mating	38
Drosophila melanogaster	0.001	24		6.90	plants	tremulation	mating	38

	<i>Drosophila persimilis</i>	0.001	23		5.78	plants	tremulation	mating	39
	<i>Drosophila sechellia</i>		25		7.13	plants	tremulation	mating	40
	<i>Drosophila suzukii</i>		25.5		12.99	plants	tremulation	mating	38
	<i>Drosophila suzukii</i>		25	0.450		plants	wing fanning	mating	38
	<i>Drosophila yakuba</i>		25		6.80	plants	tremulation	mating	40
	Embioptera								
	Clothodidae								
	<i>Antipaluria urichi</i>	0.100	21	0.473		plants	tremulation	mating	41
	Hemiptera								
	Aleyrodidae								
	<i>Aleurothrixus floccosus</i>	0.001	25	0.293	1.10	plants	tremulation	mating	42
	<i>Siphoninus phillyreae</i>	0.001	25	0.300	8.16	plants	tremulation	mating	42
	<i>Trialeurodes vaporariorum</i>	0.001	25	0.323	15.37	plants	stridulation	mating	43
	Aphalaridae								
	<i>Anoeconeossa bundoorensis</i>	0.001	21	0.660	1.11	plants	stridulation	mating	44
	Cicadellidae								
	<i>Amrasca devastans</i>		28		5.78	plants	tymbalation	mating	45
	<i>Aphrodes makarovi</i>	0.008	22.5	0.398	21.98	plants	tymbalation	mating	46
	<i>Empoasca vitis</i>	0.001	22	0.253	25.64	plants	tymbalation	mating	47
	<i>Graphocephala atropunctata</i>	0.008	25	0.235	11.62	plants	tremulation	mating	48
	<i>Homalodisca liturata</i>	0.041	25	0.095	7.90	plants	tymbalation	mating	48
	<i>Homalodisca vitripennis</i>	0.054	25	0.059	10.55	plants	tymbalation	mating	48
	<i>Psammotettix alienus</i>	0.002	22.5	0.492	166.67	plants	tymbalation	mating	49
	<i>Psammotettix alienus</i>	0.002	22.5	0.140	3.33	plants	wing fanning	mating	49
	<i>Scaphoideus titanus</i>	0.004	23.5	0.560		plants	tremulation	mating	50
	Cixiidae								
	<i>Apartus michalki</i>	0.005	20	0.375		plants	tymbalation	mating	51
	<i>Hyalesthes obsoletus</i>	0.003	23.5	0.576	14.29	plants	tymbalation	mating	52
	<i>Typhlobrixia namorokensis</i>		20.15	0.155	8.33	plants	tymbalation	mating	53

Coreidae								
Leptoglossus occidentalis	0.151	24.5	0.145		plants	percussion	mating	54
Flatidae								
Metcalfa pruinosa	0.005	23.5	0.442	40.00	plants	unknown	mating	55
Homotomidae								
Macrohomotoma gladiata	0.001	25	0.906		plants	stridulation	mating	56
Macrohomotoma robusta		25	0.672		plants	stridulation	mating	56
Membracidae								
Ennya chrysur		24.5	0.107		plants	tremulation	mating	57
Umbonia crassicornis	0.030	21	0.150		plants	tremulation	mating	58
Umbonia crassicornis	0.030	21		15.50	plants	tymbalation	mating	58
Miridae								
Macrolophus costalis	0.002	25	0.300		plants	unknown	mating	59
Macrolophus pygmaeus	0.002	25	0.295		plants	unknown	mating	59
Peloriidiidae								
Hackeriella veitchi	0.002	21	0.082		plants	unknown	unknown	60
Pentatomidae								
Acrosternum hilare	0.124	26	0.082	0.58	plants	tremulation	mating	61
Aelia acuminata	0.010	23	0.166	6.13	plants	tremulation	mating	62
Arma custos	0.054	24	0.125		plants	tremulation	unknown	63
Carpocoris fuscispinus	0.061	23	0.151	12.32	plants	tremulation	mating	62
Carpocoris pudicus	0.041	23	0.140	3.09	plants	tremulation	mating	62
Carpocoris purpuriepennis	0.054	23	0.149	4.39	plants	tremulation	mating	62
Chinavia impicticornis		26	0.093	3.99	plants	tremulation	mating	64
Chinavia ubica		26	0.123	5.98	plants	tremulation	mating	64
Chlorochroa juniperina		23	0.157	0.32	plants	tremulation	mating	62
Chlorochroa ligata	0.124	25	0.089	0.05	plants	tremulation	mating	65
Chlorochroa pinicola		23	0.172	0.20	plants	tremulation	mating	62
Chlorochroa sayi		25	0.108	0.09	plants	tremulation	mating	65

<i>Chlorochroa uhleri</i>	0.086	25	0.093	0.05	plants	tremulation	mating	65
<i>Codophila varia</i>	0.056	23	0.173	2.73	plants	tremulation	mating	62
<i>Dichelops melacanthus</i>	0.040	26	0.110	0.94	plants	tremulation	mating	66
<i>Dolycoris baccarum</i>	0.047	23	0.151		plants	tremulation	mating	62
<i>Edessa meditabunda</i>		26	0.080	3.27	plants	tremulation	mating	67
<i>Eurydema oleracea</i>	0.007	23	0.183		plants	tremulation	mating	62
<i>Eurydema ornata</i>	0.015	23	0.107		plants	tremulation	mating	62
<i>Euschistus heros</i>	0.054	26		2.33	plants	percussion	mating	68
<i>Euschistus heros</i>	0.054	26	0.200		plants	tremulation	mating	68
<i>Euschistus servus</i>	0.049	21	0.136	1.83	plants	tremulation	mating	69
<i>Graphosoma lineatum</i>		23	0.151	0.43	plants	tremulation	mating	62
<i>Graphosoma semipunctatum</i>		23	0.193	0.49	plants	tremulation	mating	62
<i>Halyomorpha halys</i>	0.095	21	0.067		plants	tremulation	mating	70
<i>Holcostethus strictus</i>		23	0.155	0.03	plants	tremulation	mating	71
<i>Holcostethus vernalis</i>		23	0.167	3.12	plants	tremulation	mating	62
<i>Murgantia histrionica</i>	0.025	26	0.087	8.55	plants	tremulation	mating	72
<i>Nezara viridula</i>	0.072	26	0.094	0.29	plants	tremulation	mating	61
<i>Palomena prasina</i>		21	0.095	3.22	plants	tremulation	mating	73
<i>Pentatoma rufipes</i>	0.078	23	0.086	0.33	plants	tremulation	mating	62
<i>Picromerus bidens</i>	0.061	23	0.107	4.37	plants	tremulation	mating	74
<i>Podisus maculiventris</i>	0.054	23	0.097	2.05	plants	percussion	mating	75
<i>Podisus maculiventris</i>	0.054	23	0.009	3.34	plants	tremulation	mating	75
<i>Podisus nigrispinus</i>	0.044	26	0.127	3.48	plants	tremulation	mating	76
<i>Staria lunata</i>	0.010	23	0.125		plants	tremulation	mating	62
<i>Thyanta custator accerra</i>		26	0.097	0.93	plants	tremulation	mating	77
<i>Thyanta pallidovirens</i>	0.062	26	0.101	1.40	plants	tremulation	mating	77
<i>Troilus luridus</i>	0.041	22	0.113	4.04	plants	tremulation	mating	63
<i>Zicrona caerulea</i>	0.015	23	0.110		plants	tremulation	unknown	63

Psyllidae

<i>Cacopsylla multijuga</i>		25	0.794	6.69	plants	stridulation	mating	78
<i>Cacopsylla oluanpiensis</i>		25	1.015	4.15	plants	stridulation	mating	78
<i>Cacopsylla picta</i>		18	0.257	1.05	plants	stridulation	mating	79
<i>Cacopsylla pyri</i>	0.002	22.5	0.689	19.61	plants	stridulation	mating	80
<i>Cacopsylla tobirae</i>	0.001	25	0.993	6.89	plants	stridulation	mating	78
<i>Diaphorina citri</i>	0.001	25.5	0.205		plants	wing fanning	mating	81
Reduviidae								
<i>Dipetalogaster maxima</i>	0.013	25	2.200	5.78	plants	stridulation	distress	82
<i>Mepraia spinolai</i>	0.179	24.9	1.524	30.54	plants	stridulation	distress	83
<i>Rhodnius prolixus</i>	0.055	25	1.500	35.60	plants	stridulation	mating	84
<i>Triatoma guasayana</i>	0.181	25	1.900	5.22	plants	stridulation	distress	82
<i>Triatoma infestans</i>	0.173	25	0.750	23.80	plants	stridulation	mating	85
<i>Triatoma sordida</i>	0.135	25	2.100	9.64	plants	stridulation	distress	82
Rhopalidae								
<i>Jadera haematoloma</i>	0.047	26	3.250	20.50	plants	stridulation	facilitating social relationships	86
Trioziidae								
<i>Aacanthocnema dobsoni</i>	0.000	20	0.240		plants	stridulation	mating	87
<i>Schedotrioza apicobystra</i>		24	0.615	74.13	plants	stridulation	mating	88
<i>Schedotrioza distorta</i>		24	0.576	12.22	plants	stridulation	mating	88
<i>Schedotrioza marginata</i>		24	0.683	61.24	plants	stridulation	mating	88
<i>Schedotrioza multitudinea</i>		24	0.869	22.89	plants	stridulation	mating	88
<i>Trioza acuminatissima</i>	0.004	25	0.717		plants	stridulation	mating	89
<i>Trioza neolitseacola</i>		25	0.728		plants	stridulation	mating	89
Hymenoptera								
Apidae								
<i>Apis cerana</i>	0.025		0.290		hives	tremulation	facilitating social relationships	90
<i>Apis mellifera</i>	0.075		0.400		hives	tremulation	facilitating social relationships	91
<i>Bombus impatiens</i>	0.070	30	0.595	0.04	plants	wing fanning	facilitating social relationships	92
<i>Bombus terrestris audax</i>	0.105	21.5	0.313		plants	wing fanning	foraging	93

Dactylurina schmidti	0.007			3.90	hives	tremulation	facilitating social relationships	94
Hypotrigona gribodoi	0.001			8.40	rock	tremulation	facilitating social relationships	94
Melipona seminigra		21	0.500	4	hives	tremulation	foraging	95
Meliponula bocandei	0.012			1.20	plants	tremulation	facilitating social relationships	94
Meliponula ferruginea	0.008		1.100	6.30	plants	tremulation	facilitating social relationships	94
Plebeina hildebrandti	0.003		0.290	4.40	hives	tremulation	facilitating social relationships	94
<b>Braconidae</b>								
Cotesia flavipes	0.001	25	0.562		plants	wing fanning	mating	96
Cotesia marginiventris	0.001	25	0.310		plants	wing fanning	mating	97
Diachasmimorpha longicaudata	0.002	27	0.165	6.36	plants	wing fanning	mating	98
Doryctobracon crawfordi	0.005	27	0.165	5.53	plants	wing fanning	mating	98
Opius hirtus		27	0.144	7.14	plants	wing fanning	mating	98
Utetes anastrephae	0.003	27	0.152	7.12	plants	wing fanning	mating	98
<b>Formicidae</b>								
Atta cephalotes	0.064	21		6.00	plants	stridulation	facilitating social relationships	99
Atta laevigata	0.034	21	0.698		plants	stridulation	facilitating social relationships	100
Atta vollenweideri	0.006	25	1.000	4.00	plants	stridulation	facilitating social relationships	101
Camponotus herculeanus	0.015	21		20.00	plants	percussion	facilitating social relationships	102
Crematogaster rogenhoferi	0.002	25	1.904		plants	stridulation	facilitating social relationships	103
Novomessor albisetosus	0.014	21		4.30	plants	stridulation	facilitating social relationships	104
Novomessor cockerelli	0.015	21		4.60	plants	stridulation	facilitating social relationships	104
Pogonomyrmex barbatus		21		665.00	plants	stridulation	facilitating social relationships	105
Pogonomyrmex desertorum				770.00	plants	stridulation	facilitating social relationships	105
Pogonomyrmex maricopa				835.00	plants	stridulation	facilitating social relationships	105
Pogonomyrmex rugosus				770.00	plants	stridulation	facilitating social relationships	105
<b>Ichneumonidae</b>								
Pimpla turionellae	0.043	25	4.510		plants	percussion	foraging	106
<b>Mutillidae</b>								
Dasymutilla foxi			0.600	10.00	soil	stridulation	mating	107



	Vespidae								
	Polistes dominulus		26	0.904	17.40	hives	percussion	facilitating social relationships	108
	Polistes fuscatus	0.177	27.5		28.96	hives	percussion	facilitating social relationships	109
	Isoptera								
	Archotermopsidae								
	Zootermopsis nevadensis	0.282	21	2.000	20.00	plants	percussion	distress	110
	Rhinotermitidae								
	Coptotermes gestroi	0.005	27		15.70	soil	percussion	alarm calling	111
	Lepidoptera								
	Drepanidae								
	Drepana arcuata	0.001	21		1.16	plants	percussion	agonistic interactions	112
	Drepana bilineata		23.5	0.097	1.28	plants	percussion	agonistic interactions	113
	Tethea or		23	0.007	3.45	plants	mandible scraping	territoriality	114
	Lycaenidae								
	Cacyreus marshalli	0.027	24	0.520	14.26	plants	unknown	facilitating social relationships	115
	Cupido argiades	0.010	24	0.477	29.57	plants	unknown	facilitating social relationships	115
	Feniseca tarquinius	0.069	24	0.302		plants	unknown	foraging	116
	Jalmenus evagoras	0.159	22	0.849	1.76	plants	stridulation	facilitating social relationships	117
	Lycaeides argyrognomon		24	0.532	15.75	plants	unknown	facilitating social relationships	115
	Lycaena dispar	0.061	24	0.689	24.30	plants	unknown	facilitating social relationships	115
	Lycaena helle	0.090	24	0.730	16.41	plants	unknown	facilitating social relationships	115
	Lycaena phlaeas	0.051	24	0.609	25.93	plants	unknown	facilitating social relationships	115
	Maculinea alcon	0.026	24	0.161	10.73	plants	unknown	facilitating social relationships	115
	Plebejus argus	0.051	24	0.318	10.52	plants	unknown	facilitating social relationships	115
	Polyommatus bellargus	0.075	24	0.528	16.51	plants	unknown	facilitating social relationships	115
	Polyommatus coridon	0.076	24	0.374	9.25	plants	unknown	facilitating social relationships	115
	Polyommatus icarus	0.051	24	0.653	20.00	plants	unknown	facilitating social relationships	115
	Scolitantides orion	0.001	24	0.508	9.85	plants	unknown	facilitating social relationships	115
	Riodinidae								

	<i>Eurybia elvina</i>	0.061	21	13.50	plants	stridulation	facilitating social relationships	118	
Neuroptera									
Chrysopidae									
	<i>Chrysoperla calocedrii</i>		25	0.071	plants	tremulation	mating	119	
	<i>Chrysoperla downesi</i>		25	0.071	plants	tremulation	mating	120	
	<i>Chrysoperla heidarii</i>		25	0.091	0.89	plants	tremulation	mating	121
	<i>Chrysoperla nipponensis</i>		25	0.049	plants	tremulation	mating	122	
	<i>Chrysoperla plorabunda</i>		25	0.064	plants	tremulation	mating	120	
Mantophasmatodea									
Austrophasmatidae									
	<i>Austrophasma caledonense</i>	0.054	20	8.15	plants	percussion	mating	123	
	<i>Austrophasma gansbaaiense</i>	0.145	20	1.98	plants	percussion	mating	123	
	<i>Austrophasma rawsonvillense</i>	0.086	20	7.07	plants	percussion	mating	123	
	<i>Hemilobophasma montaguense</i>	0.064	20	1.75	plants	percussion	mating	123	
	<i>Karoophasma biedouwensis</i>	0.069	20	10.11	plants	percussion	mating	123	
	<i>Karoophasma botterkloofense</i>	0.049	20	10.74	plants	percussion	mating	123	
	<i>Lobatophasma redelinghuysense</i>	0.082	20	15.00	plants	percussion	mating	123	
	<i>Namaquaphasma ookiepense</i>	0.088	20	10.73	plants	percussion	mating	123	
	<i>Viridiphasma clanwilliamense</i>	0.036	20	5.59	plants	percussion	mating	123	
Mantophasmatidae									
	<i>Praedatophasma maraisi</i>	0.350	20	19.12	plants	percussion	mating	123	
	<i>Sclerophasma paresisense</i>	0.135	20	11.19	plants	percussion	mating	123	
Orthoptera									
Anostostomatidae									
	<i>Deinacrida rugosa</i>	9.780	20	0.037	0.35	soil	tremulation	agonistic interactions	124
	<i>Hernideina fernovata</i>		22	1.375		plants	stridulation	distress	125
Gryllidae									
	<i>Agnotecous obscurus</i>		21	0.048		plants	tremulation	mating	126

	<i>Cardiodactylus muria</i>	21	0.038		plants	tremulation	mating	126	
	<i>Lebinthus luae</i>	21	0.084		plants	tremulation	mating	126	
	<i>Ponca hebaridi</i>	23.5	0.097		plants	tremulation	mating	127	
	Gryllotalpidae								
	<i>Gryllotalpa major</i>	2.598	20	1.971	3.14	plants	stridulation	mating	128
	Rhaphidophoridae								
	<i>Troglophilus neglectus</i>	0.136	21	0.031	0.45	rock	tremulation	mating	129
	Tetrigidae								
	<i>Bolivaritettix chinensis</i>		30		17.54	plants	unknown	mating	130
	<i>Criottetix bispinosus</i>		24		15.50	plants	unknown	mating	130
	<i>Parattetix hirsutus</i>		24		28.57	plants	unknown	mating	130
	<i>Tetrix ceperoi</i>		25	0.336		plants	tremulation	mating	131
	<i>Tetrix ceperoi</i>		25	0.585		plants	wing fanning	mating	131
	Tettigoniidae								
	<i>Conocephalus nigropleurum</i>	0.135	27.5	0.288	0.43	plants	tremulation	mating	132
	<i>Copiphora rhinoceros</i>		20		1.30	plants	tremulation	mating	133
	<i>Copiphora vigorosa</i>	0.790	30	0.012	24.00	plants	tremulation	mating	134
	<i>Gnathoclitia sodalis</i>		27.5		36.00	plants	tremulation	agonistic interactions	135
	<i>Meconema thalassinum</i>	0.066	26.1		45.50	plants	percussion	mating	136
	<i>Onomarchus uninotatus</i>	2.027	21	0.048	1.25	plants	tremulation	mating	137
	<i>Tettigonia cantans</i>	1.200	21	8.000	30	plants	stridulation	mating	138
	Plecoptera								
	Capniidae								
	<i>Isocapnia grandis</i>	0.056	22		3.48	rock	percussion	mating	139
	Chloroperlidae								
	<i>Kathroperla perdita</i>	0.045	23		27.78	rock	percussion	mating	139
	<i>Paraperla frontalis</i>	0.056	22		19.61	rock	percussion	mating	139
	Leuctridae								
	<i>Zealeuctra arnoldi</i>	0.006	24		15.84	rock	percussion	mating	140

<i>Zealeuctra claasseni</i>		24		11.53	rock	percussion	mating	140
<i>Zealeuctra hitei</i>	0.010	24		16.25	rock	percussion	mating	140
<i>Zealeuctra warreni</i>	0.006	24.5		14.60	rock	percussion	mating	141
Nemouridae								
<i>Soyedina carolinensis</i>		22.5		40.32	rock	percussion	mating	141
Peltoperlidae								
<i>Peltoperla arcuata</i>	0.025	23	0.550	7.19	rock	percussion	mating	142
<i>Peltoperla tarteri</i>		23.5	0.400	0.27	rock	percussion	mating	142
<i>Sierraperla cora</i>		23.5		6.85	rock	percussion	mating	143
<i>Soliperla fenderi</i>		22		3.27	rock	percussion	mating	139
<i>Soliperla quadrispinula</i>		22		27.03	rock	percussion	mating	139
<i>Soliperla sierra</i>	0.043	23		30.30	rock	percussion	mating	139
<i>Soliperla thyra</i>		22		52.63	rock	percussion	mating	139
<i>Tallaperla anna</i>	0.016	24.5	0.350		rock	percussion	mating	142
<i>Tallaperla elisa</i>	0.012	23	0.600	0.86	rock	percussion	mating	142
<i>Tallaperla lobata</i>	0.014	24.5	0.550		rock	percussion	mating	142
<i>Tallaperla maria</i>	0.014	23.5	0.700	2.22	rock	percussion	mating	142
<i>Viehoferla ada</i>		24		16.39	rock	percussion	mating	142
<i>Yoraperla brevis</i>		23	1.000	6.37	rock	percussion	mating	142
Perlidae								
<i>Acroneuria mela</i>	0.206	23.5		42.02	rock	percussion	mating	142
<i>Attaneuria ruralis</i>	0.378	24		28.90	rock	percussion	mating	142
<i>Beloneuria georgiana</i>		24.5		23.20	rock	percussion	mating	142
<i>Eccopectera xanthenes</i>		23.5		27.62	rock	percussion	mating	142
<i>Hansonoperla appalachia</i>	0.057	23.5		12.42	rock	percussion	mating	142
<i>Neoperla clymene</i>	0.033	24		17.89	rock	percussion	mating	142
<i>Paragnetina media</i>	0.030	25.5		21.32	rock	percussion	mating	142
Perlodidae								
<i>Acroneuria abnormis</i>	0.229	22.5		30.30	rock	percussion	mating	144

<i>Acroneuria carolinensis</i>	0.254	25	32.26	rock	percussion	mating	144
<i>Acroneuria evoluta</i>	0.104	24	20.70	rock	percussion	mating	145
<i>Acroneuria lycorias</i>	0.185	23	23.26	rock	percussion	mating	144
<i>Calineuria californica</i>	0.454		10.00	rock	percussion	mating	144
<i>Calliperla luctuosa</i>	0.064	21	40.00	rock	percussion	mating	139
<i>Chernokrilus misnomus</i>	0.064	22.75	2.49	rock	percussion	mating	146
<i>Claassenia sabulosa</i>	0.454	24	13.16	rock	percussion	mating	144
<i>Clioperla clio</i>	0.056	24	3.20	rock	percussion	mating	147
<i>Doroneuria baumanni</i>	0.542	23.5	5.43	rock	percussion	mating	145
<i>Helopicus nalatus</i>	0.118	24	12.20	rock	percussion	mating	139
<i>Hesperoperla pacifica</i>		25	8.33	rock	percussion	mating	144
<i>Hydroperla crosbyi</i>	0.043	23.5	25.00	rock	percussion	mating	143
<i>Isogenoides elongatus</i>	0.148	22	4.83	rock	percussion	mating	139
<i>Isogenoides zionensis</i>	0.185	23	2.84	rock	percussion	mating	141
<i>Isoperla burksi</i>		23.5	5.59	rock	percussion	mating	148
<i>Isoperla couchatta</i>	0.009	24.25	6.82	rock	percussion	mating	148
<i>Isoperla decepta</i>		24	1.97	rock	percussion	mating	148
<i>Isoperla fulva</i>		22	38.61	rock	percussion	mating	149
<i>Isoperla holochlora</i>	0.037	24	2.10	rock	percussion	mating	148
<i>Isoperla mohri</i>		24.25	2.83	rock	percussion	mating	148
<i>Isoperla montana</i>	0.033	23.5	5.93	rock	percussion	mating	148
<i>Isoperla mormona</i>		23	21.79	rock	percussion	mating	149
<i>Isoperla namata</i>		23.5	1.75	rock	percussion	mating	146
<i>Isoperla ouachita</i>	0.010	24	3.77	rock	percussion	mating	148
<i>Isoperla phalerata</i>		23	40.65	rock	percussion	mating	149
<i>Isoperla punctata</i>		22	23.09	rock	percussion	mating	149
<i>Isoperla sagittata</i>	0.010	22	5.51	rock	percussion	mating	148
<i>Isoperla signata</i>		23	4.43	rock	percussion	mating	146
<i>Isoperla similis</i>	0.064	24	0.43	rock	percussion	mating	148

Oconoperla innubila	0.033	24	28.17	rock	percussion	mating	147
Osobenus yakamae		24	2.56	rock	percussion	mating	147
Paragnetina fumosa	0.056	25	27.03	rock	percussion	mating	144
Paragnetina kansensis	0.082	25	10.99	rock	percussion	mating	144
Perlinella drymo	0.082	26.5	26.32	rock	percussion	mating	144
Perlinodes aurea	0.093	23.5	1.89	rock	percussion	mating	147
Phasganophora capitata		23.5	2.43	rock	percussion	mating	145
Pictetiella expansa	0.082	22	31.06	rock	percussion	mating	146
Yugus arinus	0.148	24.75	14.60	rock	percussion	mating	147
Pteronarcyidae							
Allonarcys biloba	0.753	23.5	1.90	rock	percussion	mating	150
Allonarcys proteus	0.753	23	3.25	rock	percussion	mating	150
Pteronarcella badia Alaska	0.405	25	14.93	rock	percussion	mating	150
Pteronarcella badia Colorado	0.405	22	14.08	rock	percussion	mating	150
Pteronarcella californica		25	4.55	rock	percussion	mating	140
Pteronarcella regularis	0.118	23	10.42	rock	percussion	mating	143
Pteronarcys californica	1.012	22	3.80	rock	percussion	mating	150
Pteronarcys dorsata	1.652	23.5	3.70	rock	percussion	mating	150
Pteronarcys princeps		23.5	2.58	rock	percussion	mating	143
Taeniopterygidae							
Taeniopteryx burksi		21.5	8.31	rock	percussion	mating	141
Taeniopteryx lonicera		23	7.04	rock	percussion	mating	148
Taeniopteryx maura		24	12.36	rock	percussion	mating	148
Taeniopteryx nivalis		20.5	10.00	rock	percussion	mating	139
Taeniopteryx ugola		22.5	5.13	rock	percussion	mating	141
Malacostraca							
Decapoda							
Ocypodidae							
Uca mjoebergi		0.544		soil	stridulation	mating	151

	Uca rapax		30	0.185		soil	stridulation	mating	152
	Sesamidae								
	Perisesarma eumolpe		21.5		5.00	soil	stridulation	agonistic interactions	153
	Perisesarma indiarum		21.5		7.00	soil	stridulation	agonistic interactions	153
	Mammalia								
	Carnivora								
	Phocidae								
	Halichoerus grypus	250000.000	37		30	soil	percussion	agonistic interactions	154
	Diprotodontia								
	Macropodidae								
	Macropus giganteus			0.652	8.20	plants	percussion	predator avoidance	155
	Proboscidae								
	Elephantidae								
	Elephas maximus		37		20	soil	percussion	distress	156
	Loxodonta africana	5150000.000	37	0.027		soil	vocal cords	facilitating social relationships	157
	Rodentia								
	Bathyergidae								
	Georchus capensis	270.500	24.5		26.00	soil	percussion	mating	158
	Heteromyidae								
	Dipodomys spectabilis		37		19.15	soil	percussion	predator avoidance	159
	Spalacidae								
	Spalax ehrenbergi	141.000	37	0.200	13.23	soil	percussion	agonistic interactions	160
	Tachyoryctes daemon	210.000	25		11.76	soil	percussion	facilitating social relationships	161
	Squamata								
	Iguania								
	Chamaeleonidae								
	Chamaeleo calypttratus		24	0.112		plants	tremulation	mating	162

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