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Stochastic Technology Choice Model for Consequential Life Cycle Assessment

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SCHOLARONE[™] Manuscripts

1	Stochastic Technology Choice Model
2	for Consequential Life Cycle Assessment
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9	Discussions on Consequential Life Cycle Assessment (CLCA) have relied largely on partial or
10	general equilibrium models. Such models are useful for integrating market effects into CLCA,
11	but also have well-recognized limitations such as the poor granularity of the sectoral definition
12	and the assumption of perfect oversight by all economic agents. Building on the Rectangular-
13	Choice-of-Technology (RCOT) model, this study proposes a new modeling approach for CLCA,
14	the Technology Choice Model (TCM). In this approach, the RCOT model is adapted for its use
15	in CLCA and extended to incorporate parameter uncertainties and suboptimal decisions due to
16	market imperfections and information asymmetry in a stochastic setting. In a case study on rice
17	production, we demonstrate that the proposed approach allows modeling of complex production
18	technology mixes and their expected environmental outcomes under uncertainty, at a high level
19	of detail. Incorporating the effect of production constraints, uncertainty, and suboptimal
20	decisions by economic agents significantly affects technology mixes and associated greenhouse
21	gas (GHG) emissions of the system under study. The case study also shows the model's ability to

changes in the quantity of final demand.

determine both the average and marginal environmental impacts of a product in response to

23



24 25

26 INTRODUCTION

Addressing global sustainability imperatives demands a substantial shift in today's production and consumption patterns. The 2 degree Celsius climate target and Sustainable Development Goals, for example, require a material change within policy, technology, market, and consumer behaviors.^{1–4} Understanding the consequences of such changes for the environment, however, is challenging, in part because of our limited capacity to model the wide range of market transformations that such changes may trigger.⁵ It is therefore crucial to better understand the potential effects of policies on market responses to support environmental decision-making.

An approach aiming to provide such understanding is Consequential Life Cycle Assessment (CLCA).⁶⁻¹¹ CLCA aspires to determine the environmental consequences of decisions such as the introduction of a new technology,¹²⁻¹⁴ implementing a new policy,^{15, 16} or an increase in product demand.¹⁷⁻¹⁹ While the literature on CLCA has steadily increased in recent years, debate continues regarding its operational models for the choice of technologies, and the implications of uncertainties.^{7, 20}

For modeling market effects, CLCA studies have relied largely on partial equilibrium (PE) and computational general equilibrium (CGE) models.^{9, 21–25} While PE models focus on a subset of markets within an economy, CGE models consider all sectors of an economy. The advantage of both PE and CGE models lies in their ability to determine the quantity and price of products jointly based on econometrically derived underlying data. Furthermore, such models often cover multiple regions. Kløverpris et al., for example, modeled the effect of corn consumption on the use of land based on a CGE model covering 57 sectors in 87 regions.^{25, 26}

Nevertheless, equilibrium models are also criticized for their underlying theoretical 48 assumptions, which may not be observed in real markets.^{27, 28} For example, equilibrium models 49 assume that all economic agents possess perfect oversight and that independently made decisions 50 by each agent lead to a global economic optimum. In reality, however, economic agents may 51 make suboptimal decisions, for example due to imperfect information, and most economic 52 decision-making by individual agents may not lead to a global optimum. Equilibrium models 53 also assume that all markets are in equilibrium, and that prices and demands are determined 54 based on fixed elasticities. These elasticities are, in principle, econometrically inferred, but they 55 are often based on outdated values or proxy data.²⁷ In addition, in PE and CGE models, sectoral 56 or product resolutions are generally poor. Consequently, PE and CGE models are not suitable for 57 modeling changes introduced at a detailed process-level, or determining substitution effects 58 among alternative technologies serving the same market. These substitution effects, however, are 59 crucial, for example, for modeling the environmental impact of introducing a new technology 60 producing an established product, as discussed by the authors in previous work.²⁹ 61

62

Moreover, CLCA is often exposed to large uncertainties.^{27, 30, 31} A main source of uncertainty in 63 CLCA arises from difficulties in modeling changes in the composition of technology mixes used 64 to supply markets or to produce intermediate flows as a response to changes under study.^{31, 32} 65 One way to address such uncertainties is to test multiple scenarios showing the range of potential 66 outcomes under different assumptions.^{31, 32} This approach, however, may result in an extremely 67 wide range of possible outcomes and must rely on an often subjective choice of selective 68 scenarios. Therefore, methods to systematically address large uncertainties in CLCA have been 69 called for.^{7, 33} 70

While various methods for uncertainty assessment have been utilized for LCA including 71 Analytical Error Propagation^{34, 35} and Monte Carlo simulation^{36–38}, they are typically not applied 72 to CLCAs. A major barrier to using such methods in CLCA lies in the choice of technologies, 73 74 which is typically made independently from the mathematical formulation of the consequential model, rather than being integrated into the model. Therefore, the interdependence between 75 technology choices and other sources of uncertainty, for example in process parameters and 76 prices, cannot be explored by means of mathematical uncertainty assessment in existing CLCA 77 models. 78

79

In this work, we propose the Technology Choice Model (TCM) as a new operational framework for CLCA. The model simultaneously determines technology choices in multiple markets, while systematically considering parameter uncertainties, suboptimal decisions, and factor constraints. As a result, TCM enables the modeling of both market effects and environmental impacts at a high level of detail. In the following section, we introduce the basic structure of TCM. The basic structure integrates technology choices and lays the foundation for a comprehensive uncertainty assessment. The model inherits its basic structure from the Rectangular Choice-of-Technology (RCOT) model,³⁹ which is an economic input-output model that allows for more than one technology for producing one product. Practical applications of the RCOT model have been presented for the analysis of the choice between irrigation technologies for agricultural production,⁴⁰ and the choice between bio- and petroleum-based fuels.⁴¹

In the subsequent section, we expand the basic structure of TCM to account for uncertainty and variability in process parameters, prices, factor constraints, and final demand. In addition, we allow for suboptimal decisions by economic agents. The application of TCM is demonstrated in a hypothetical case study on rice production.

96

97 BASIC STRUCTURE OF THE TECHNOLOGY CHOICE MODEL (TCM)

In RCOT, technology choices directly result from the model: given all existing technology options, a cost minimization objective determines which technologies are used to which extent, taking into account constraints in factor availability. In this section, the RCOT model of the economic input-output literature is adapted into a CLCA formulation. For this purpose, we follow the notations of the generalized calculus for LCA by Heijungs and Suh,^{42 43} which are shortly summarized in the following.

104

105 Generalized calculus for LCA

In Heijungs and Suh,⁴² the life cycle inventory model is described as a product system consisting of unit processes that exchange intermediate flows (functional flows) and elementary flows (to and from the environment). The exchange of intermediate flows between unit processes is described in the technology matrix A. In this matrix, rows represent intermediate flows, while columns represent processes. A coefficient a_{ij} of the technology matrix A describes the intermediate flow i, which is produced (for $a_{ij} > 0$) or absorbed (for $a_{ij} < 0$) by process j. The net intermediate flows leaving the product system are specified in the functional unit vector frepresenting the unit quantity that the LCA study is based upon. For an invertible technology matrix A and a given functional unit vector f, a scaling vector s can be calculated from

115
$$s = A^{-1}f.$$
 (1)

The elementary flow matrix B describes the elementary flows of the unit processes. In the ISO 116 standards on LCA (ISO 14040⁴⁴ and 14044⁴⁵), elementary flows are defined as "material or 117 energy entering the system being studied that has been drawn from the environment without 118 previous human transformation, or material or energy leaving the system being studied that is 119 released into the environment without subsequent human transformation". In the elementary flow 120 matrix, elementary flows are represented by rows, while the columns represent the same 121 processes as in the technology matrix A. The matrix is defined such that a coefficient b_{ei} shows 122 the elementary flow e of unit process j entering (for $b_{ej} < 0$) or leaving (for $b_{ej} > 0$) the 123 system. Multiplying the elementary flow matrix B with the scaling vector s yields the Life Cycle 124 Inventory (LCI) vector g representing the total elementary flows associated with the functional 125 unit *f*: 126

$$g = Bs = BA^{-1}f.$$
 (2)

127 The characterization matrix Q contains characterization factors transforming the elementary 128 flows into environmental impact flows. The characterization factor of elementary flow e for impact category z is represented by the coefficient q_{ze} of the characterization matrix Q. The total environmental impacts expressed in impact vector h are calculated such that

$$h = Qg = QBA^{-1}f.$$
 (3)

131

132 **Basic formulation of the Technology Choice Model**

The generalized calculus for LCA by Heijungs and Suh⁴² presented in the previous section 133 requires a square technology matrix A, i.e., the same number of processes and intermediate 134 flows. Which technology is utilized to produce which intermediate flow is thus pre-determined 135 for the application of Equations 1-3. In contrast, the RCOT model provides a platform under 136 which technology choices are determined by the model within market conditions and constraints. 137 138 Following RCOT, suppose that intermediate flows (functional flows such as products) can be produced by more than one process, in which case the number *m* of processes in the product 139 system exceeds the number *n* of intermediate flows.³⁹ The resulting $n \times m$ technology matrix is 140 rectangular since there are more columns than rows. For a given functional unit f and a 141 rectangular technology matrix A, multiple feasible solutions for the scaling vector s may exist, 142 because intermediate flows can be produced by more than one process. In other words, the 143 system is underdetermined. To determine a unique solution, one or more additional criteria are 144 needed to choose between feasible solutions, which in fact serve as criteria for technology 145 choices. 146

Further suppose that products will be produced in the most economical way, i.e. using technologies that belong to the least expensive production pathway. The term "production pathway" is thereby defined as the entire chain of processes used to produce a product, from the extraction of raw materials to the production of the final product. In addition, factor constraints

determine the maximum potential production volume of technologies. The production of agricultural products, for example, is constrained by the amount of cropland available. To implement criteria for technology choices and factor constraints in CLCA, we introduce the linear programming formulation of the RCOT model³⁹ to the generalized calculus for LCA.⁴²

We define a $o \times m$ factor requirement matrix F containing the factor inputs of the unit 155 processes measured in physical units. Typical factors are labor, taxes, and natural resources. The 156 factor requirement matrix F is constructed such that a coefficient f_{ki} describes the amount of 157 factor k needed by process j. We further define a $o \times 1$ unit price vector κ . An element κ_k of κ 158 represents the unit price of factor k. In addition, we define a vector c of factor constraints (the 159 term noted as 'factor endowment' in Duchin and Levine³⁹). An element c_k of c quantifies the 160 maximum available amount of factor k. The coefficients for unconstrained factors are set to ∞ . 161 Finally, we define a final demand vector y containing the final demand y_i for each commodity i. 162 In contrast to the functional unit vector f in the previous section, the final demand vector y in this 163 model contains the total final demand for products from the production system. The formulation 164 of basic TCM follows: 165

$$Min Z = \kappa Fs$$
(4)

166
$$s.t. As = y$$

- 167 $s_i \ge 0$
- $Fs \le c.$

The objective function Z of the model represents the factor costs associated with the final demand y. The constraint As = y ensures that the final demand y is produced. The constraint $s_j \ge 0$ specifies that the output of each process is positive. The last constraint ensures that the total factor use calculated by Fs does not exceed the factor availability expressed by the constraints vector c.

Finally, once the scaling vector s is calculated, the total elementary flows g as well as the environmental impacts h associated with the final demand y can be calculated from

176

$$g = Bs \tag{5}$$

177 and

178

 $h = Qg = QBs. \tag{6}$

179

180 Basic TCM has the following properties:

(1) All processes are chosen in a way that minimizes the factor costs associated with thefinal demand *y*.

(2) If an intermediate flow can be produced only by processes with identical outputs along the production chain (either single outputs or the same combination of multiple outputs), it is supplied by exactly one process.

- (3) If an intermediate flow can be produced by processes with different multiple outputs
 along the production chain (co-products), it will be supplied by either one process or by
 a combination of processes.
- (4) If there are binding factor constraints for any process on the least expensive production
 pathway, this production pathway will be used until the constraint applies and then
 complemented by one or several other production pathways representing the next lowest-cost option for the supply of the final demand.

TCM, as a generalized framework, is independent of the resolution in the underlying data. Therefore, it can utilize detailed, engineering-level data suitable for process LCAs, as well as input-output tables that RCOT utilizes. This feature also enables connecting detailed foreground systems with an input-output background system via hybrid approaches.^{43, 46}

In addition to economic factor constraints, product systems may also be constrained in the availability of environmental factors of production, e.g. emissions of pollutants, generation of waste, use of natural resources and ecosystem services. The implementation of environmental factor constraints in basic TCM is discussed in section 1 of the supporting information.

202

203 STOCHASTIC TECHNOLOGY CHOICE MODEL

The basic TCM presented in the previous section estimates the environmental impacts hassociated with the final demand y in a deterministic way based on the specified input parameters. This deterministic approach disregards the uncertainty inherent in these input parameters and may therefore yield misleading results.⁴⁷ For example, the results of a comparative LCA of two products may be misleading if the differences are not significant. Addressing uncertainties in LCA is therefore crucial for reliable decision support.⁴⁷

Uncertainties in LCA are manifold and have been categorized in various ways.⁴⁸ Huijbregts⁴⁹ distinguishes between (1) parameter uncertainty, (2) model uncertainty, (3) uncertainty due to choices, (4) spatial variability, (5) temporal variability, and (6) variability between objects/sources. While all of these uncertainties are applicable to basic TCM, in this work, we focus on (1), and (3) to (6). The term "uncertainty" is understood as discrepancy between a quantity used in the study and the true value of that quantity,¹⁰ whereas the term "variability" refers to inherent variations in the real world, e.g. seasonal and spatial variation of

precipitation.^{50, 47} In addition to uncertainties in the data typically used in LCA, TCM is subject 217 to uncertainties in factor constraints and in the final demand. For example, the amount of natural 218 resources classed as a factor constraint, and future final demands are generally estimated. 219 Various methods have been used to quantify uncertainties in LCA, e.g. analytical error 220 propagation,³⁴ calculations with fuzzy logic,⁵¹ and stochastic methods such as Monte Carlo 221 simulation.^{36–38} While the application of these methods is becoming increasingly common in 222 attributional LCA, examples and methods for a comprehensive quantitative uncertainty 223 assessment in CLCA are still missing. 224

In this section, we expand basic TCM (Equations 4-6) to account for uncertainties and variability. We therefore first identify parameters subject to uncertainty and variability in TCM, and discuss their impact on technology choices and environmental impacts. Subsequently, we stochastically propagate these uncertainties into the CLCA results using Monte Carlo simulation. The combination of basic TCM and Monte Carlo simulation is denoted 'stochastic Technology Choice Model (stochastic TCM)'.

231

232 Effect of uncertainty in TCM

233 Uncertainty and variability in input parameters

In TCM, the input parameters in *A*, *B*, *Q*, *F*, κ , *c* and *y* may be subject to both uncertainty and variability. Uncertainties in *A*, *F*, and κ affect the factor costs of the product system, which determine the choice of technologies. To illustrate the effect of uncertainties in *A*, *F*, and κ on the choice of technologies, we assume a product system in which the final demand can be produced via 3 different production pathways. The factor costs of the 3 production pathways are illustrated in Figure 1, and represent the sum of factor costs along the entire production pathway that are

associated with the production of one unit of output. The bars show single value estimates of the 240 factor costs, while the error bars specify the uncertainties. Without uncertainties, the choice of 241 technologies is obvious: all technologies on production pathway P_1 are chosen, because P_1 is the 242 lowest-cost option. Considering uncertainty, however, we find that there is a certain probability 243 for each production pathway to be the lowest-cost option. Pathway P_3 , for example, may be 244 cheaper than the other pathways if its factor cost is overestimated while the costs of the other 245 pathways are underestimated. Thus, the expected environmental impact associated with the final 246 demand is not equal to the environmental impact of a single production pathway. It is given by 247 the sum of the environmental impacts of all 3 production pathways weighted by their respective 248 probability of being the lowest-cost option. 249

Uncertainties in y directly affect the total amount of products produced, while uncertainties in both y and c may affect technology mixes if factor constraints are binding. The effect of changes in y is further discussed in the case study section.

Uncertainties in *B* and *Q* affect the elementary flows and environmental impact flows associated with a given scaling vector *s*. Uncertainties in *B* and *Q* may also affect the choice of technology in the model with environmental factor constraints (see Equation S1). 256



257

Figure 1: Factor costs of three hypothetical production pathways P_1 , P_2 , P_3 for a product. The columns show the expected values, while the error bars refer to the inherent uncertainty.

260 Uncertainty due to suboptimal decisions

In basic TCM, technologies are chosen based on minimization of factor costs. This approach 261 reflects the assumption that economic agents make optimal decisions to minimize their factor 262 costs. In reality, however, decisions in the market may not be optimal for several reasons. One is 263 imperfect information or information asymmetry in the market.⁵² Due to imperfect information, 264 decision-makers may not know all decision alternatives (e.g. suppliers for a certain raw material) 265 and the potential outcome of the decision alternatives (e.g. present and future prices). Another 266 reason lies in non-market influences such as personal relationships and patriotism. A business 267 owner, for example, may choose a supplier with whom he/she has a friendly relationship or 268 family ties. More fundamentally, human decisions are not always rational.^{53, 54} 269

In TCM, the effect of suboptimal decisions can be simulated as random noise in the choice of technologies. This random noise causes a diversification of raw material supplies and technologies used. The integration of uncertainties and suboptimal decisions is discussed in the following section.

275 Modeling of uncertainties in stochastic TCM

Stochastic TCM addresses uncertainties in basic TCM using Monte Carlo simulation. Monte Carlo simulation is a widely accepted method to quantify uncertainties in LCA.¹⁰ The idea of Monte Carlo simulation is to stochastically propagate uncertainties in input parameters into the model's results. To perform a Monte Carlo simulation of TCM, we first need to quantify the uncertainties inherent in the model.

Uncertainties and variability in input parameters can be quantified as probability distributions. Data on probability distributions of process parameters can be obtained from LCA databases. Ecoinvent,⁵⁵ for example, specifies probability distributions for almost all data items. Probability distribution for prices can be determined based on historic data under the assumption that the volatility of future prices equals the volatility of a certain timeframe in the past.

The quantification of uncertainties due to suboptimal decisions is more difficult, because no literature values on these uncertainties exist. In the context of TCM, suboptimal decisions are understood as random noise in decision-making, and hence technology choices. The choice of technologies is based on the costs of individual processes determined by factor requirements and prices. Introducing a random error to the determination of the factor cost therefore results in a random error in the choice of technologies. In this way, uncertainty due to suboptimal decisions can be translated into additional parameter uncertainty.

To simulate the effect of suboptimal decisions, we add an additional row to the factor requirement matrix F. This row specifies the requirement of an additional factor representing suboptimal decisions, which is normally distributed and has an expected value of zero. The standard deviation of the amount of this additional factor required by each process is defined as

297 percentage i of the expected cost of running the respective process. Consequently, the standard 298 deviations for the additional row in *F* are calculated by

299

$$\sigma_{C,add} = i\kappa' F. \tag{7}$$

In this equation, κ contains the expected factor prices and *F* the expected factor requirements before adding the additional row. Subsequently, we add an additional entry to the factor price vector κ , which corresponds to the newly added row in *F*. This additional entry has an expected value of 1 and a standard deviation of 0. Due to the modifications in *F* and κ , we introduce an error in the factor cost calculations for each process representing suboptimal decisions.

Once the probability distributions of all input parameters are defined, we can perform the Monte Carlo simulation. The Monte Carlo simulation is a stepwise procedure: in each step v, the parameters of TCM are varied according to their respective probability distribution, and the scalars s_{v} , g_{v} , and h_{v} are calculated. After n steps, we determine the arithmetic mean of the nresults from the Monte Carlo simulation for all elements in s, g, and h. The mean values represent the expected values of these elements, and hence the CLCA results.

The resulting mean-scaling vector s_{mean} allows us to determine the expected production 311 volumes of each process, as well as technology mixes for the production of each flow. 312 Production volumes are determined by the scaling vector entry of a technology in s_{mean} 313 multiplied by the expected value of the output of this technology specified in A. The share of a 314 technology in a technology mix is determined by the ratio of the production volume of this 315 technology and the sum of the production volumes of all technologies producing the same 316 product. The elements in g_{mean} and h_{mean} represent the expected elementary flows and 317 environmental impacts of the production system, respectively. 318

320 CASE STUDY

In this section, we apply stochastic TCM to a case study investigating the environmental 321 impacts of producing processed rice. The results from stochastic TCM are compared with those 322 from basic TCM to show the effect of uncertainties and suboptimal decisions. In the case study, 323 the final demand for processed rice is 1 Mt, while the functional unit is defined as "production of 324 1 kg processed rice". The case study is designed to demonstrate the application of the model and 325 326 does not aim to provide implications for real-life rice production. It is therefore based on hypothetical data. The rice production system, however, is inspired by an existing production site 327 in the Punjab Province in Pakistan. The data base for this case study, as well as the Matlab⁵⁶-file 328 used to determine the case study results are available in the supporting information. 329

330

331 Rice production system

The rice production system is illustrated in Figure 2. Producing processed rice in the rice 332 factory requires raw rice, electricity, and thermal energy. The raw rice is produced in small-scale 333 farms surrounding the factory. Electricity is provided by a coal-fired power plant. Thermal 334 energy can be generated by three different types of boilers fueled by natural gas, wood pellets, 335 and rice husk, respectively. Natural gas is supplied by the national gas grid. Wood pellets are 336 produced in the vicinity of the rice factory. Rice husk is co-produced with raw rice at the farms. 337 The use of rice husk for thermal energy generation, however, requires the collection of rice husk 338 at farms, involving an additional transportation demand. For simplicity, we assume that the farms 339 are located in five different zones with average transportation distances of 100 km (zone 1), 200 340 km (zone 2), 300 km (zone 3), 400 km (zone 4), and 500 km (zone 5). The amount of rice husk 341

available in each zone is assumed to be sufficient to satisfy about 20% of the thermal energydemand of the rice factory. Rice husk not used at the factory is burned at the fields.

344 Consequently, there are technology choices for the supply of rice husk and for thermal energy.

Rice husk can be provided by farms in 5 different zones, while thermal energy can be produced

346 from rice husk but also from natural gas and wood pellets. These technology choices ultimately

lead to 7 potential fuel choices for thermal energy generation: rice husk from 5 zones, natural

348 gas, and wood pellets.

For the 7 fuel types, we assume the following order of factor costs associated with thermal energy production beginning with the lowest: (1) rice husk from zone 1, (2) rice husk from zone 2, (3) natural gas, (4) rice husk from zone 3, (5) wood pellets, (6) rice husk from zone 4, and (7) rice husk from zone 5. The expected values of these factor costs and their standard deviations are illustrated in Figure S1 of the supporting information.



354

355

Figure 2: Process flow sheet of the rice production system

357 Mathematical formulation of TCM for the case study

The product system is comprised of the following 15 processes: rice factory, rice farming, rice 358 359 husk boiler, natural gas boiler, wood pellet boiler, 5 processes for rice husk collection, natural gas supply, wood pellet supply, burning of rice husk, power plant, and transportation by truck. 360 The unit processes are represented by the columns of the matrices A, B, and F. The product flows 361 include processed rice, unprocessed rice, thermal energy, rice husk at factory, rice husk at farm, 362 natural gas, wood pellets, electricity, and transportation. These flows are specified by the rows of 363 the technology matrix A and the coefficients of the final demand vector y. Furthermore, 11 364 production factors are defined: operation of rice factory, cultivation of land, 5 factors 365 representing purchases of rice husk in the 5 zones, extraction of natural gas, operation of power 366 plant, extraction of coal (used in power plant), and operation of truck. The cost factors are 367 collected in the cost factor matrix F, while the respective prices are expressed by the price vector 368 κ . For simplicity, we only consider 2 elementary flows, namely CO₂ and CH₄, which are 369 370 expressed by the rows of the intervention matrix B. The respective characterization factors are specified in the characterization matrix Q. Finally, factor constraints are defined in the 371 constraints vector c for the quantity of rice husk available in each zone. 372

We assume that parameters in *A*, *B*, *F*, κ , and *c*, except for the main process outputs in *A*, follow a normal distribution with a standard deviation of 10%. The main outputs of processes in *A* have a standard deviation of 0, because the uncertainty of other parameters is defined in relation to the main outputs. Furthermore, all uncertainties are assumed to be independent from each other.

All matrices and vectors used in the case study are provided in the Excel-file 'TCM Case Study Data.xlsx' in the supporting information.

380 Case study results

381 *Results of basic TCM*

The case study results of basic TCM are shown in columns (1) and (2) in Figure 3. Without 382 considering factor constraints (column (1)), all thermal energy is produced from rice husk from 383 zone 1 in the basic model. Rice husk from zone 1 is chosen, because it has the lowest expected 384 factor costs, and is assumed to be available without any limits. In the model with factor 385 constraints (column (2)), in contrast, thermal energy is produced from rice husk from zones 1 386 and 2, as well as from natural gas. Multiple production pathways are chosen, because the amount 387 of rice husk available in each zone is constrained. Rice husk from zone 1 represents the lowest 388 cost option, but can only satisfy about 20% of the total thermal energy demand due to factor 389 constraints. It is therefore complemented by rice husk from zone 2, which is the second lowest-390 cost option and can provide another 20% of the total energy demand. After all rice husk from 391 zones 1 and 2 is utilized, additional thermal energy needed is generated from natural gas. This 392 393 production pathway represents the next-lowest cost option and is available without limits in the model. 394

In a decision-making context, it is important to note that the global warming impact in the model result is substantially higher when considering factor constraints (about 0.9 kg CO2e/kg v.s. 1.3 kg CO2e/kg), because rice husk must be fetched from afar. Fetching rice husk from afar increases the transportation needs, and hence, also the cost of rice husk to the extent that natural gas becomes economically favorable. Without considering such constraints, costs, and associated market responses, the environmental benefit of a new technology can be overestimated.



401

Figure 3: Thermal energy production mixes (columns, left axis) and resulting climate change
 impacts (markers on solid line, right axis) based on basic TCM and stochastic TCM with and
 without binding factor constraints and suboptimal decisions.

405 *Results of stochastic TCM*

Columns (3) to (6) in Figure 3 show the results of stochastic TCM reflecting uncertainties in *A*, *B*, *F*, κ , and *c*. The consideration of these uncertainties leads to a diversification of the thermal energy production mix, where wood pellets and rice husk from zone 5 enter the market, absorbing part of the nearby rice husk supplies and natural gas. Parametric uncertainties considered in these results can arise from various origins including price fluctuations, measurement errors, and seasonality in supplies. Additionally, the effect of suboptimal decisions is illustrated in columns (4) to (6) in Figure 3. The percentage *i* specified in the column labels

states the factor *i* used to model suboptimal decisions in the respective scenario (cf. Equation 7). 413 The higher this factor *i* is, the more equally distributed the shares of individual technologies in 414 technology mixes are. Hence, suboptimal decisions lead to an increase in market share of less 415 competitive technologies in TCM. In our particular example, the consideration of uncertainties 416 and suboptimal decisions in stochastic TCM reduces the expected global warming impact of rice 417 production (solid line in Figure 3). Modeling the potential outcome of suboptimal decisions, e.g. 418 due to market imperfections and behavioral aspects, is important, because it helps understand the 419 potential discrepancy between the designed and real outcomes of a policy. 420

Figure 4 shows the effect of varying the final demand from 0 to 1 Mt, using TCM with factor 421 constraints and suboptimal decisions (i = 10%). Varying the final demand allows us to show 422 423 what would happen if consumers made different consumption decisions, e.g. due to changes in prices and incomes, legislation, or desire to contribute to sustainability. The shaded areas 424 illustrate the shares of energy carriers in the technology mix (left vertical axis). The dashed line 425 represents the marginal climate impact per functional unit, while the solid line shows the average 426 climate impact per functional unit. For a final demand below 0.2 Mt of processed rice, no factor 427 constraints apply. In this range, the amount of rice husk available in each zone is sufficient to 428 satisfy the total thermal energy demand. Consequently, mostly rice husk from zone one is used, 429 because it has the lowest factor costs. As long as no factor constraints apply, the technology mix 430 is constant and the marginal climate impact equals the average climate impact per functional 431 unit. For a final demand above 0.2 Mt, one or more factor constraints become binding: the 432 thermal energy demand of the rice factory exceeds the amount of thermal energy that can be 433 434 generated from rice husk from zone 1. This constraint results in an increased use of rice husk from other zones, natural gas, and wood pellets. Since these changes in the technology mix are 435

induced by changes in final demand, the resulting climate impact also depends on the final demand. Consequently, in presence of factor constraints, the marginal climate impact per functional unit differs from the average climate impact per functional unit. The non-linear relationship between climate impact and final demand can be determined using TCM.



440

Figure 4: Effect of varying the final demand in TCM with factor constraints and suboptimal decisions (i = 10%; cf. Equation 7) on thermal energy production mix (shaded areas, left vertical axis), marginal climate impact per functional unit (dashed black line, right vertical axis), and average climate impact per functional unit (solid black line, right vertical axis)

This result casts light on another important variable for decision-making that traditional LCAs often ignore: the scale of production and consumption.^{57–60} In the presence of constraints, the market may favor different technologies under different scales of production and consumption, which in turn affect the policy outcome. Figure 4 shows that the system may generate largerthan-expected GHG emissions when the demand for processed rice rises beyond 0.2 Mt/year.

450 DISCUSSION

There is no doubt that understanding the environmental consequences of an action, which 451 consequential LCA aims at measuring, is of great importance for policy-makers. An operational 452 framework enabling the modeling of technology choices at an engineering-level detail, however, 453 has been lacking in CLCA literature. In this paper, we presented an operational approach to 454 model discrete technology choices at process-level detail, while systematically considering 455 456 parameter uncertainty, suboptimal decisions, and factor constraints. In TCM, products are produced by technology mixes rather than single technologies, where the shares of individual 457 technologies within these technology mixes are determined by the cost of production, factor 458 constraints, uncertainties, and suboptimal decisions. By this means, TCM models market effects 459 and environmental impacts of a product system and its changes at a high level of detail. The case 460 study shows that considering uncertainty is not only essential for understanding the quality of 461 CLCA results but may also lead to substantially different results. 462

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464 Equilibrium models v.s. Technology Choice Model

Here we briefly discuss the advantages and limitations of TCM as compared to other modeling 465 approaches widely used for CLCAs. In the literature, Partial Equilibrium (PE) models and 466 General Equilibrium Models (GEMs)-and CGE models as their operational form-are 467 frequently used as a modeling framework for CLCA. Unlike PE and CGE models, TCM can be 468 formulated as a purely process-based model based on engineering estimates and measurement 469 data. PE and CGE models, in contrast, are based on elasticities derived from econometric 470 analyses. These elasticities are determined on an aggregated sector basis which typically 471 distinguishes less than 60 sectors per country. Therefore, the level of detail of PE and CGE 472

models is substantially lower compared to TCM. A high level of detail as provided by TCM is
crucial, for example, for modeling substitution effects among technologies producing the same
product, or to explore the effect of a policy on technological change.²⁹

PE and CGE models are criticized for assuming perfect cost-minimization for all economic agents. This assumption has been relaxed in stochastic TCM by allowing for suboptimal decisions. In the case study, considering suboptimal decisions has been shown to significantly affect technology choices as well as expected climate impacts. In this work, suboptimal decisions are understood as technology choices that do not result in the minimization of factor costs.

In contrast to TCM, on the other hand, PE and CGE models determine demand changes due to 482 price changes in all considered markets. In TCM, these changes are captured for intermediate 483 products, but not for final product markets, i.e. markets of products with non-zero entries in y. 484 Nevertheless, if reliable data on the price elasticity of demand is available, TCM can be 485 expanded to account for changes in final product demand due to price change. For this purpose, 486 the demand at time t + 1 can be determined in response to the price in time t using the price 487 elasticity of demand.²⁹ Also, the data requirement for TCM can be higher than for PE or CGE, 488 and unlike CGE models, TCM captures only a part of the entire economic system. However, 489 TCM can be extended to cover broader economic systems by employing hybrid approaches, 490 where the foreground system is modeled in process-level detail, while the rest of the economy is 491 modeled using input-output data.43,46 492

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Limitations and future research

In TCM, suboptimal decisions are treated as random phenomena. In reality, however, 494 suboptimal decisions may be systematic: a potential reason for suboptimal decisions is that 495 decision-makers may have objectives other than cost minimization. A company, for example, 496 may decide to use renewable feedstocks along the entire supply chain to demonstrate 497 commitment to environmental protection, even if these feedstocks may be more expensive. In 498 this case, technology choices are suboptimal regarding the cost minimization objective, but 499 optimal with regard to the decision-makers' objective. Furthermore, in this case, decisions on 500 feedstock choices are correlated. Correlations between suboptimal decisions, as well as 501 conflicting objectives are currently not captured in TCM and remain a topic for future research. 502 In addition, the degree to which economic agents make suboptimal decisions still needs to be 503 investigated in the context of TCM. Thus, our modeling approach for suboptimal decisions 504 represents only an initial step towards acknowledging the effect of suboptimal decisions in the 505 model, and needs further refinement to capture the full complexity of the topic. 506

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TCM allows the consideration of factor constraints in LCA, which have rarely been addressed 508 quantitatively in the literature. These constraints may largely affect technology mixes, as shown 509 in the case study. While our case study demonstrated the use of constraints for factor availability 510 only, in reality, production systems may be subject to additional constraints. Chemical industries, 511 for example, may be constrained in the use of certain chemicals imposing material health risks.⁶¹ 512 Operation of energy technologies is affected by regulation, transmission and storage capacity, 513 among other things.^{62, 63} Additional constraints may be further incorporated into the general 514 515 TCM framework.

It is also important to note that all parameter uncertainties in our case study were assumed to be independent from each other. In reality, uncertainties may be correlated, in which case, this correlation can be incorporated into the Monte Carlo simulation, if the correlation between input parameters can be estimated.⁴⁹

TCM presented in this paper aims at operationalizing CLCA focusing on the question of technology choice based on a transparent and reproducible modeling framework. While taking into account economic principles, uncertainty, and suboptimal decisions in technology choices, there are many more aspects to be further explored, including technology lock-in and path dependency, rebound effect, delay mechanisms, and inertia in market behavior, to name a few. Follow-up research on such topics as well as additional case studies using TCM would be desirable.

527 Operationalizing TCM at a large scale requires additional information to existing and 528 established process life cycle inventory (LCI) databases. In particular, reliable price data is 529 crucial in applying TCM to existing process LCI databases. In addition, understanding factor 530 constraints and capacity limitations would be essential in transforming existing process LCI 531 databases for TCM. In order to materialize the benefits of TCM, efforts must be made to develop 532 related meta-data protocols and collect additional data required.

533 ASSOCIATED CONTENT

Supporting information. TCM with environmental factor constraints, factor costs of thermal
 energy generation, implementation of TCM in MATLAB, MATLAB-code and data (Excel-file)
 used to calculate the case study results. This material is available free of charge via the internet at
 http://pubs.acs.org

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550 **REFERENCES**

Adoption of the Paris Agreement; FCCC/CP/2015/L.9/Rev.1; United Nations Framework Convention on
 Climate Change (UNFCCC), 2015. http://unfccc.int/resource/docs/2015/cop21/eng/l09r01.pdf.

553 2 *Fifth Assessment Report (AR5)*; Intergovernmental Panel on Climate Change (IPCC), 2014. 554 http://www.ipcc.ch/report/ar5/.

Transforming our world: the 2030 Agenda for Sustainable Development; A/RES/70/1; United Nations,
 https://sustainabledevelopment.un.org/post2015/transformingourworld/publication.

Gibon, T.; Wood, R.; Arvesen, A.; Bergesen, J. D.; Suh, S.; Hertwich, E. G. A Methodology for Integrated,
Multiregional Life Cycle Assessment Scenarios under Large-Scale Technological Change. *Environ. Sci. Technol.* **2015**, *49* (18), 11218–11226; DOI 10.1021/acs.est.5b01558.

560 5 Cucurachi, S.; Suh, S. A Moonshot for Sustainability Assessment. *Environ. Sci. Technol.* 2015, 49 (16),
 561 9497–9498; DOI 10.1021/acs.est.5b02960.

562 6 Ekvall, T.; Weidema, B. System boundaries and input data in consequential life cycle inventory analysis.
563 *Int J Life Cycle Assess* 2004, 9 (3), 161–171; DOI 10.1007/BF02994190.

- Zamagni, A.; Guinée, J.; Heijungs, R.; Masoni, P.; Raggi, A. Lights and shadows in consequential LCA. *Int J Life Cycle Assess* 2012, *17* (7), 904–918; DOI 10.1007/s11367-012-0423-x.
- 566 8 Guinée, J. B.; Heijungs, R.; Huppes, G.; Zamagni, A.; Masoni, P.; Buonamici, R.; Ekvall, T.; Rydberg, T. 567 Life Cycle Assessment: Past, Present, and Future. *Environ. Sci. Technol.* 2011, 45 (1), 90–96; DOI 568 10.1021/es101316v.
- 569 9 Earles, J.; Halog, A. Consequential life cycle assessment: a review. *Int J Life Cycle Assess* 2011, *16* (5),
 570 445–453; DOI 10.1007/s11367-011-0275-9.

571	10 Finnveden, G.; Hauschild, M. Z.; Ekvall, T.; Guinée, J.; Heijungs, R.; Hellweg, S.; Koehler, A.;
572	Pennington, D.; Suh, S. Recent developments in Life Cycle Assessment. J. Environ. Manage. 2009, 91 (1), 1-21;
573	DOI 10.1016/j.jenvman.2009.06.018.
574	11 Sandén, B. A.; Karlström, M. Positive and negative feedback in consequential life-cycle assessment. J.
575	Clean. Prod. 2007, 15 (15), 1469–1481; DOI 10.1016/j.jclepro.2006.03.005.
576	12 Pehnt, M.; Oeser, M.; Swider, D. J. Consequential environmental system analysis of expected offshore
577	wind electricity production in Germany. <i>Energy</i> 2008 , <i>33</i> (5), 747–759; DOI 10.1016/j.energy.2008.01.007.
578	13 Zink, T.; Maker, F.; Geyer, R.; Amirtharajah, R.; Akella, V. Comparative life cycle assessment of
579	smartphone reuse: repurposing vs. refurbishment. Int J Life Cycle Assess 2014, 19 (5), 1099-1109; DOI
580	10.1007/s11367-014-0720-7.
581	14 Miller, S. A.; Keoleian, G. A. Framework for Analyzing Transformative Technologies in Life Cycle
582	Assessment. Environ. Sci. Technol. 2015, 3067–3075; DOI 10.1021/es505217a.
583	15 Rajagopal, D. Consequential Life Cycle Assessment of Policy Vulnerability to Price Effects. J. Ind. Ecol.
584	2013 , 164–175; DOI 10.1111/jiec.12058.
585	16 Bento, A. M.; Klotz, R.; Landry, J. R. Are there Carbon Savings from US Biofuel Policies? The Critical
586	Importance of Accounting for Leakage in Land and Fuel Markets. EJ 2015, 36 (3), 75-109; DOI
587	10.5547/01956574.36.3.aben.
588	17 Thomassen, M. A.; Dalgaard, R.; Heijungs, R.; Boer, I. de. Attributional and consequential LCA of milk
589	production. Int J Life Cycle Assess 2008, 13 (4), 339–349; DOI 10.1007/s11367-008-0007-y.
590	18 Schmidt, J. System delimitation in agricultural consequential LCA. Int J Life Cycle Assess 2008, 13 (4),
591	350–364; DOI 10.1007/s11367-008-0016-x.

Lan, J.; Lenzen, M.; Dietzenbacher, E.; Moran, D.; Kanemoto, K.; Murray, J.; Geschke, A. Structural
Change and the Environment. *J. Ind. Ecol.* 2012, *16* (4), 623–635; DOI 10.1111/j.1530-9290.2012.00518.x.

Plevin, R. J.; Delucchi, M. A.; Creutzig, F. Using Attributional Life Cycle Assessment to Estimate Climate-

Change Mitigation Benefits Misleads Policy Makers. *J. Ind. Ecol.* 2014, *18* (1), 73–83; DOI 10.1111/jiec.12074.
Vázquez-Rowe, I.; Rege, S.; Marvuglia, A.; Thénie, J.; Haurie, A.; Benetto, E. Application of three independent consequential LCA approaches to the agricultural sector in Luxembourg. *Int J Life Cycle Assess* 2013, *18* (8), 1593–1604; DOI 10.1007/s11367-013-0604-2.

594

20

Marvuglia, A.; Benetto, E.; Rege, S.; Jury, C. Modelling approaches for consequential life-cycle
assessment (C-LCA) of bioenergy: Critical review and proposed framework for biogas production. *Renew. Sustainable Energy Rev.* 2013, 25, 768–781; DOI 10.1016/j.rser.2013.04.031.

Earles, J. M.; Halog, A.; Ince, P.; Skog, K. Integrated Economic Equilibrium and Life Cycle Assessment
Modeling for Policy-based Consequential LCA. *J. Ind. Ecol.* 2012, 375–384; DOI 10.1111/j.15309290.2012.00540.x.

Dandres, T.; Gaudreault, C.; Tirado-Seco, P.; Samson, R. Assessing non-marginal variations with
consequential LCA: Application to European energy sector. *Renew. Sustainable Energy Rev.* 2011, *15* (6), 3121–
3132; DOI 10.1016/j.rser.2011.04.004.

Kløverpris, J.; Wenzel, H.; Nielsen, P. H. Life cycle inventory modelling of land use induced by crop
consumption. *Int J Life Cycle Assess* 2008, *13* (1), 13–21; DOI 10.1065/lca2007.10.364.

610 26 Kløverpris, J.; Baltzer, K.; Nielsen, P. H. Life cycle inventory modelling of land use induced by crop
611 consumption. *Int J Life Cycle Assess* 2010, *15* (1), 90–103; DOI 10.1007/s11367-009-0132-2.

612 27 Suh, S.; Yang, Y. On the uncanny capabilities of consequential LCA. *Int J Life Cycle Assess* 2014, *19* (6),
613 1179–1184; DOI 10.1007/s11367-014-0739-9.

Dale, B. E.; Kim, S. Can the Predictions of Consequential Life Cycle Assessment Be Tested in the Real
World? Comment on "Using Attributional Life Cycle Assessment to Estimate Climate-Change Mitigation...". J. *Ind. Ecol.* 2014, 18 (3), 466–467; DOI 10.1111/jiec.12151.

617	29	Kätelhön, A.; Assen, N. von der; Suh, S.; Jung, J.; Bardow, A. Industry-Cost-Curve Approach for
618	Modeli	ing the Environmental Impact of Introducing New Technologies in Life Cycle Assessment. Environ. Sci.
619	Techno	<i>bl.</i> 2015, <i>49</i> (13), 7543–7551; DOI 10.1021/es5056512.
620	30	Baitz, M.; Albrecht, S.; Brauner, E.; Broadbent, C.; Castellan, G.; Conrath, P.; Fava, J.; Finkbeiner, M.;
621	Fischer	r, M.; Fullana i Palmer, P.; Krinke, S.; Leroy, C.; Loebel, O.; McKeown, P.; Mersiowsky, I.; Möginger, B.;
622	Pfaadt,	M.; Rebitzer, G.; Rother, E.; Ruhland, K.; Schanssema, A.; Tikana, L. LCA's theory and practice: like ebony
623	and ive	bry living in perfect harmony? Int J Life Cycle Assess 2013, 18 (1), 5–13; DOI 10.1007/s11367-012-0476-x.
624	31	Brandão, M.; Clift, R.; Cowie, A.; Greenhalgh, S. The Use of Life Cycle Assessment in the Support of
625	Robust	(Climate) Policy Making: Comment on "Using Attributional Life Cycle Assessment to Estimate Climate-
626	Change	e Mitigation". J. Ind. Ecol. 2014, 18 (3), 461–463; DOI 10.1111/jiec.12152.
627	32	Mathiesen, B. V.; Münster, M.; Fruergaard, T. Uncertainties related to the identification of the marginal
628	energy	technology in consequential life cycle assessments. J. Clean. Prod. 2009, 17 (15), 1331-1338; DOI
629	10.101	6/j.jclepro.2009.04.009.
630	33	Höjer, M.; Ahlroth, S.; Dreborg, KH.; Ekvall, T.; Finnveden, G.; Hjelm, O.; Hochschorner, E.; Nilsson,
631	M.; Palm, V. Scenarios in selected tools for environmental systems analysis. J. Clean. Prod. 2008, 16 (18), 1958-	
632	1970; I	DOI 10.1016/j.jclepro.2008.01.008.
633	34	Heijungs, R. Identification of key issues for further investigation in improving the reliability of life-cycle
634	assessn	nents. J. Clean. Prod. 1996, 4 (3-4), 159–166; DOI 10.1016/S0959-6526(96)00042-X.
635	35	Jung, J.; Assen, N. von der; Bardow, A. Sensitivity coefficient-based uncertainty analysis for multi-
636	functio	nality in LCA. Int J Life Cycle Assess 2014, 19 (3), 661–676; DOI 10.1007/s11367-013-0655-4.
637	36	Kennedy, D. J.; Montgomery, D. C.; Quay, B. H. Data quality. Int J Life Cycle Assess 1996, 1 (4), 199-
638	207; D	OI 10.1007/BF02978693.

Huijbregts, M. A. J. Part II: Dealing with parameter uncertainty and uncertainty due to choices in life cycle
assessment. *Int J Life Cycle Assess* 1998, *3* (6), 343–351; DOI 10.1007/BF02979345.

Maurice, B.; Frischknecht, R.; Coelho-Schwirtz, V.; Hungerbühler, K. Uncertainty analysis in life cycle
inventory. Application to the production of electricity with French coal power plants. *J. Clean. Prod.* 2000, *8* (2),
95–108; DOI 10.1016/S0959-6526(99)00324-8.

Duchin, F.; Levine, S. H. Sectors May Use Multiple Technologies Simultaneously: The Rectangular
Choice-of-Technology Model with Binding Factor Constraints. *Econ. Systems Res.* 2011, 23 (3), 281–302; DOI
10.1080/09535314.2011.571238.

40 López-Morales, C. A.; Duchin, F. Economic Implications of Policy Restrictions on Water Withdrawals
648 from Surface and Underground Sources. *Econ. Systems Res.* 2014, 27 (2), 154–171; DOI
649 10.1080/09535314.2014.980224.

- 41 Dilekli, N.; Duchin, F. Prospects for Cellulosic Biofuel Production in the Northeastern United States: A
 Scenario Analysis. J. Ind. Ecol. 2016, 20 (1), 120–131; DOI 10.1111/jiec.12291.
- 42 Heijungs, R.; Suh, S. *The Computational Structure of Life Cycle Assessment*; Eco-Efficiency in Industry
 and Science 11; Springer Netherlands: Dordrecht, 2002.
- 43 Suh, S. Functions, commodities and environmental impacts in an ecological–economic model. *Ecol. Econ.*2004, 48 (4), 451–467; DOI 10.1016/j.ecolecon.2003.10.013.
- 44 ISO 14040:2006 Environmental management Life cycle assessment Principles and framework;
 International Organization for Standardization, 2006.
- 658 45 SO 14044:2006 Environmental management Life cycle assessment Requirements and guidelines;
 659 International Organization for Standardization, 2006.
- 46 Suh, S.; Lenzen, M.; Treloar, G. J.; Hondo, H.; Horvath, A.; Huppes, G.; Jolliet, O.; Klann, U.; Krewitt,
- 661 W.; Moriguchi, Y.; Munksgaard, J.; Norris, G. System Boundary Selection in Life-Cycle Inventories Using Hybrid
- 662 Approaches. Environ. Sci. Technol. 2004, 38 (3), 657–664; DOI 10.1021/es0263745.

663	47 Geisler, G.; Hellweg, S.; Hungerbühler, K. Uncertainty Analysis in Life Cycle Assessment (LCA): Ca		
664	Study on Plant-Protection Products and Implications for Decision Making. Int J Life Cycle Assess 2005, 10 (3), 184		
665	192; DOI 10.1065/lca2004.09.178.		
666	48 Heijungs, R.; Huijbregts, M. A. J. A Review of Approaches to Treat Uncertainty in LCA. <i>Complexity a</i>		
667	integrated resources management meeting of the international environmental modelling and software society, vol.		
668	pp 332–339.		
669	49 Huijbregts, M. A. J. Application of uncertainty and variability in LCA. <i>Int J Life Cycle Assess</i> 1998 , <i>3</i> (
670	273–280; DOI 10.1007/BF02979835.		
671	50 Sonnemann, G. W.; Schuhmacher, M.; Castells, F. Uncertainty assessment by a Monte Carlo simulation		
672	a life cycle inventory of electricity produced by a waste incinerator. J. Clean. Prod. 2003, 11 (3), 279-292; DC		
673	10.1016/S0959-6526(02)00028-8.		
674	51 Chevalier, JL.; Le Téno, JF. Life cycle analysis with ill-defined data and its application to buildi		
675	products. Int J Life Cycle Assess 1996, 1 (2), 90–96; DOI 10.1007/BF02978652.		
676	52 Healy, P. M.; Palepu, K. G. Information asymmetry, corporate disclosure, and the capital markets:		
677	review of the empirical disclosure literature. J. Account. Econ. 2001, 31 (1-3), 405-440; DOI 10.1016/S016		
678	4101(01)00018-0.		
679	53 Becker, G. S. Irrational behavior and economic theory. J. Polit. Econ. 1962, 70 (1), 1–13; D.		
680	10.1086/258584.		
681	54 Shiv, B.; Carmon, Z.; Ariely, D. Placebo Effects of Marketing Actions: Consumers May Get What Th		
682	Pay For. J. Mark. Res. 2005, 42 (4), 383–393; DOI 10.1509/jmkr.2005.42.4.383.		
683	55 Ecoinvent. http://www.ecoinvent.ch.		
684	56 MathWorks. MATLAB. http://uk.mathworks.com/products/matlab/.		
685	57 Andersson, K.; Ohlsson, T. Life cycle assessment of bread produced on different scales. Int J Life Cycle		
686	Assess 1999, 4 (1), 25–40; DOI 10.1007/BF02979392.		

687	58 Piccinno, F.; Hischier, R.; Seeger, S.; Som, C. From laboratory to industrial scale: A scale-up framework
688	for chemical processes in life cycle assessment studies. J. Clean. Prod. 2016, 135, 1085-1097; DOI
689	10.1016/j.jclepro.2016.06.164.
690	59 Gavankar, S.; Suh, S.; Keller, A. A. The Role of Scale and Technology Maturity in Life Cycle Assessment
691	of Emerging Technologies: A Case Study on Carbon Nanotubes. J. Ind. Ecol. 2015, 19 (1), 51-60; DOI
692	10.1111/jiec.12175.
693	60 K. Suh; S. Suh; B. Walseth; J. Bae; R. Barker. Optimal Corn Stover Logistics for Biofuel Production: A
694	Case in Minnesota. Transactions of the ASABE 2011, 54 (1), 229–238; DOI 10.13031/2013.36240.
695	61 European Chemicals Agency (ECHA). REACH. http://echa.europa.eu/regulations/reach.
696	62 Kuo, YM.; Fukushima, Y. Greenhouse Gas and Air Pollutant Emission Reduction Potentials of
697	Renewable Energy-Case Studies on Photovoltaic and Wind Power Introduction Considering Interactions among
698	Technologies in Taiwan. J. Air Waste Manag. Assoc. 2009, 59 (3), 360–372; DOI 10.3155/1047-3289.59.3.360.
699	63 Fukushima, Y.; Kuo, Y. Evaluation of GHG Emission Reduction Potentials of PV System Considering
700	Power Mix Shifts. J. Energy Eng. 2008, 134 (2), 58–62; DOI 10.1061/(ASCE)0733-9402(2008)134:2(58).