

UC Irvine

UC Irvine Previously Published Works

Title

On the Utility of ToxCast™ and ToxPi as Methods for Identifying New Obesogens

Permalink

<https://escholarship.org/uc/item/2qf7z62v>

Journal

Environmental Health Perspectives, 124(8)

ISSN

1542-4359

Authors

Janesick, Amanda Shaine
Dimastrogiovanni, Giorgio
Vanek, Lenka
[et al.](#)

Publication Date

2016-08-01

DOI

10.1289/ehp.1510352

Peer reviewed



ENVIRONMENTAL HEALTH PERSPECTIVES

<http://www.ehponline.org>

On the Utility of ToxCast™ and ToxPi as Methods for Identifying New Obesogens

**Amanda Shaine Janesick, Giorgio Dimastrogiovanni,
Lenka Vanek, Christy Boulos, Raquel Chamorro-García,
Weiyi Tang, and Bruce Blumberg**

<http://dx.doi.org/10.1289/ehp.1510352>

Received: 18 June 2015

Accepted: 7 December 2015

Advance Publication: 13 January 2016

Note to readers with disabilities: *EHP* will provide a [508-conformant](#) version of this article upon final publication. If you require a 508-conformant version before then, please contact ehp508@niehs.nih.gov. Our staff will work with you to assess and meet your accessibility needs within 3 working days.



National Institute of
Environmental Health Sciences

On the Utility of ToxCastTM and ToxPi as Methods for Identifying New Obesogens

Amanda Shaine Janesick^{1,*}, Giorgio Dimastrogiovanni^{2,*}, Lenka Vanek¹, Christy Boulos¹,
Raquel Chamorro-García¹, Weiyi Tang¹, and Bruce Blumberg^{1,3}

¹Department of Developmental and Cell Biology, University of California, Irvine, Irvine, California, USA; ²Department of Environmental Chemistry, IIQAB-CSIC (Superior Council of Scientific Investigations), Barcelona, Spain; ³Department of Pharmaceutical Sciences, University of California, Irvine, California, USA. *These authors contributed equally to this work

Address correspondence to B. Blumberg, Developmental and Cell Biology, U.C. Irvine, 2011 BioSci 3, Irvine, CA 92697-2300 USA. Telephone: 949-824-8573. Fax: 949-824-4709. E-mail: blumberg@uci.edu

Running title: The utility of ToxCast for identifying obesogens

Acknowledgments: Thank you to Kris Thayer at the National Toxicology Program for providing ToxCast chemicals, and Vickie Walker, David Reif, and Kris Thayer for constructing the ToxPi models. Thank you to Heidi Käch, Cecile Hurup Hansen, Ane Nunes, and Sathya Balachander for their technical assistance. This work was supported by grants from NIH (ES015849, ES015849-03S1) and a research agreement with the National Toxicology Program at NIEHS. G.D. was supported by a predoctoral fellowship from Ministerio de Economía y Competitividad (EEBB-I-13-06750).

Competing financial interests: B.B. is a named inventor on U.S. patents 5,861,274, 6,200,802, 6,815,168 and 7,250,273 related to PPAR γ . The other authors declare they have no actual or potential competing financial interests.

Abstract

Background: In ToxCast Phase I, EPA commissioned screening of 320 pesticides, herbicides, fungicides and other chemicals in a series of high throughput assays. EPA also developed a toxicological prioritization tool, ToxPi to facilitate using ToxCast assays to predict biological function.

Objectives: We asked whether top-scoring PPAR γ activators identified in ToxCast Phase I were genuine PPAR γ activators and inducers of adipogenesis. Next, we identified ToxCast assays that should predict adipogenesis, developed an adipogenesis ToxPi and asked how well the ToxPi predicted adipogenic activity.

Methods: We used transient transfection to test the ability of ToxCast chemicals to modulate PPAR γ and RXR α , and differentiation assays employing 3T3-L1 preadipocytes and mouse bone marrow derived mesenchymal stem cells (mBMSCs) to evaluate the adipogenic capacity of ToxCast chemicals.

Results: Only 5/21 of the top scoring ToxCast PPAR γ activators were activators in our assays, 3 were PPAR γ antagonists, the remainder were inactive. The *bona fide* PPAR γ activators we identified induced adipogenesis in 3T3-L1 cells and mBMSCs. Only 7/17 chemicals predicted to be active by the ToxPi promoted adipogenesis, 1 inhibited adipogenesis, and 2/7 predicted negatives were also adipogenic. Of these 9 adipogenic chemicals, 3 activated PPAR γ and 1 activated RXR α .

Conclusions: ToxCast PPAR γ and RXR α assays do not correlate well with laboratory measurements of PPAR γ and RXR α activity. The adipogenesis ToxPi performed poorly, perhaps due to the performance of ToxCast assays. We observed a modest predictive value of ToxCast for PPAR γ and RXR α activation and adipogenesis and it is likely that many obesogenic chemicals remain to be identified.

Introduction

In 1996, the Food Quality Protection Act (FQPA 1996) and the Safe Drinking Water Act Amendments (SDWA Amendments 1996) directed the United States Environmental Protection Agency (EPA) to develop a screening program that would identify endocrine disrupting chemicals (EDCs) targeting the androgen, estrogen and thyroid signaling pathways. One key outcome is that the EPA developed the ToxCastTM program in 2007 (Dix et al. 2007). The stated goal of ToxCast was to employ high-throughput screening (HTS) assays to prioritize chemicals and use this information to inform regulatory decisions regarding thousands of environmental contaminants (Dix et al. 2007). The rationale was that a vanishingly small number of chemicals had been tested adequately for toxicity, and even fewer for endocrine-disrupting endpoints. Currently, 8 million unique, organic compounds are available for purchase (Chuprina et al. 2010), and approximately 84,000 chemicals are registered with EPA under the Toxic Substances Control Act of 1976 (TSCA 1976). The EPA Chemical Data Reporting revealed that over 7,000 chemicals are in wide use (annual production volume >100,000 pounds) (U.S. EPA 2014b). Other estimates which include data sources from the US, Canada, and Europe conclude that 30,000 chemicals are in wide commercial use (>1 ton/year) (Muir and Howard 2006). Health and toxicity data for most chemicals remains elusive because TSCA grandfathered tens of thousands of chemicals that were already on the market before 1976, none of which underwent EPA review and for which scant safety data are available.

In 2007, the National Research Council recommended *in vitro* assays to determine which toxicity pathways contribute to human disease (Collins et al. 2008; Kavlock et al. 2009). As a result, ToxCast implemented “Phase 1” *in vitro* testing. ToxCast Phase 1 was a proof-of-concept study whereby 320 pesticides, mostly agrochemicals, were selected based on historical

toxicological evidence, including *in vivo* carcinogenicity, reproductive, and developmental defects (Dix et al. 2007). Phase 1 chemicals were subjected to over 450 assays and prioritized by cluster and discriminant analysis using multiple inputs: *in silico* predictions from physicochemical properties, radioligand/enzyme biochemistry, transcription reporter assays, microarray, cytotoxicity, cell growth kinetics, and more (Dix et al. 2007). In Phase 2, 700 additional chemicals (for which toxicological data is more sparse compared to Phase 1) were tested (Kavlock et al. 2012). Since its 2007 inception, ToxCast has been reformulated as a "front end" to the EPA Endocrine Disruptor Screening Program (EDSP) to prioritize chemicals for subsequent, *in vivo* testing.

The peroxisome proliferator activated receptor (PPAR γ) is a key regulator of adipogenesis (Tontonoz and Spiegelman 2008). PPAR γ heterodimerizes with the 9-cis retinoic acid receptor (RXR) and directly promotes transcription of such key adipogenic genes as fatty acid binding protein 4 (FABP4), lipoprotein lipase (LPL) and adiponectin (ADIPOQ) (Tontonoz and Spiegelman 2008). Some environmental EDCs activate PPAR γ and RXR, thereby promoting adipogenesis whereas others promote adipogenesis by as yet unknown pathways (Janesick and Blumberg 2011b). These "obesogens" typically act at low, environmentally relevant doses (often below the established no-observed-adverse-effect-level, NOAEL) during critical windows of prenatal or postnatal development to promote obesity later in life (Grun and Blumberg 2006; Janesick and Blumberg 2011a). Obesogens can also alter the epigenetic memory of cells, creating lasting, transgenerational effects on obesity and metabolic endpoints (Chamorro-Garcia and Blumberg 2014; Chamorro-Garcia et al. 2013; Janesick et al. 2014).

When we began this project, there were no published studies investigating the reliability of ToxCast assays. Subsequently, EPA scientists have evaluated the performance of estrogen

and androgen assays as pre-screens for chemicals to be further tested in EDSP (Reif et al. 2010; Rotroff et al. 2013). Since several ToxCast assays measure the ability of chemicals to bind to, or activate PPAR γ , we first sought to test how reliable the assays (performed by commercial contractors) were in a laboratory setting. Next, prompted by a meeting hosted by NIEHS to evaluate the evidence for the involvement of EDCs in obesity and diabetes (Thayer et al. 2012), we identified a set of ToxCast assays that should predict the adipogenic potential of chemicals. These assays were used to generate a toxicological priority index (ToxPi) (Reif et al. 2010) that we expected to predict the ability of chemicals to promote adipogenesis in cell culture models. In principle, ToxCast assays and ToxPi should be useful tools for identifying chemicals that target various adverse outcome pathways. However, we show here that the results of ToxCast PPAR γ and RXR α assays do not correlate well with activity measured in a laboratory setting and that there is little agreement among ToxCast assays on the same endpoints. We further found that the ToxPi we designed for adipogenesis performed poorly in identifying potential obesogens and that the results were rife with false positives. Despite the poor overall performance of ToxCast assays and the ToxPi, some obesogens and potential obesogens were identified. We expect that if poorly performing ToxCast assays were improved (or replaced) the utility of ToxCast and ToxPi could be improved markedly and the promise of this important program realized.

Methods

ToxCastTM Phase 1 Assays

We used publically available data from three main assays reported in ToxCast Phase 1: Attagene FactorialTM Transcription Reporter System, NIH Chemical Genomics Center (NCGC) Invitrogen GeneBLAzer technology, and NovaScreen Direct Binding (Supplemental Material,

Table S1). Attagene FactorialTM is a high-throughput assay that uses capillary gel electrophoresis to track multiple reporters within the same population of transiently transfected cells simultaneously (Romanov et al. 2008). Trans-FactorialTM assays use receptor ligand binding domains (LBD) fused to the GAL4-DNA binding domain (DBD) whereas Cis FactorialTM assays use identified nuclear hormone receptor response elements without added receptors (Romanov et al. 2008). NovaScreen® (Caliper/Perkin-Elmer/Perkin Elmer Biosciences) uses fluorescence polarization (Jameson and Sawyer 1995; Jolley 1981) or scintillation proximity (Sweetnam et al. 1993) technology to detect binding of chemicals to hPPAR γ in competition with fluorescent ciglitazone, or binding to hGR in competition with [³H]-dexamethasone. These binding assays cannot differentiate whether a chemical is an activator or antagonist of a receptor, but measure apparent binding affinity, in vitro. NCGC GeneBLAzerTM technology (Invitrogen), utilizes a GAL4 DBD, nuclear receptor LBD, GAL4_{UAS} β -lactamase reporter and a FRET-based substrate which creates blue color when modified by β -lactamase (Knight et al. 2009; Zlokarnik et al. 1998). We tested the top 20 ranked activators of PPAR γ from ToxCast Phase 1 (Supplemental Material Table S2). These chemicals were supplied by the National Toxicology Program (NTP) from the same stocks that were utilized in ToxCast Phase 1. We also included chlorothalonil, which the NovaScreen PPAR γ direct binding assay indicated bound strongly to PPAR γ (Supplemental Material, Table S2). For analysis of the ToxPi, all chemicals tested were supplied by NTP and derived from ToxPi scoring of 16 different assays (Supplemental Material, Tables S1, S3), which is explained in further detail below.

Phase I ToxPi Construction

We supplied a list of gene targets to Kris Thayer (NIH/NIEHS) that literature and experience suggested could be useful to predict adipogenesis. Sixteen assays from Attagene, NovaScreen, and NCGC (Supplemental Material, Table S1) interrogated these targets and were incorporated into ToxPi models constructed by David Reif and Vickie Walker (NTP/NIEHS) (Supplemental Material, Table S3). These 16 assays were chosen because they were relevant to the biological process of adipogenesis. Three-out-of-sixteen assays showed no activation by any of the 320 ToxCast chemicals. ToxPi scores and ranking were achieved using analysis previously published (Filer et al. 2014; Reif et al. 2010; Reif et al. 2013). Briefly, each slice of the ToxPi is composed of one assay or a collection of assays. For example, the LXRE slice is one assay (Attagene, DR4-CIS assay), but the PPAR γ slice represents 3 assays (Supplemental Material, Table S1). Scores/ranking were generated by summing the AC₅₀ values of the assays within each slice for each chemical. Highly ranked chemicals either have very low AC₅₀ values for 1-2 assays, or moderately low AC₅₀ values across many assays (Supplemental Material, Table S3). To generate the input data for the ToxPi analyses we used AC₅₀ values available in the January 14, 2011 ToxCast Phase I release (Knudsen et al. 2011). For any chemical where the AC₅₀ was N/A, the AC₅₀ for that particular assay was set to 1,000,000 (1 Molar).

Transient transfection analysis

pCMX-GAL4, pCMX-GAL4-mPPAR γ , pCMX-GAL4-hRXR α were previously described (Grun et al. 2006). Transient transfections were performed in COS7 cells as described (Chamorro-Garcia et al. 2012). Briefly, COS7 cells were seeded at 15,000 cells per well in 96-well tissue culture plates in 10% calf bovine serum. The following day, cells were transfected in

Opti-MEM at ~90% confluency. 1 μg of CMX-GAL4 effector plasmid was co-transfected with 5 μg tk-(MH100)₄-luciferase reporter and 5 μg of CMX- β -galactosidase transfection control plasmids in Opti-MEM® using Lipofectamine® 2000 reagent (Invitrogen™ Life Technologies), following the manufacturer's recommended protocol. All chemicals were solvated in dimethyl sulfoxide (DMSO). After overnight incubation, the medium was replaced with Dulbecco's Modified Eagle Medium (DMEM), 10% resin charcoal stripped fetal bovine serum (FBS) plus ligands at concentrations indicated in the figure legends for an additional 24 h. DMSO concentration was maintained at 0.05% across all chemical treatments. Cells were lysed and assayed for luciferase and β -galactosidase activity as described (Forman et al. 1995). All transfections were performed in triplicate and reproduced in multiple experiments. Data are reported as fold induction or reduction over vehicle (0.1% DMSO) controls \pm S.E.M. using standard propagation of error (Bevington and Robinson 2003). EC₅₀ and IC₅₀ values (half maximal effective or inhibitory concentration) for the active chemicals were obtained using nonlinear regression, variable slope in GraphPad Prism 5.0 (Supplemental Material, Figure S1). Spirodiclofen did not plateau, therefore, it was constrained at the top dose. EC₁₀ and IC₁₀ values (10% maximal effective or inhibitory concentration) were calculated in GraphPad Quick Calc (Compute EC_{anything} from EC₅₀). The EC₅₀, EC₁₀, IC₅₀, IC₁₀ values from NCGC and ToxCast are reported from gain AC₅₀ and AC₁₀ values in the ToxCast 2014 release (Filer et al. 2014).

Adipogenesis Assays -- Cell Culture (Supplemental Material, Figure S2)

3T3-L1 cells (ATCC) were maintained in high-glucose DMEM supplemented with 10% fetal bovine serum, 2 mM L-glutamine, 50 IU/mL penicillin, and 50 $\mu\text{g}/\text{mL}$ streptomycin. 2×10^4 cells per well were seeded in 12-well plates. After 48 hours, cells were exposed to the adipogenic

cocktail MDI (500 μ M isobutylmethylxanthine, 0.25 nM dexamethasone and 5 μ g/mL insulin), 8 μ g/mL biotin and 8 μ g/mL pantothenate for 2 days. Induction media was removed and cells were exposed to test chemicals during 5 days, replacing the media every 2 days. Rosiglitazone (ROSI) and tributyltin (TBT) were used as positive controls at 100 nM and 50 nM final concentrations, respectively. All ToxCast chemicals were tested at 0.02, 0.2, 2, and 20 μ M, and DMSO concentration maintained at 0.1% across all treatments. If the chemical was toxic at 20 μ M we repeated the experiment at 10 μ M.

8×10^4 cells/well mouse bone marrow derived mesenchymal stem cells (mBMSCs) (OriCell™) were seeded in 12-well plates in basic medium: high-glucose Dulbecco's Modified Eagle Medium (DMEM; Hyclone) containing 10% calf bovine serum (Premium Select, Atlanta Biologicals), 100 IU/mL penicillin, 100 μ g/mL streptomycin, and 1 mM sodium pyruvate. mBMSCs were induced to differentiate in differentiation media (low glucose α MEM containing 15% fetal bovine serum (Premium Select, Atlanta Biologicals), 100 IU/mL penicillin, 100 μ g/mL streptomycin and 2mM L-glutamine) with adipogenic cocktail (500 μ M isomethylbutylxanthine, 1 μ M dexamethasone, 5 μ g/mL insulin) and either 500 nM ROSI, 50 nM TBT or ToxCast chemicals (as noted for 3T3-L1 cells) for 14 days, replacing the media every 3 days.

Adipogenesis Assays -- Quantitation

Cells were either fixed in 3.7% formaldehyde in phosphate buffered saline (PBS) for 30 min at room temperature (RT) for lipid quantitation, or homogenized in TriPure (Roche) for gene expression analysis. For lipid quantitation, fixed cells were washed twice with PBS and maintained in PBS overnight at 4° C to release residual Phenol Red. Background relative

fluorescent units (RFUs) of cells were measured prior to staining. Cells were stained with 1 $\mu\text{g/mL}$ Nile Red (to detect lipid accumulation) and 1 $\mu\text{g/mL}$ Hoechst 33342 (to detect nuclei as a surrogate for cell number) in PBS for 15 minutes in the dark at RT and washed twice with PBS. RFUs were measured for Hoechst 33342 (355 excitation, 460 emission) and Nile Red (485 excitation, 590 emission) in a Spectra Max Gemini XS 96-well spectrofluorometer. Background values of Hoechst and Nile Red were subtracted from the RFUs after staining and the ratio $\text{RFU}_{\text{Nile Red}}/\text{RFU}_{\text{Hoechst}}$ was calculated. Data are represented as mean \pm SEM. Total RNA was isolated using TriPure (Roche) as recommended by the manufacturer. Reverse transcription and quantitative real time RT-PCR (QPCR) were performed using Transcriptor RT and Sybr Green Master Mix (Roche). cDNA was quantitated in a Light Cycler 480 System (Roche) using primer sets listed in Supplemental Material, Table S4. Each primer set amplified a single band as determined by gel electrophoresis and melting curve analysis. QPCR data were analyzed using the $2^{-\Delta\Delta\text{Ct}}$ method (Livak and Schmittgen 2001) relative to ribosomal protein 36B4, normalizing to 0.1% DMSO vehicle. Error bars represent the SEM from four to six biological replicates calculated using standard propagation of error (Bevington and Robinson 2003).

Phase II PPAR γ Activators and ToxPi Construction

Phase II ToxCast data includes 1858 chemicals (Filer et al. 2014) which have associated Z-score corrections for each chemical-assay pair. We obtained Z-scores and $\log(\text{AC}_{50})$ values from the ToxCast Data Summary Files (<http://www.epa.gov/ncct/toxcast/data.html>) for the 320 Phase I chemicals. Z-scores are a measurement of potency relative to cytotoxicity and are often employed to remove false positive chemicals (U.S. EPA 2014a). Phase II ToxPi diagrams were constructed using Phase II 2014 release data (Filer et al. 2014) with the Phase I chemical library.

Recently, Phase II ToxPi diagrams have been constructed by first removing chemicals with low Z-scores, and then incorporating the Z-score into the magnitude of the Pi slice (Kris Thayer, personal communication). Following this method, we created new ToxPi diagrams either solely based on AC₅₀ values, or by removing chemicals with cytotoxicity Z-scores less than 3 and then correcting the magnitude of the Pi slice by incorporating Z-scores. This was accomplished by converting the AC₅₀ value to negative log molar units (e.g., 1 μM = 6), then adding the Z-score value. Final rankings were generated using the ToxPi algorithm (Reif et al. 2010). For re-evaluation of PPARγ activators using Phase II data for three PPARγ assays (Attagene, NVS, NCGC/Tox21) (Supplemental Material, Table S5), the Phase I chemical library was ranked based solely on AC₅₀ (half maximal activity concentration) values, or by removing chemicals with cytotoxicity Z-scores less than 3 (U.S. EPA 2014a) and then adding the Z-score to the AC₅₀ as described above. Final rankings were generated using the ToxPi algorithm weighting all three PPARγ assays equally.

Statistical Analysis

Statistical analysis and graphing was conducted in GraphPad Prism 5.0. One-way ANOVA was used to determine differences in relative mRNA abundance or staining among ToxCast treatment groups and negative control (DMSO). This was followed by a Dunnett's post-hoc test to ascertain statistical significance for each ToxCast treatment group compared to control (DMSO). Unpaired t-test was used to determine the significance of effects elicited by the positive controls, ROSI or TBT relative to DMSO. $P \leq 0.05$ was considered statistically significant.

Results

PPAR γ activation by Phase I ToxCastTM PPAR γ activators

Kris Thayer, David Reif and Vickie Walker (NTP/NIEHS) provided us with the 20 highest ranked activators of PPAR γ from ToxCast Phase I. AC₅₀ values were largely driven by the Attagene Gal-PPAR γ trans-FactorialTM Transcription Reporter assays (Martin et al. 2010) (Supplemental Material, Table S1). An additional chemical, chlorothalonil, was negative in Attagene assays but was reported to bind avidly (AC₅₀ = 0.6 μ M) in the NovaScreen PPAR γ ligand binding assay. We tested these activators and chlorothalonil in Cos7 cells using transient transfection assays and found that only 4/21 chemicals (spirodiclofen, zoxamide, triphenyltin, and triflumizole) activated GAL-PPAR γ (Figure 1A). Three-out-of-twenty of the reported PPAR γ agonists (fluazinam, alachlor, and acetochlor) did not activate PPAR γ ; rather they were weak antagonists in competition assays against 50 nM Rosiglitazone (Figure 1B). The GAL4 DBD alone was not activated or repressed by any of the chemicals tested or the controls (data not shown). Quinoxifen was subsequently identified as weakly active at 100 μ M and 33 μ M doses (see Figure 4A). The EC₅₀, EC₁₀, IC₅₀, and IC₁₀ values for the active chemicals were calculated, reported, and compared to Attagene and NCGC/Tox21 assay data (release 2014) (Filer et al. 2014) in Figure 1C. Subsequent to our studies, EPA scientists have re-evaluated the results of ToxCast Phase 1 with respect to non-specific induction of reporter gene activity in some assays and tested additional chemicals to yield ToxCast Phase 2 (Kavlock et al. 2012).

Effects of spirodiclofen and zoxamide on adipogenesis in cell culture models

We next tested whether spirodiclofen or zoxamide induced adipogenesis in mouse bone marrow derived mesenchymal stem cells (mBMSCs) and 3T3-L1 preadipocytes since most (but

not all) PPAR γ activators increase adipogenesis (Janesick and Blumberg 2011b). Triflumizole and triphenyltin are not reported here because we previously published the results obtained with triflumizole (Li et al. 2012) and triphenyltin is a known obesogen (Kanayama et al. 2005). In 3T3-L1 preadipocytes, spirodiclofen induced adipogenesis at all doses tested, and zoxamide induced adipogenesis at the lowest dose (Figure 2A). In mBMSCs, spirodiclofen induced adipogenesis at 10 and 20 μ M whereas zoxamide induced differentiation at 2 and 10 μ M (Figure 2B). Zoxamide was toxic to 3T3-L1 cells at ≥ 10 μ M and to mBMSCs at ≥ 20 μ M. QPCR evaluated expression of genes known to be involved in different phases of adipogenic differentiation (FABP4 for preadipocytes, Fsp27 for lipid droplet accumulation, and LPL for terminal differentiation). Corresponding increases in adipogenic gene expression were observed for spirodiclofen and zoxamide (Figure 2A, B).

Selection of assays for ToxPi construction

ToxPi (Toxicological priority index) is a prioritization tool that combines information from several assays to link chemicals with a particular biological process (Reif et al. 2010). Each ToxPi slice represents one assay or a collection of assays on the same target (see key, Figure 3A). Sixteen assays were chosen for the adipogenesis ToxPi, which were grouped into 8 slices (Supplemental Material, Table S1). For example, the PPAR γ slice consists of 3 assays because these assays were all performed on the same target, PPAR γ , but executed by different companies with different methods (e.g., binding assay versus activation assay). The size of each ToxPi slice reflects the magnitude of the AC₅₀ values (the lower the values, the larger the slice). PPAR γ and RXR α were chosen for their ability to regulate fat cell development (Grun and Blumberg 2006; Rosen and MacDougald 2006; Tontonoz and Spiegelman 2008). Proteins of the

C/EBP family function downstream and upstream of PPAR γ to stabilize the adipogenic fate (Darlington et al. 1998; Rosen and MacDougald 2006). The glucocorticoid receptor (GR) and sterol regulatory element-binding protein (SREBP) both regulate lipid metabolism (Peckett et al. 2011; Raghov et al. 2008). LXR is responsible for adipocyte function and regulates SREBP-1c expression (Calkin and Tontonoz 2012).

No chemical was found to be active in all 16 assays. The highest scoring chemicals were active in 5-6 assays or 4-5 slices. Medium scoring chemicals were active in 1-3 slices/assays and low scoring chemicals were not active in any assays. The low scoring “negatives” also did not demonstrate activity in any other ToxPis that represented collective assays on feeding behavior, islet cell function, and insulin sensitivity (Kris Thayer, personal communication). Figure 3A and Supplemental Material Table S3 show 24 top, medium, low (zero/negative) scoring chemicals obtained by ToxPi analysis. Figure 3B shows how these 24 chemicals rank in context with all ToxCast Phase I chemicals. We tested these chemicals in PPAR γ and RXR α activation assays and found that pyridaben, quinoxifen, and triphenyltin activated PPAR γ (Figure 4A) and fludioxonil activated RXR α (Figure 4B). This means that only 2/11 high scoring ToxPi chemicals and 2/6 medium scoring ToxPi chemicals could activate PPAR γ or RXR α despite that Attagene assays reported all as PPAR γ or RXR α activators. Triphenyltin, a known PPAR γ and RXR α agonist, was not on the Attagene list of RXR α activators (false negative). As expected, none of the low (zero) scoring chemicals activated PPAR γ or RXR α .

Effects of Phase I ToxPi predicted adipogenic chemicals on adipogenesis.

The main goal of the adipogenesis ToxPi was not to assess individual receptor or transcription factor activators, but rather, to predict which chemicals might activate one or more

key pathways that collectively promote adipogenesis. We tested all of the top, medium, low (zero/negative) scoring chemicals obtained by ToxPi analysis in the 3T3-L1 preadipocyte model (Figure 5). We did not test triphenyltin, triflumizole, or bisphenol A here, because these have been previously published (Chamorro-Garcia et al. 2012; Kanayama et al. 2005; Li et al. 2012; Masuno et al. 2002). Counting these known obesogens, 7/17 of the top and medium scoring chemicals demonstrated adipogenic activity. 2/7 of the negative, zero-scoring chemicals, acetamiprid and pymetrozine, promoted adipogenesis. Pyridaben strongly inhibited adipogenesis, despite its ability to activate PPAR γ (Supplemental Material, Figure S3). Figure 5 shows the results of an example 3T3-L1 assay, performed in triplicate and corresponding gene expression in Figure 6. QPCR in Figure 6 evaluated expression of genes known to be involved in different phases of adipogenic differentiation (Zfp423 for early commitment, FABP4 for preadipocytes and LPL for terminal differentiation). These assays were repeated multiple times by a succession of laboratory personnel and showed similar results (data not shown). All chemicals that were positive in 3T3-L1 assays were tested in mBMSCs to evaluate which chemicals can promote differentiation of MSCs into adipocytes. We identified fludioxonil and quinoxifen as obesogenic chemicals that could differentiate these uncommitted stem cells into adipocytes (Figure 7), whereas the others could only induce differentiation in cells already committed to the adipocyte lineage (preadipocytes). Gene expression analysis by QPCR verified the in vitro cell culture results (Figure 7).

ToxCast/NCGC Discrepancies and Summary of Phase I ToxCast and ToxPi Data

We created summary tables that describe all chemicals we tested and compare the various activation and adipogenic assays available. Table 1 is a summary of the ToxCast, PPAR γ

activation analysis and Table 2 is a summary of the ToxPi analysis. Supplemental Material Tables S6 and S7 are a continuation of these tables and show a comparison of AC_{50} values from our assays, Attagene (ATG), NovaScreen (NVS), and NCGC/Tox21. Notable discrepancies between the assay platforms on the same receptor endpoint are apparent. The most fundamental problem is that the three main nuclear receptor assays (ATG, NVS, & NCGC) do not overlap nearly as well as would be expected, even using the most current Phase II ToxCast data release (Filer et al. 2014) (Figure 8). The Attagene PPAR γ agonist assay has proportionally more overlap with the NCGC/Tox21 PPAR γ *antagonist*, rather than the agonist assay. When taking the intersection of Phase II, 2014 release data for ATG, NVS and NCGC/Tox21 agonist assays, we find 17 chemicals, and even these are unlikely to be true activators. For example, docusate sodium is a detergent and while it was recently shown to activate PPAR γ (Temkin et al. 2015), it is also likely to be a pan-assay interference compound (Dahlin et al. 2015) because it is positive in 214 ToxCast assays. Fluazinam is a PPAR γ antagonist (Figure 1) that did not promote adipogenesis, and tetrac is a thyroid hormone analog, unlikely to have affinity for PPAR γ . Moreover, Attagene identifies 100 chemicals that activate RXR β but not RXR α (Supplemental Material, Figure S4), despite that receptor selective rexinoids are not known to exist because the same residues make contact with ligand in all three RXR subtypes, RXR α , RXR β and RXR γ (Love et al. 2002). These results are *prima facie* implausible and should have indicated to the screeners that one or more of the assays are problematic.

Z-score corrections are currently being implemented by ToxCast as a mechanism to remove false positive chemicals. Z-scores are a measurement of potency relative to cytotoxicity of each chemical-assay pair (U.S. EPA 2014a). Typically, a chemical with a Z-score less than 3 is either removed, or flagged as a “non-selective” hit (U.S. EPA 2014a). Recently, Phase II

ToxPi diagrams have been constructed by first removing chemicals with low Z-scores, and then correcting the magnitude of the Pi slice by adding the Z-score to the $\log(AC_{50})$ (Kris Thayer, personal communication). We have used Phase II, 2014 release data (Filer et al. 2014) to regenerate a list of PPAR γ activators and adipogenesis ToxPi employing Z-score corrections for the Phase I chemical library. On the positive side, applying Z-scores, nearly all false negative ToxPi chemicals are lost, or ranked very low. Incredibly, all true positives we identified are also lost (Supplemental Material, Figure S5A). (Z,E)-Fenproximate continues to be ranked high in all analyses, but we showed that this chemical was not adipogenic. Acetamiprid and Pymetrozine, which we found to be adipogenic, were false negatives in ToxPi Phase I 2011 release (Knudsen et al. 2011) and their ranking does improve slightly in 2014 release (Filer et al. 2014), especially with Z-scores incorporated. Top scoring chemicals from the ToxPi using Phase I chemical library but Phase II data are shown with (Supplemental Material, Figure S6A) and without (Supplemental Material, Figure S6B) Z-score correction. Pyridaben appears in the high ranking, non-Z-score corrected hits, yet, we found that this chemical inhibited adipogenesis (Supplemental Material, Supplemental Figure S3).

When investigating the PPAR γ assays only, we found that Phase II data using Z-score correction identifies an almost entirely new set of top-scoring chemicals (Supplemental Material, Table S5). As we observed with the Phase II ToxPi data, Z-scores can alter results dramatically (Supplemental Material, Figure S5B). Nine-out-of-twelve false positive chemicals are removed, while 3/12 (tebufenpyrad, pyraclostrobin, and dimethenamid) remain. Unfortunately, the bona fide PPAR γ activators, quinoxifen and triflumizole are also eliminated by their Z-scores, and triphenyltin is an inactive chemical in all three PPAR γ assays. Zoxamide is only called active in 1/3 assays and has a relatively small Z-score, and therefore is ranked quite low. Instead,

chemicals such as atrazine and 2,4-D are ranked higher. Since these chemicals were in hand, we tested them on PPAR γ and RXR α and found them to be inactive (Supplemental Material, Figure S7). We created a second table without Z-score correction and found that the list of chemicals also differs from the original list (Supplemental Material, Table S5). This is primarily due to the poor correlation between assay results in Phase I 2011 release (Knudsen et al. 2011) versus Phase II 2014 release (Filer et al. 2014) (Supplemental Material, Figure S8). We tested the top-scoring chemical, triclosan, and found it to be inactive on PPAR γ (Supplemental Material, Figure S7). Taken together, these data suggests that recent "refinements" made to analysis of ToxCast data alone do not improve their ability to measure PPAR γ or RXR activity or to predict adipogenic capacity.

Discussion

The ToxCastTM program is a high-throughput screening effort initiated by the EPA to predict chemical toxicity, potentially for use in risk assessment. More recently, ToxCast has been repurposed as a "front-end" or prescreen to identify chemicals that should be screened using the full battery of tests in the endocrine disruptor screening program (EDSP). This approach shows some promise for androgen and estrogen receptor disruption (Rotroff et al. 2013; Rotroff et al. 2014). We evaluated the ability of ToxCast assays to predict obesogenic chemicals by measuring their ability to activate PPAR γ and RXR α in transient transfection assays and to promote adipogenesis in MSCs and 3T3-L1 preadipocytes. Our results show that ToxCast assays were able to successfully predict some bona fide obesogens, however, this success was accompanied by numerous false positives and a few false negatives. In our first study, we worked with a list of 21 chemicals reported to be PPAR γ activators in ToxCast Phase 1 assays.

We could only validate 5/21 of these chemicals as PPAR γ activators (triphenyltin, zoxamide, triflumizole, spirodiclofen, and quinoxifen); moreover, 3/21 were weak PPAR γ antagonists (fluazinam, acetochlor, and alachlor). On the positive side, each of the bona fide PPAR γ activators promoted adipogenesis in 3T3-L1 cells and mBMSCs, suggesting that well-executed PPAR γ -activation assays could be informative.

In our second study, we identified a set of gene targets whose regulation could be relevant to adipogenesis and used ToxCast phase 1 assays and ToxPi software to generate a list of candidate chemicals for testing. We found that 5/11 high scoring and 2/6 medium scoring ToxPi chemicals were adipogenic in 3T3-L1 preadipocytes. Surprisingly, 2/7 of the predicted negatives were adipogenic. Of the 9 chemicals that could differentiate 3T3-L1 preadipocytes, only 3 (quinoxifen, triflumizole (Li et al. 2012), and fludioxonil), were able to induce adipogenesis in uncommitted mBMSCs. Perhaps not surprisingly, these 3 chemicals were PPAR γ (quinoxifen, triflumizole) or RXR (fludioxonil) activators. One additional notable result was that the PPAR γ activator pyridaben was a strong inhibitor of adipogenesis. While uncommon, this result is not unprecedented; mycophenolic acid is a known PPAR γ activator that inhibits adipogenesis (Ubukata et al. 2007). There are two possible conclusions for the poor predictive power of the adipogenesis ToxPi. The first is that we did not identify an appropriate group of assays and that this resulted in the poor performance of the ToxPi. The second possibility is that the very poor correlation between receptor activation reported in ToxCast assays and bona fide receptor activation we measured resulted in poor predictive power of the adipogenesis ToxPi.

It is interesting to consider why the current ToxCast strategy is relatively effective for the androgen and estrogen receptors (Rotroff et al. 2014) but performs poorly for predicting PPAR γ

activity or obesogenicity. One possibility is that PPAR γ has a large binding pocket with relatively few high-affinity endogenous ligands, whereas ER and AR bind endogenous ligands at subnanomolar levels. For this reason, many PPAR γ activators might be identified by ToxCast without being biologically relevant. Another possibility is that the much larger number of ER and AR assays in ToxCast limits the damage caused by a few poorly performing assays to the predictive power of the overall assay system. However, false positive problems were also recently observed in ToxCast AR and ER endocrine disruption and neurotoxicity assays (Silva et al. 2015). Another possibility is that Phase I data only considered AC₅₀ values. Currently, in ToxCast Phase II, a more sophisticated approach is implemented that incorporates measures of cytotoxicity and chemicals are assigned a so-called 'Z-score'. It is recommended that chemicals with Z-scores less than 3 should be removed, or at a minimum, flagged/filtered (U.S. EPA 2014a). When we apply these cytotoxicity measures to the ToxCast Phase I chemicals, most false negatives are removed, however, all true positives are also lost. A new collection of PPAR γ activators or adipogenic chemicals can be identified using Z-score corrections, but our data suggest that this new approach alone does not improve the ability of ToxPi models to predict adipogenic activity or PPAR γ activators.

Although we identified new obesogens from the ToxCast phase 1 dataset, the low validation rate of PPAR γ and RXR α activation assays, coupled with the poor predictive power of the adipogenesis ToxPi is troublesome and prevents these data from being used as effective predictors of adipogenic activity. It was recognized by early computer programmers that the quality of the output from any software is dependent on the quality of the input; reliable models cannot be produced from inaccurate data or results that cannot be reproduced. It is a *sine qua non* of high throughput screening that HTS assays alone cannot accurately predict the activity of

chemicals against any endpoint. In their seminal paper on high throughput screening assays, Inglese et al. note that “It is essential to view the primary HTS as the initial step of an integrated process” (Inglese et al. 2007). All HTS assay results need to be confirmed by counter-screens and secondary screens that reduce the number of false positives and false negatives (Hughes et al. 2012). To the best of our knowledge, ToxCast assays as currently practiced are not constructed in this manner and this limits their accuracy. Until the HTS assays reflect actual receptor activity, these data must be interpreted with caution. This is particularly important now that EPA is proposing to use ToxCast as a substitute for Endocrine Disruptor Screening Program Tier 1 assays (Browne et al. 2015).

Another issue is that the assays used in ToxCast were largely pre-existing commercial assays which were adopted from the philosophy and approach of the pharmaceutical industry. Assays for drug discovery are designed to identify only the strongest hits in large libraries of structurally-similar chemicals (millions or more) to limit the subsequent screening required to develop lead compounds for preclinical studies. This is philosophically the opposite of a proper chemical genomics approach to identify potential bad actors that should be selected for further scrutiny. Such assays would seek to identify EVERY chemical that activates a particular pathway in a statistically significant way and then rank these for further testing. The ability of ToxCast assays to predict *in vivo* toxicity is often evaluated by comparing the effects of a chemical in ToxCast with effects from guideline studies, *in vivo* (Rotroff et al. 2013). However, the endpoints in guideline studies are not always sensitive to chemical effects on the endocrine system (Zoeller et al. 2012); thus, limiting their utility as validators of ToxCast assays for endocrine activity.

Of particular concern with respect to ToxCast is the lack of agreement among assays on the same endpoint. Exacerbating this problem is that Attagene assays identified ~2-3x more chemicals than other assays on both PPAR γ and ER endpoints (Figure 8 of this paper, and Supplemental Figure 6, Rotroff et al. 2014). Attagene trans-FactorialTM assays and NCGC GeneBLAzer assays utilize well-understood and thoroughly tested GAL4 DNA-binding domain, nuclear receptor ligand binding domain chimeras that perform well in a laboratory setting. Therefore, an open and important question is why such well-established principles of nuclear receptor biology have been implemented so poorly, yet are relied upon without further validation. Furthermore, the suitability of Attagene cis-FactorialTM assays to identify effects on specific gene-regulatory pathways is fundamentally questionable due to obvious off-target effects. For example, the PPAR response element used in the PPRE cis assay will bind PPAR δ and PPAR α , RXR α,β,γ , COUP-TF α,β , and HNF4 (Nakshatri and Bhat-Nakshatri 1998) at a minimum. Therefore, the results of such an assay cannot reasonably be called PPAR-specific; one wonders why such assertions are accepted uncritically. Despite these weaknesses, ToxCast publications continually average results across all similar assays equally to create composite curves (Rotroff et al. 2014). The lack of correspondence between distinct, yet mechanistically similar or identical assays is ignored, and all assays are weighted equally (Rotroff et al. 2014). This may work for cases such as the estrogen receptor where the large number of assays reduces the negative impact of the poorly performing assays, but will necessarily fail when assay numbers are small (such as for PPAR γ , RXR, etc.). It would be beneficial for other measures of action on particular endpoints to be included.

Conclusions

We have identified several problems in the ToxCastTM Phase I and Phase II data that impair the ability of these assays to predict activity on PPAR γ , RXR and adipogenesis. We recommend eliminating the practice of averaging results across assays in favor of eliminating poorly performing assays. We recommend incorporating reliable counter screens and secondary screens to validate the results of primary HTS assays before these are used for prioritization of chemical lists for further testing or to inform regulatory testing. Such modifications would be very beneficial and could improve the performance of ToxCast such that it can be as useful as originally envisioned. It is time that ToxCast assays and approaches are modified such that they produce accurate results that can be validated in subsequent experiments by multiple laboratories at high frequency.

References

- Bevington PR, Robinson DK. 2003. Data Reduction and Error Analysis for the Physical Sciences. New York: McGraw-Hill Education.
- Browne P, Judson RS, Casey WM, Kleinstreuer NC, Thomas RS. 2015. Screening Chemicals for Estrogen Receptor Bioactivity Using a Computational Model. *Environ Sci Technol* 49(14): 8804-8814.
- Calkin AC, Tontonoz P. 2012. Transcriptional integration of metabolism by the nuclear sterol-activated receptors LXR and FXR. *Nat Rev Mol Cell Biol* 13(4): 213-224.
- Chamorro-Garcia R, Blumberg B. 2014. Transgenerational effects of obesogens and the obesity epidemic. *Curr Opin Pharmacol* 19: 153-158.
- Chamorro-Garcia R, Kirchner S, Li X, Janesick A, Casey SC, Chow C, et al. 2012. Bisphenol A diglycidyl ether induces adipogenic differentiation of multipotent stromal stem cells through a peroxisome proliferator-activated receptor gamma-independent mechanism. *Environ Health Perspect* 120(7): 984-989.
- Chamorro-Garcia R, Sahu M, Abbey RJ, Laude J, Pham N, Blumberg B. 2013. Transgenerational inheritance of increased fat depot size, stem cell reprogramming, and hepatic steatosis elicited by prenatal exposure to the obesogen tributyltin in mice. *Environ Health Perspect* 121(3): 359-366.
- Chuprina A, Lukin O, Demoiseaux R, Buzko A, Shivanyuk A. 2010. Drug- and lead-likeness, target class, and molecular diversity analysis of 7.9 million commercially available organic compounds provided by 29 suppliers. *J Chem Inf Model* 50(4): 470-479.
- Collins FS, Gray GM, Bucher JR. 2008. Toxicology. Transforming environmental health protection. *Science* 319(5865): 906-907.
- Dahlin JL, Nissink JW, Strasser JM, Francis S, Higgins L, Zhou H, et al. 2015. PAINS in the assay: chemical mechanisms of assay interference and promiscuous enzymatic inhibition observed during a sulfhydryl-scavenging HTS. *J Med Chem* 58(5): 2091-2113.
- Darlington GJ, Ross SE, MacDougald OA. 1998. The role of C/EBP genes in adipocyte differentiation. *J Biol Chem* 273(46): 30057-30060.
- Dix DJ, Houck KA, Martin MT, Richard AM, Setzer RW, Kavlock RJ. 2007. The ToxCast program for prioritizing toxicity testing of environmental chemicals. *Toxicol Sci* 95(1): 5-12.
- Filer D, Patisaul HB, Schug T, Reif D, Thayer K. 2014. Test driving ToxCast: endocrine profiling for 1858 chemicals included in phase II. *Curr Opin Pharmacol* 19: 145-152.

Forman BM, Goode E, Chen J, Oro AE, Bradley DJ, Perlmann T, et al. 1995. Identification of a nuclear receptor that is activated by farnesol metabolites. *Cell* 81(5): 687-693.

FQPA. 1996. Public Law. 104-170.

Grun F, Blumberg B. 2006. Environmental obesogens: organotins and endocrine disruption via nuclear receptor signaling. *Endocrinology* 147(6 Suppl): S50-55.

Grun F, Watanabe H, Zamanian Z, Maeda L, Arima K, Cubacha R, et al. 2006. Endocrine-disrupting organotin compounds are potent inducers of adipogenesis in vertebrates. *Mol Endocrinol* 20(9): 2141-2155.

Hughes M, Inglese J, Kurtz A, Andalibi A, Patton L, Austin C, et al. 2012. Early Drug Discovery and Development Guidelines: For Academic Researchers, Collaborators, and Start-up Companies. In: *Assay Guidance Manual* [Internet] (Sittampalam G, Coussens N, Nelson H, et al., eds). Bethesda: Eli Lilly & Company and the National Center for Advancing Translational Sciences. Available from: <http://www.ncbi.nlm.nih.gov/books/NBK92015>.

Hulsen T, de Vlieg J, Alkema W. 2008. BioVenn - a web application for the comparison and visualization of biological lists using area-proportional Venn diagrams. *BMC Genomics* 9: 488.

Inglese J, Johnson RL, Simeonov A, Xia M, Zheng W, Austin CP, et al. 2007. High-throughput screening assays for the identification of chemical probes. *Nat Chem Biol* 3(8): 466-479.

Jameson DM, Sawyer WH. 1995. Fluorescence Anisotropy Applied to Biomolecular Interactions. *Method Enzymol* 246: 283-300.

Janesick A, Blumberg B. 2011a. Endocrine disrupting chemicals and the developmental programming of adipogenesis and obesity. *Birth Defects Res C Embryo Today* 93(1): 34-50.

Janesick A, Blumberg B. 2011b. Minireview: PPARgamma as the target of obesogens. *J Steroid Biochem Mol Biol* 127(1-2): 4-8.

Janesick AS, Shioda T, Blumberg B. 2014. Transgenerational inheritance of prenatal obesogen exposure. *Mol Cell Endocrinol* 398(1-2): 31-35.

Jolley ME. 1981. Fluorescence Polarization Immunoassay. *Clin Chem* 27(6): 972-973.

Kanayama T, Kobayashi N, Mamiya S, Nakanishi T, Nishikawa J. 2005. Organotin compounds promote adipocyte differentiation as agonists of the peroxisome proliferator-activated receptor gamma/retinoid X receptor pathway. *Mol Pharmacol* 67(3): 766-774.

Kavlock R, Chandler K, Houck K, Hunter S, Judson R, Kleinstreuer N, et al. 2012. Update on EPA's ToxCast program: providing high throughput decision support tools for chemical risk management. *Chem Res Toxicol* 25(7): 1287-1302.

Kavlock RJ, Austin CP, Tice RR. 2009. Toxicity testing in the 21st century: implications for human health risk assessment. *Risk Anal* 29(4): 485-487; discussion 492-487.

Knight AW, Little S, Houck K, Dix D, Judson R, Richard A, et al. 2009. Evaluation of high-throughput genotoxicity assays used in profiling the US EPA ToxCast chemicals. *Regul Toxicol Pharmacol* 55(2): 188-199.

Knudsen TB, Houck KA, Sipes NS, Singh AV, Judson RS, Martin MT, et al. 2011. Activity profiles of 309 ToxCast chemicals evaluated across 292 biochemical targets. *Toxicology* 282(1-2): 1-15.

Li X, Pham HT, Janesick AS, Blumberg B. 2012. Triflumizole is an obesogen in mice that acts through peroxisome proliferator activated receptor gamma (PPARgamma). *Environ Health Perspect* 120(12): 1720-1726.

Livak KJ, Schmittgen TD. 2001. Analysis of relative gene expression data using real-time quantitative PCR and the 2(-Delta Delta C(T)) Method. *Methods* 25(4): 402-408.

Love JD, Gooch JT, Benko S, Li C, Nagy L, Chatterjee VK, et al. 2002. The structural basis for the specificity of retinoid-X receptor-selective agonists: new insights into the role of helix H12. *J Biol Chem* 277(13): 11385-11391.

Martin MT, Dix DJ, Judson RS, Kavlock RJ, Reif DM, Richard AM, et al. 2010. Impact of environmental chemicals on key transcription regulators and correlation to toxicity end points within EPA's ToxCast program. *Chem Res Toxicol* 23(3): 578-590.

Masuno H, Kidani T, Sekiya K, Sakayama K, Shiosaka T, Yamamoto H, et al. 2002. Bisphenol A in combination with insulin can accelerate the conversion of 3T3-L1 fibroblasts to adipocytes. *J Lipid Res* 43(5): 676-684.

Muir DC, Howard PH. 2006. Are there other persistent organic pollutants? A challenge for environmental chemists. *Environ Sci Technol* 40(23): 7157-7166.

Nakshatri H, Bhat-Nakshatri P. 1998. Multiple parameters determine the specificity of transcriptional response by nuclear receptors HNF-4, ARP-1, PPAR, RAR and RXR through common response elements. *Nucleic Acids Res* 26(10): 2491-2499.

Peckett AJ, Wright DC, Riddell MC. 2011. The effects of glucocorticoids on adipose tissue lipid metabolism. *Metabolism* 60(11): 1500-1510.

Raghow R, Yellaturu C, Deng X, Park EA, Elam MB. 2008. SREBPs: the crossroads of physiological and pathological lipid homeostasis. *Trends Endocrinol Metab* 19(2): 65-73.

Reif DM, Martin MT, Tan SW, Houck KA, Judson RS, Richard AM, et al. 2010. Endocrine profiling and prioritization of environmental chemicals using ToxCast data. *Environ Health Perspect* 118(12): 1714-1720.

Reif DM, Sypa M, Lock EF, Wright FA, Wilson A, Cathey T, et al. 2013. ToxPi GUI: an interactive visualization tool for transparent integration of data from diverse sources of evidence. *Bioinformatics* 29(3): 402-403.

Romanov S, Medvedev A, Gambarian M, Poltoratskaya N, Moeser M, Medvedeva L, et al. 2008. Homogeneous reporter system enables quantitative functional assessment of multiple transcription factors. *Nat Methods* 5(3): 253-260.

Rosen ED, MacDougald OA. 2006. Adipocyte differentiation from the inside out. *Nat Rev Mol Cell Biol* 7(12): 885-896.

Rotroff DM, Dix DJ, Houck KA, Knudsen TB, Martin MT, McLaurin KW, et al. 2013. Using in vitro high throughput screening assays to identify potential endocrine-disrupting chemicals. *Environ Health Perspect* 121(1): 7-14.

Rotroff DM, Martin MT, Dix DJ, Filer DL, Houck KA, Knudsen TB, et al. 2014. Predictive endocrine testing in the 21st century using in vitro assays of estrogen receptor signaling responses. *Environ Sci Technol* 48(15): 8706-8716.

SDWA Amendments. 1996. Public Law. 104-182.

Silva M, Pham N, Lewis C, Iyer S, Kwok E, Solomon G, et al. 2015. A Comparison of ToxCast Test Results with In Vivo and Other In Vitro Endpoints for Neuro, Endocrine, and Developmental Toxicities: A Case Study Using Endosulfan and Methidathion. *Birth Defects Res B Dev Reprod Toxicol* 104(2): 71-89.

Sweetnam PM, Caldwell L, Lancaster J, Bauer C, Mcmillan B, Kinnier WJ, et al. 1993. The Role of Receptor-Binding in Drug Discovery. *J Nat Prod* 56(4): 441-455.

Temkin AM, Bowers RR, Magaletta ME, Holshouser S, Maggi A, Ciana P, et al. 2015. Effects of Crude Oil/Dispersant Mixture and Dispersant Components on PPARgamma Activity and : Identification of Dioctyl Sodium Sulfosuccinate (DOSS; CAS #577-11-7) as a Probable Obesogen. *Environ Health Perspect*. DOI:10.1289/ehp.1409672

Thayer KA, Heindel JJ, Bucher JR, Gallo MA. 2012. Role of environmental chemicals in diabetes and obesity: a National Toxicology Program workshop review. *Environ Health Perspect* 120(6): 779-789.

Tontonoz P, Spiegelman BM. 2008. Fat and beyond: the diverse biology of PPARgamma. *Annu Rev Biochem* 77: 289-312.

TSCA. 1976. Public Law. 94–469.

U.S. EPA. 2014a. Integrated Bioactivity and Exposure Ranking: A Computational Approach for Prioritization and Screening Chemicals in the Endocrine Disruptor Screening Program. Washington, D.C. Available: <http://www.regulations.gov/#!documentDetail;D=EPA-HQ-OPP-2014-0614-0003>.

U.S. EPA. 2014b. TSCA Chemical Substance Inventory Basic Information. Washington, D.C. Available: <http://www.epa.gov/oppt/existingchemicals/pubs/tscainventory/basic.html>.

Ubukata M, Takamori H, Ohashi M, Mitsuhashi S, Yamashita K, Asada T, et al. 2007. Mycophenolic acid as a latent agonist of PPARgamma. *Bioorg Med Chem Lett* 17(17): 4767-4770.

Zlokarnik G, Negulescu PA, Knapp TE, Mere L, Burres N, Feng LX, et al. 1998. Quantitation of transcription and clonal selection of single living cells with beta-lactamase as reporter. *Science* 279(5347): 84-88.

Zoeller RT, Brown TR, Doan LL, Gore AC, Skakkebaek NE, Soto AM, et al. 2012. Endocrine-disrupting chemicals and public health protection: a statement of principles from The Endocrine Society. *Endocrinology* 153(9): 4097-4110.

Table 1. Summary of results from Figures 1 and 2.

ToxCast Chemical ^a	Adipogenesis ^b	Activation ^c	
		3T3-L1	COS7
Triphenyltin ^d	Positive	PPAR γ Activator	0.02
Fluazinam ^d	Not Tested	PPAR γ Antagonist	7.2
Niclosamide	Not Tested	Inactive	
Pyraclostrobin	Not Tested	Inactive	
Zoxamide	Positive	PPAR γ Activator	1.31
Acetochlor	Negative	PPAR γ Antagonist	46.7
Butachlor	Not Tested	Not Tested	
Triflumizole	Positive	PPAR γ Activator	11.5
Prochloraz	Not Tested	Inactive	
Spirodiclofen	Positive	PPAR γ Activator	12.76
Alachlor	Not Tested	PPAR γ Antagonist	4.48
Tebufenpyrad ^d	Not Tested	Inactive	
Dimethenamid	Not Tested	Inactive	
Tebufenozide	Not Tested	Inactive	
Quinoxifen ^e	Not Tested	Inactive	
Indoxacarb	Not Tested	Inactive	
Fenpyroximate (Z,E) ^d	Not Tested	Inactive	
S-Bioallethrin	Not Tested	Inactive	
Dimethomorph	Not Tested	Inactive	
Cyazofamid ^d	Not Tested	Inactive	
Chlorothalonil	Not Tested	Inactive	

^aList of the chemicals used in PPAR γ activation or antagonism assays (Figure 1). ^bResults of the 3T3-L1 adipogenesis assay. Only those chemicals that were positive activators on PPAR γ were tested, and all those tested were adipogenic. ^cResults of the Cos-7 transient transfection assays, with the AC₅₀ value (in μ M) listed. Supplemental Material, Table S6 has a continuation of this table where our AC₅₀ values are compared to commercial assays. ^dChemicals which were also tested in ToxPi assays (Table 2). ^eQuinoxifen was later shown to be active at 100 μ M (see Table 2).

Table 2. Summary of ToxPi results derived from Figures 3-7.

ToxPi Chemical^a	Prioritization^b	Adipogenesis^c	Activation^d	
Chemical Name	ToxPI Score	3T3-L1	COS7	AC₅₀
Tebupirimfos	HIGH	Positive	Inactive	
Prallethrin	HIGH	Negative	Inactive	
d-cis/trans Allethrin	HIGH	Negative	Inactive	
Fludioxonil ^f	HIGH	Positive	RXR α Activator	14.3
Cyazofamid ^e	HIGH	Negative	Inactive	
Flusilazole	HIGH	Positive	Inactive	
Fenthion	HIGH	Negative	Inactive	
Fenpyroximate (Z,E) ^e	HIGH	Negative	Inactive	
Forchlorfenuron	HIGH	Positive	Inactive	
Triphenyltin ^e	HIGH	Positive	PPAR γ Activator	0.02
Tebufenpyrad ^e	HIGH	Negative	Inactive	
Bisphenol A	MEDIUM	Positive	Inactive	
PFOS	MEDIUM	Negative	Inactive	
Quinoxifen ^{e,f}	MEDIUM	Positive	PPAR γ Activator	33.4
Imazalil	MEDIUM	Negative	Inactive	
Pyridaben	MEDIUM	Inhibitor	PPAR γ Activator	3.0
Fluazinam ^e	MEDIUM	Negative	PPAR γ Antagonist	7.2
Methylene dithiocyanate	NEGATIVE	Negative	Inactive	
Maleic hydrazide	NEGATIVE	Negative	Inactive	
Monocrotophos	NEGATIVE	Negative	Inactive	
Asulam	NEGATIVE	Negative	Inactive	
Flumetsulam	NEGATIVE	Negative	Inactive	
Acetamiprid	NEGATIVE	Positive	Inactive	
Pymetrozine	NEGATIVE	Positive	Inactive	

^aList of the chemicals used in PPAR γ or RXR α activation assays and adipogenesis assays (Figures 4-7). ^bToxPi score (Reif et al. 2010; Reif et al. 2013) (Figure 3). ^cResults of the 3T3-L1 adipogenesis assay. ^dCos-7 transient transfection assays, with the AC₅₀ value (in μ M) listed. Supplemental Material, Table S7 has a continuation of this table where our AC₅₀ values are compared to commercial assays. ^eChemicals which were also tested in ToxCast assays (Table 1). ^fThese chemicals promoted adipogenesis in mBMSCs in addition to 3T3-L1 cells.

Figure Legends

Figure 1. ToxCast Chemical Activity on PPAR γ . The ability of a graded dose series of ToxCast chemicals to (A) activate or (B) antagonize GAL4-mPPAR γ was tested in transiently transfected COS7 cells. (A, B) Data points are averages of triplicate transfections (3 biological replicates). Cytotoxicity, as measured by decreased β -galactosidase activity was observed at 100 μ M for spirodiclofen, triflumizole, alachlor and fluazinam, ≥ 10 μ M for zoxamide, and ≥ 1 μ M for triphenyltin. Data are depicted as (A) fold induction or (B) reduction over vehicle (0.05% DMSO) controls \pm S.E.M. (A) ToxCast chemicals were tested in 3-fold serial dilutions from 100 μ M through 0.137 μ M, with the final data point being 0.05% DMSO. Rosiglitazone serves as a positive control activator. (B) ToxCast chemicals were tested in 3-fold serial dilutions from 100 μ M, in competition with 50 nM rosiglitazone (Rosi). T0070907 (2-chloro-5-nitro-N-4-pyridinylbenzamide) serves as a positive control PPAR γ antagonist. (C) EC₅₀, EC₁₀, IC₅₀, and IC₁₀ values calculated from 1A, B are reported, and compared to commercial assays (see Supplemental Material, Figure S1). ATG = Attagene GAL-PPAR γ activation assay; NCGC = GeneBLAzer agonist (EC values) or antagonist (IC values) assays. Triphenyltin is previously published (Kanayama et al. 2005).

Figure 2. ToxCast chemicals zoxamide and spirodiclofen induce adipogenesis in 3T3-L1 cells and mBMSCs. Adipogenesis was induced in cells according to Supplemental Material, Figure S2. Lipid accumulation was quantified in differentiated (A) 3T3-L1 preadipocytes or (B) mouse bone marrow derived mesenchymal stem cells (mBMSCs) by measuring Nile Red fluorescence normalized by cell number (Hoechst). Rosiglitazone and tributyltin serve as positive control adipogenic chemicals. Gene expression was determined by the $2^{-\Delta\Delta CT}$ method using 36b4 as the reference gene. Data are reported as fold induction over 0.1% DMSO vehicle controls \pm S.E.M using standard propagation of error. Primer sequences can be found in Supplemental Material, Table S4. One-way ANOVA was conducted for zoxamide and spirodiclofen treatment groups and DMSO vehicle, followed by Dunnett's post-hoc test: * $P \leq 0.05$, ** $P \leq 0.01$, *** $P \leq 0.001$ compared to vehicle. Unpaired t-test was conducted for the positive controls rosiglitazone, tributyltin versus vehicle: # $P \leq 0.05$, ## $P \leq 0.01$, ### $P \leq 0.001$.

Figure 3. Selection of ToxPi Chemicals for Adipogenesis Assays. (A) Adipogenesis ToxPi where slice size (magnitude) represents the activity of a ToxCast chemical in a particular assay or collection of assays (See the assays which comprise each slice in Supplemental Material, Table S1 and the AC_{50} values associated with these assays in Supplemental Material, Table S3). PPRE = Attagene cis-PPRE reporter gene assay; PPAR γ = Attagene and NCGC trans-PPAR γ reporter gene assay and Novascreen hPPAR γ direct binding assay; GR = Attagene cis-GRE, trans-GR, and NCGC trans-GR reporter gene assay, and Novascreen hGR direct binding assay; LXR = Attagene trans-LXR α , trans- LXR β and NCGC trans-LXR β reporter gene assay; LXRE = Attagene cis-LXRE reporter gene assay; C/EBP = Attagene cis-C/EBP reporter gene assay; SREBP = Attagene cis-SREBP reporter gene assay; RXR α = Attagene and NCGC trans-RXR α reporter gene assay. Highest scoring ToxPi chemicals are predicted to be obesogenic. **(B)** Plot of the ToxPi scores for all Phase I ToxCast chemicals. Red data points are selected high-scoring chemicals. Blue data points are selected medium-scoring chemicals. Grey data points are selected zero-scoring chemicals. Black open circles are chemicals not tested in our adipogenesis assays.

Figure 4. ToxPi Chemical Activity on PPAR γ and RXR α . The ability of a graded dose series of ToxPi chemicals to activate **(A)** GAL4-mPPAR γ or **(B)** GAL4-hRXR α was tested in transiently transfected COS7 cells. **(A, B)** Data points are averages of triplicate transfections (3 biological replicates). Cytotoxicity, as measured by decreased β -galactosidase activity was observed at 1 μ M for triphenyltin. ToxPi chemicals were tested in 3-fold serial dilutions from 100 μ M through 0.137 μ M, with the final data point being 0.05% DMSO. Data are depicted as fold induction over vehicle (0.05% DMSO) controls \pm S.E.M. **(A)** Rosiglitazone serves as a positive control activator for GAL4-mPPAR γ . **(B)** LG100268 (2-[1-(3,5,5,8,8-Pentamethyl-5,6,7,8-tetrahydro-2-naphthyl)cyclopropyl]pyridine-5-carboxylic acid) serves as a positive control activator for GAL4-hRXR α . **(C)** EC_{50} and EC_{10} values calculated from **4A, B** are reported, and compared to other assays (see Supplemental Material, Figure S1). ATG = Attagene GAL-PPAR γ or GAL-RXR α activation assay; NCGC = GeneBLazer GAL-PPAR γ or GAL-RXR α activation assays. Triphenyltin is previously published (Kanayama et al. 2005).

Figure 5. ToxPi chemicals induce adipogenesis in 3T3-L1 preadipocytes. Adipogenesis was induced in 3T3-L1 cells according to Supplemental Material, Figure S2. 3T3-L1 cells were exposed to adipogenic cocktail for 2 days, then exposed to the test chemicals for 5 days. Differentiated cells were fixed and stained with Nile Red and Hoechst 33342. Lipid accumulation was quantified in cells by measuring Nile Red fluorescence normalized by cell number (Hoechst). Rosiglitazone and tributyltin serve as positive control adipogenic chemicals. One-way ANOVA was conducted for ToxPi chemical treatment groups and DMSO vehicle, followed by Dunnett's post-hoc test: * $P \leq 0.05$, ** $P \leq 0.01$, *** $P \leq 0.001$ compared to vehicle. Unpaired t-test was conducted for the positive controls rosiglitazone, tributyltin versus vehicle: #### $P \leq 0.001$.

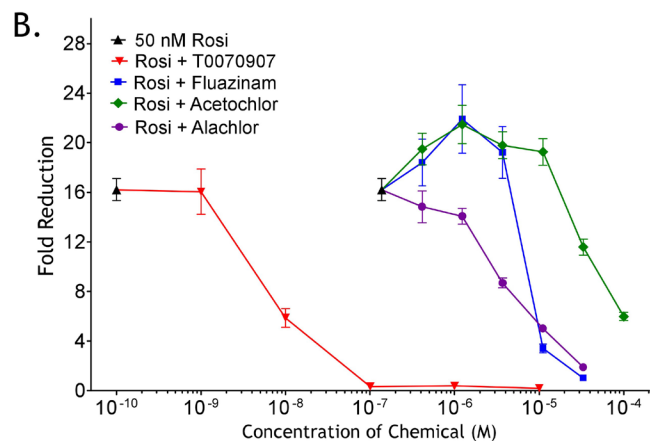
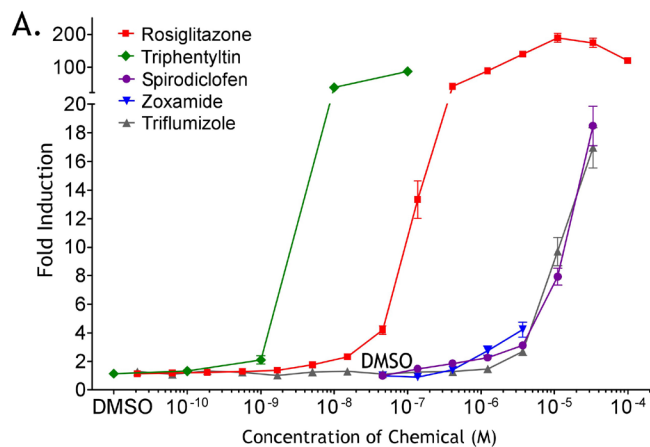
Figure 6. ToxPi chemicals induce adipogenic gene expression in 3T3-L1 preadipocytes. Adipogenesis was induced in 3T3-L1 cells according to Supplemental Material, Figure S2. 3T3-L1 cells were exposed to adipogenic cocktail for 2 days, then exposed to the test chemicals for 5 days. 3T3-L1 cells were homogenized in TriPure, total RNA was isolated, reverse transcribed, and QPCR was performed. Gene expression was determined by the $2^{-\Delta\Delta CT}$ method using 36b4 as the reference gene. Data are reported as fold induction over 0.1% DMSO vehicle controls \pm S.E.M using standard propagation of error. Primer sequences can be found in Supplemental Material, Table S4. One-way ANOVA was conducted for ToxPi treatment groups and DMSO vehicle, followed by Dunnett's post-hoc test: * $P \leq 0.05$, ** $P \leq 0.01$, *** $P \leq 0.001$ compared to vehicle. Unpaired t-test was conducted for the positive controls ROSI, TBT versus vehicle: # $P \leq 0.05$, ## $P \leq 0.01$, ### $P \leq 0.001$.

Figure 7. ToxPi chemicals quinoxifen and fludioxonil induce adipogenesis in mBMSCs. Adipogenesis was induced in mouse bone marrow derived mesenchymal stem cells (mBMSCs) according to Supplemental Material, Figure S2. mBMSCs were exposed to adipogenic cocktail plus test chemicals or positive controls for 14 days. Differentiated cells were fixed and stained with Nile Red and Hoechst 33342. Lipid accumulation was quantified in differentiated cells by measuring Nile Red fluorescence normalized by cell number (Hoechst). Rosiglitazone (ROSI) and tributyltin (TBT) serve as positive control adipogenic chemicals. mBMSCs were homogenized in TriPure, total RNA was isolated, reverse transcribed, and QPCR was performed.

Gene expression was determined by the $2^{-\Delta\Delta CT}$ method using 36b4 as the reference gene. Data are reported as fold induction over 0.1% DMSO vehicle controls \pm S.E.M using standard propagation of error. Primer sequences can be found in Supplemental Material, Table S4. One-way ANOVA was conducted for ToxPi chemical treatment groups and DMSO vehicle, followed by Dunnett's post-hoc test: * $P \leq 0.05$, ** $P \leq 0.01$, *** $P \leq 0.001$ compared to vehicle. Unpaired t-test was conducted for the positive controls rosiglitazone, tributyltin versus vehicle: # $P \leq 0.05$, ## $P \leq 0.01$, ### $P \leq 0.001$.

Figure 8. Venn diagrams comparing three main nuclear receptor commercial assays employed in ToxCast. Phase II, release 2014 (Filer et al. 2014) assay datasets (gain AC_{50} values) were obtained for five nuclear receptors: PPAR γ , androgen receptor (AR), estrogen receptor (ER), farnesoid X receptor (FXR), and glucocorticoid receptor (GR). Three assays for each receptor were evaluated: Attagene (ATG) agonist assay (red), NovaScreen (NVS) direct binding assay (green), and NCGC/Tox21 GeneBLAzer agonist assay (blue). An additional diagram (top right) was created for PPAR γ using NCGC/Tox21 antagonism assay. Chemicals scoring $AC_{50} \leq 10 \mu\text{M}$ for each assay were incorporated in the Venn diagrams, created by BioVenn (Hulsen et al. 2008).

Figure 1



C.

PPAR γ Agonists			
Chemical	Spirodiclofen		
Assay	Fig 1A	ATG	NCGC
EC50	12.76 μ M	1.85 μ M	1.64 μ M
EC10	7.27 μ M	0.10 μ M	0.17 μ M
Chemical	Zoxamide		
Assay	Fig 1A	ATG	NCGC
EC50	1.31 μ M	1.27 μ M	NA
EC10	0.31 μ M	0.18 μ M	NA

PPAR γ Antagonists		
Chemical	Alachlor	
Assay	Fig 1B	NCGC
IC50	4.48 μ M	3.19 μ M
IC10	0.78 μ M	0.15 μ M
Chemical	Acetochlor	
Assay	Fig 1B	NCGC
IC50	46.7 μ M	4.50 μ M
IC10	9.54 μ M	0.42 μ M
Chemical	Fluazinam	
Assay	Fig 1B	NCGC
IC50	7.20 μ M	1.27 μ M
IC10	4.41 μ M	0.20 μ M

Figure 2

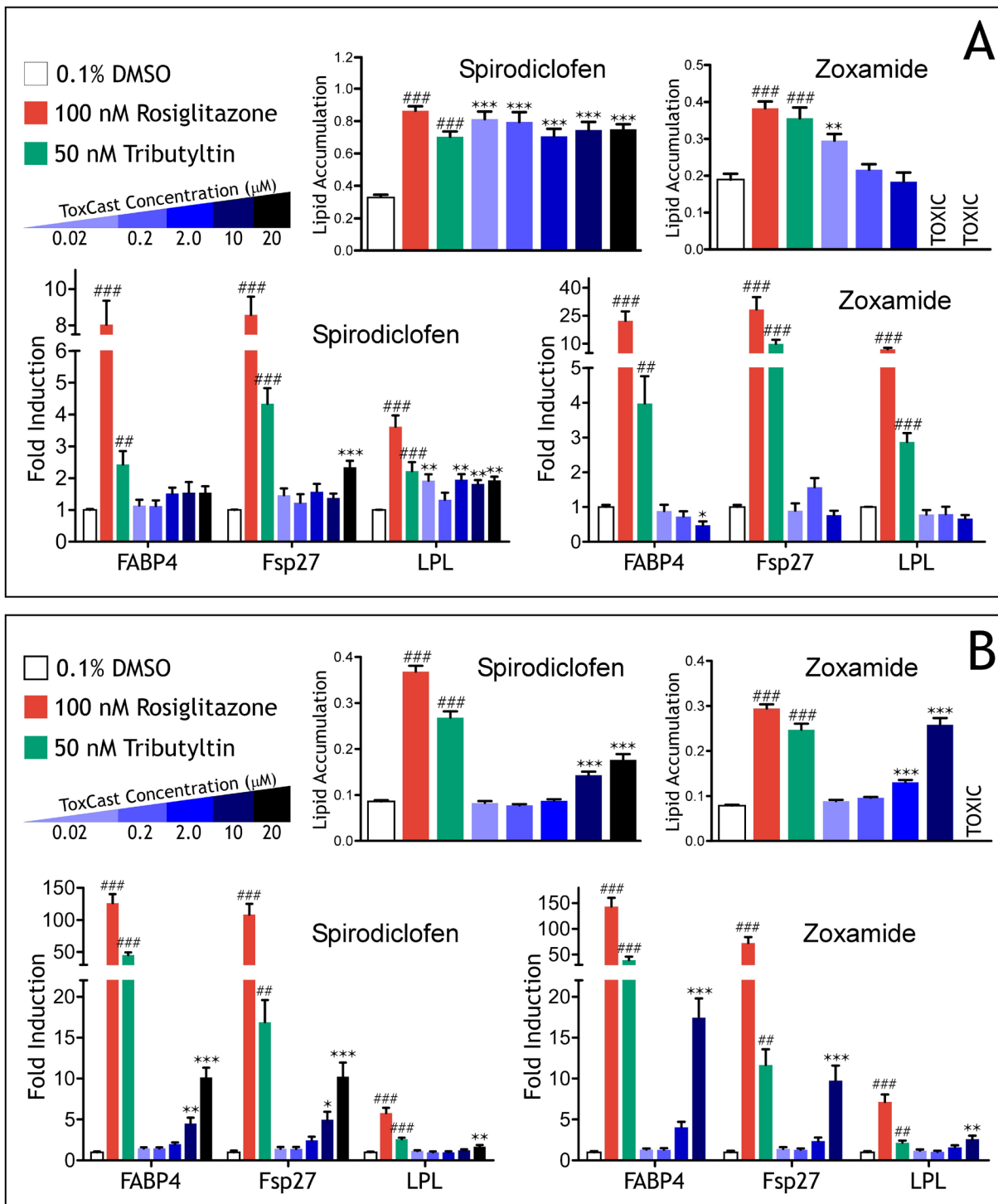


Figure 3

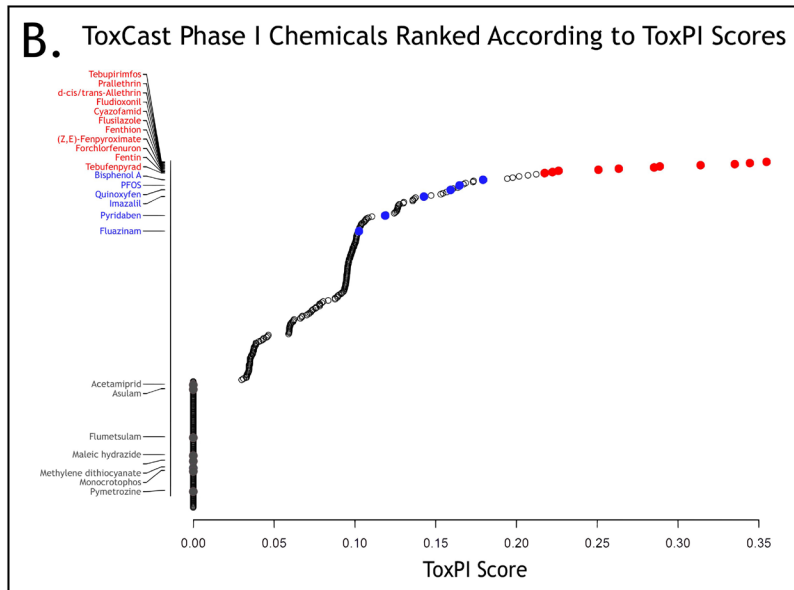
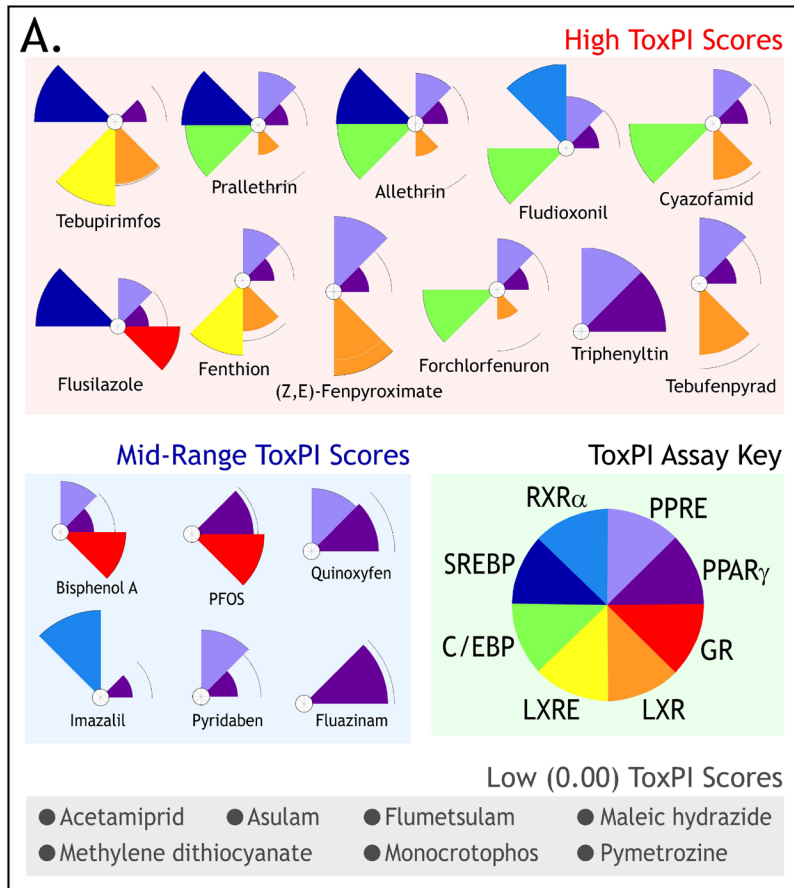
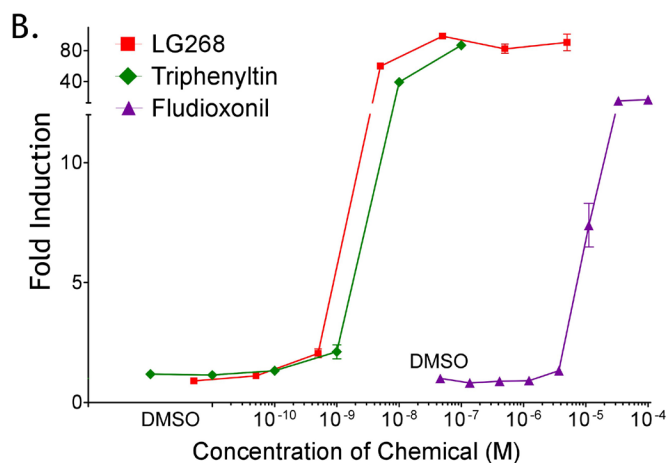
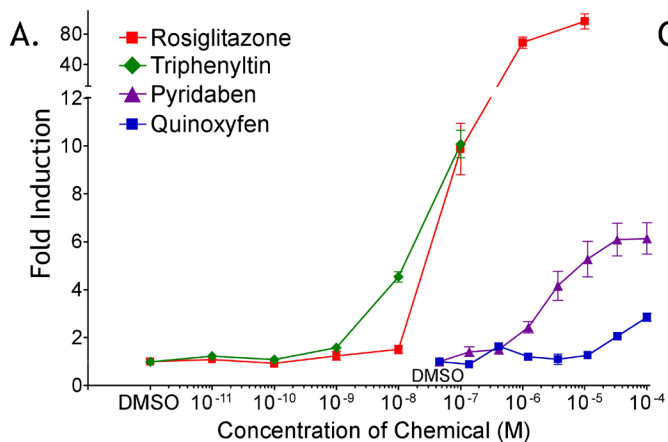


Figure 4



C.

PPAR γ Agonists			
Chemical	Pyridaben		
Assay	Fig 4A	ATG	NCGC
EC50	2.98 μ M	10.08 μ M	NA
EC10	0.53 μ M	2.57 μ M	NA
Chemical	Quinoxifen		
Assay	Fig 4A	ATG	NCGC
EC50	33.4 μ M	5.12 μ M	40.55 μ M
EC10	14.24 μ M	3.98 μ M	2.44 μ M

RXR α Agonist			
Chemical	Fludioxonil		
Assay	Fig 4B	ATG	NCGC
EC50	14.3 μ M	23.59 μ M	2.24 μ M
EC10	5.42 μ M	17.93 μ M	1.20 μ M

Figure 5

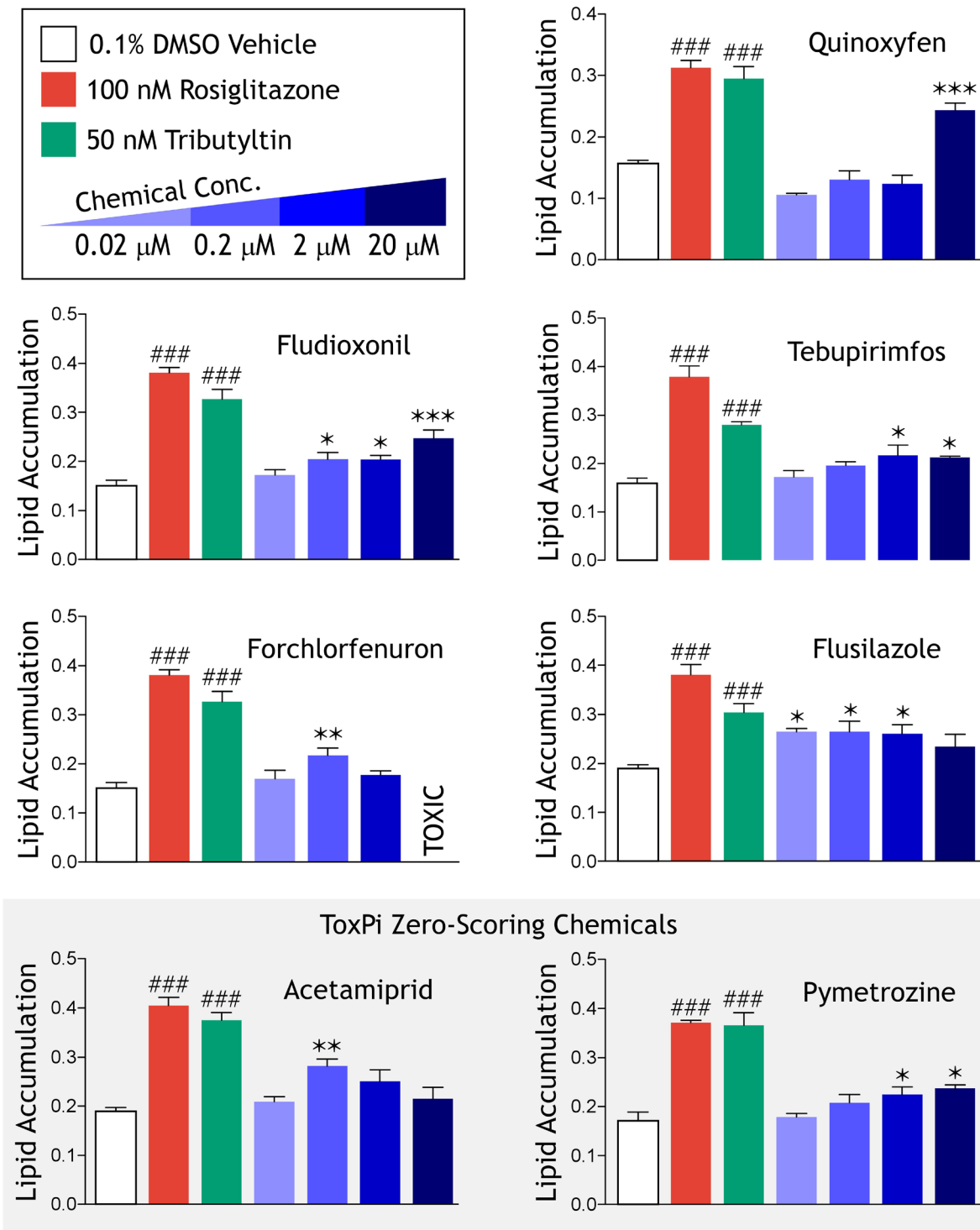


Figure 6.

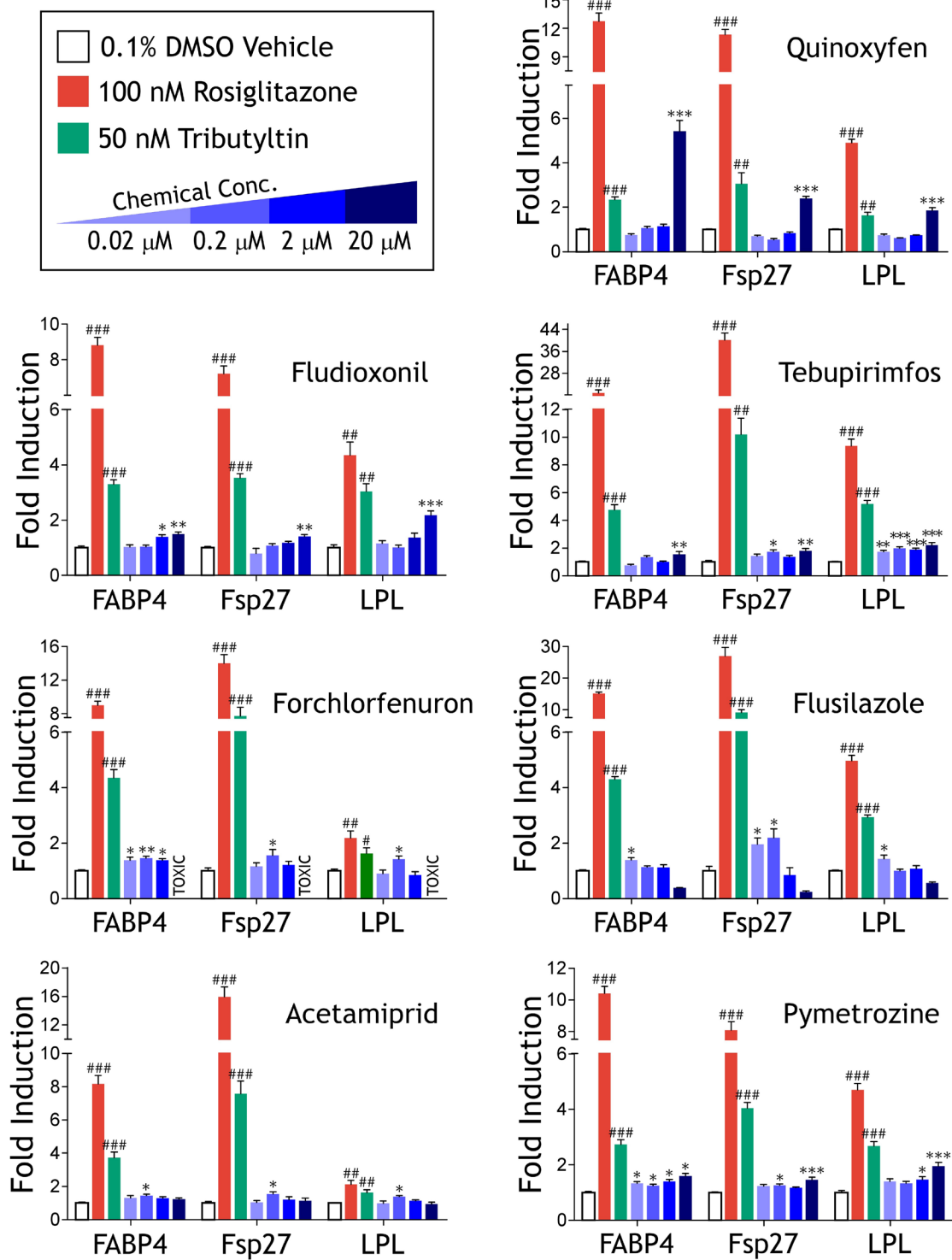


Figure 7

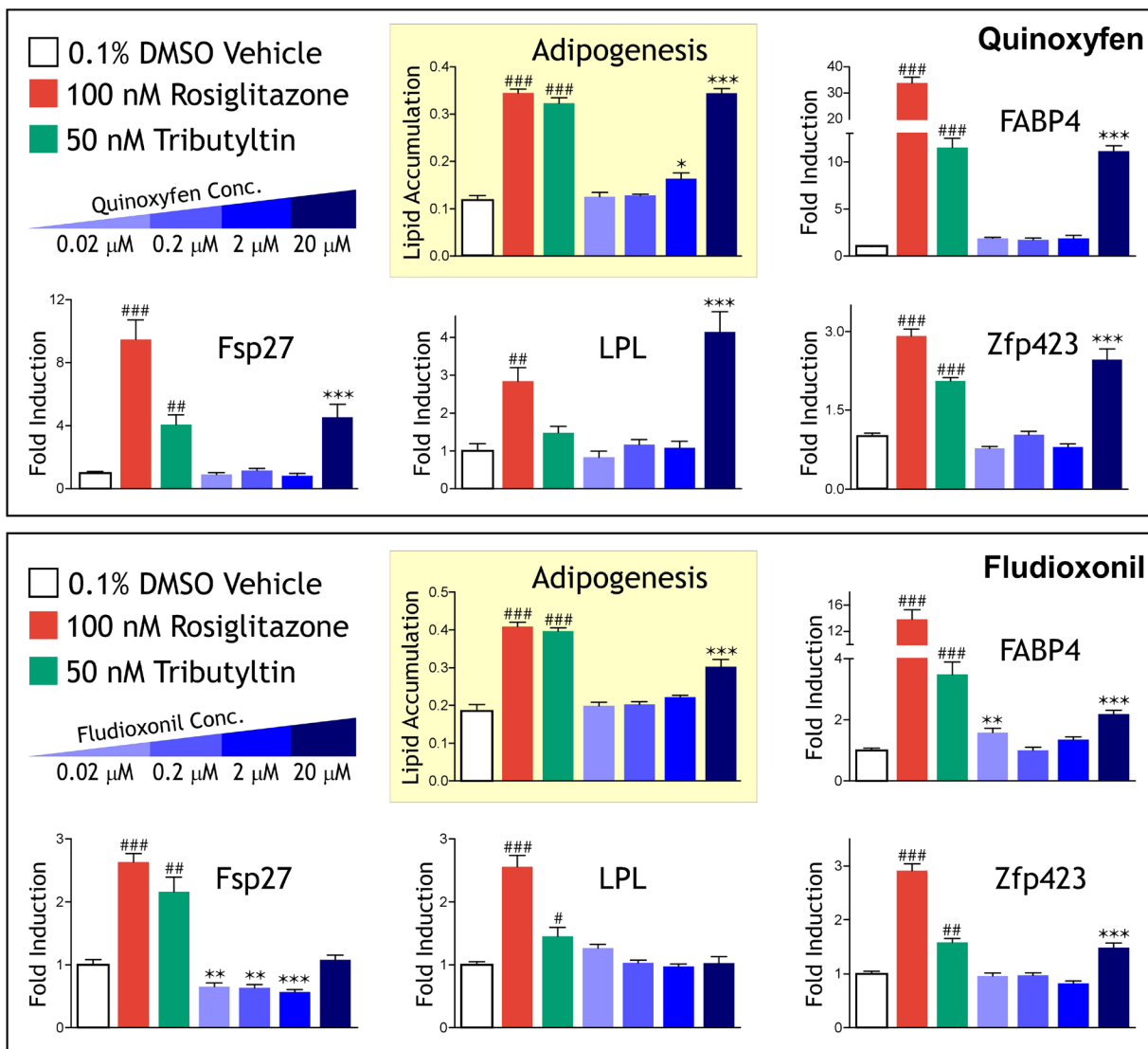


Figure 8

