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1 Title

Disentangling diverse responses to climate change among global marine ecosystem models
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- 54
- 55 Abstract

Climate change is warming the ocean and impacting lower trophic level (LTL) organisms. 56 57 Marine ecosystem models can provide estimates of how these changes will propagate to larger animals and impact societal services such as fisheries, but at present these estimates 58 vary widely. A better understanding of what drives this inter-model variation will improve our 59 60 ability to project fisheries and other ecosystem services into the future, while also helping to 61 identify uncertainties in process understanding. Here, we explore the mechanisms that 62 underlie the diversity of responses to changes in temperature and LTLs in eight global marine ecosystem models from the Fisheries and Marine Ecosystem Model Intercomparison Project 63 (FishMIP). Temperature and LTL impacts on total consumer biomass and ecosystem structure 64 65 (defined as the relative change of small and large organism biomass) were isolated using a 66 comparative experimental protocol. Total model biomass varied between -35% to +3% in response to warming, and -17% to +15% in response to LTL changes. There was little 67 consensus about the spatial redistribution of biomass or changes in the balance between 68 69 small and large organisms (ecosystem structure) in response to warming, and LTL impacts on 70 total consumer biomass varied depending on the choice of LTL forcing terms. Overall, climate 71 change impacts on consumer biomass and ecosystem structure are well approximated by the 72 sum of temperature and LTL impacts, indicating an absence of nonlinear interaction between

the models' drivers. Our results highlight a lack of theoretical clarity about how to represent fundamental ecological mechanisms, most importantly how temperature impacts scale from individual to ecosystem level, and the need to better understand the two-way coupling between LTL organisms and consumers. We finish by identifying future research needs to strengthen global marine ecosystem modelling and improve projections of climate change impacts.

- 79
- 80 Keywords

81 Climatic change, modelling, fishery oceanography, marine ecology, FishMIP, structural

- 82 uncertainty
- 83
- 84

85 **1. Introduction**

86 Water temperature and primary production play critical roles in marine processes. Higher temperatures accelerate reaction rates, with consequences ranging from the molecular to 87 ecosystem scale, while primary production provides the fundamental source of energy for 88 89 almost all marine life (Brown et al. 2004; Chavez et al. 2011). Climate change impacts on both water temperature and primary production will thus alter marine ecosystems in fundamental 90 ways (Pörtner et al. 2014). For example, a first-order expectation of these impacts is that 91 92 accelerated metabolic rates will consume energy more quickly in a warmer ocean, all else 93 being equal, so that less biomass could be supported by a given level of primary production (Heneghan et al, 2019). Yet, ecosystem-level effects emerge from individual-level processes 94 and interactions, which could lead to nonlinear effects and changes in ecosystem structure, 95 96 while shifting thermal habitats may influence the distribution of species, transforming food-97 webs to previously unknown states (Coll et al. 2020; Pinsky et al. 2020; Poloczanska et al. 2016). 98

99

100 There is a growing need to quantify and project climate change impacts on marine ecosystems to motivate mitigation (Bryndum-Buchholz et al. 2020), provide insight into potential future 101 102 threats to food security (Barange et al. 2014; Blanchard et al. 2017a; Boyce et al. 2020), and 103 identify needs for biodiversity conservation (Brito-Morales et al. 2020; Waldron et al. 2020). 104 Thus, there has been a recent proliferation of spatially-explicit marine ecosystem models that 105 simulate higher trophic level biomass and ecosystem structure at regional and global scales, 106 driven by output from climate-ocean-biogeochemical models (Tittensor et al. 2018). These 107 ecosystem models differ significantly in their design, level of complexity and implementation, 108 reflecting different choices for how to represent fundamental marine ecosystem processes,

109 as well as a diversity of model purpose and scope. As a result, there is considerable 110 uncertainty in model projections of climate change impacts on higher trophic levels (e.g. Lotze 111 et al., 2019), with projections from each model dependent upon decisions around the inclusion or simplification of many candidate processes. Structural diversity in model 112 113 projections is a strength for gaining a rich view of possible outcomes, given that each model 114 reflects a different subset of established physiological and process knowledge, implemented using different mathematical representations (Knutti, 2010; Brander et al. 2013; Lefevre et al. 115 116 2017; Payne et al. 2016). At the same time, this diversity reflects fundamental uncertainty in 117 our understanding of processes. Thus, identifying sources of structural uncertainty in ensemble projections can point to critical weaknesses and thereby accelerate model 118 119 improvement.

120

121 The Fisheries and Marine Ecosystem Model Intercomparison Project (FishMIP) was created to 122 explore this uncertainty and provide more robust assessments of climate impacts on marine ecosystems through the analysis of multi-model ensembles (Tittensor et al. 2018). A recent 123 124 FishMIP study (Lotze et al. 2019) found that projections of mean changes in animal biomass 125 from a model ensemble typically compared better with empirical data than individual models, 126 emphasising the benefits of ensemble climate impact projections. However, uncertainty in 127 ensemble projections of higher trophic level biomass is significant: Lotze et al. (2019) found that the spread of changes across the FishMIP ensemble in 21st century marine consumer 128 129 biomass under the high emissions, representative concentration pathway 8.5 (RCP 8.5) 130 climate change scenario (0 to -35%) was larger than the multi-model mean consumer biomass 131 change between the RCP 2.6 (low emissions) and RCP 8.5 scenarios (-5% to -20%). This means 132 that structural uncertainty across global marine models is greater than climate scenario

uncertainty, which is problematic for the goal of using these models to provide assessmentsof climate impacts on marine ecosystems and the societal services they provide.

135

136 For all global models in the FishMIP ensemble, temperature and lower trophic level (LTL) 137 forcings such as net primary production, phytoplankton and zooplankton biomass, and export carbon are the two main drivers of projected climate change impacts (Tittensor et al. 2018), 138 yet their implementations vary. Although there is some agreement on how temperature 139 140 impacts physiological processes in general (e.g. Kooijman, 2010), there is less agreement on 141 how these impacts vary across functional groups, body sizes, and different processes such as growth and metabolism (van Denderen et al. 2020). Similarly, although it is universally 142 understood that LTL biomass and production provide the source of energy that supports 143 144 higher trophic levels, there is less understanding about how the physiology and structure of 145 LTLs affects transfer efficiency and ecosystem structure, and how to couple lower and higher trophic levels (Eddy et al. 2020; Heneghan et al. 2016; Stock et al. 2017). Previous multi-model 146 147 ensemble studies have explored structural model uncertainty in projections of consumer 148 biomass and species distribution shifts under climate change (e.g. Jones et al. 2012; Woodworth-Jefcoats et al. 2015), but these studies did not disentangle the effects of 149 150 temperature and lower trophic level (LTL) changes, a strategy that can provide mechanistic insight on underlying processes (Carozza et al. 2018). 151

152

Here, we identify sources of structural uncertainty in marine ecosystem models, by disentangling the effects of temperature and LTL changes on model projections using eight global models from the FishMIP ensemble. We first summarise how temperature and LTL processes are incorporated in these models, highlighting common representations and

differences across the ensemble. We then isolate the impact of changes in temperature and LTL processes on consumer biomass and ecosystem structure (which we define as the relative change in small <30 cm and large \geq 30 cm consumer biomass) in a simulation protocol involving a combination of pre-industrial, historical and RCP 8.5 forcings. By illuminating key sources of structural uncertainty in marine model projections, we identify critical areas of future research necessary to improve not only climate impact projections but also our understanding of the marine ecosystem.

164

165 **2. Methods**

We used projections from eight marine ecosystem models from the Fisheries and marine 166 ecosystem Model Intercomparison Project (FishMIP, www.fishmip.org; Tittensor et al. 2018). 167 168 There are several model types (see Tables 1 and 2 for a summary of each model and key 169 references). First, models that draw on the strongly size-structured nature of marine 170 ecosystem processes to represent the ecosystem purely by body size (BOATS, 171 Macroecological) or trophic level (EcoTroph). Second, trait-based size-structured models 172 (APECOSM, DBPM, FEISTY, ZooMSS), which move beyond a purely size-based representation to include different communities and groups using functional traits other than body size. Last, 173 174 DBEM is a habitat suitability-based species-distribution model that resolves the biomass and spatial distribution of >1200 fish and invertebrate species using observational data, and 175 includes other mechanisms such as species ecophysiology and dispersal. There is large 176 177 variation in the structural complexity of the models, and a detailed description of how each model incorporates temperature and lower trophic level (LTL) impacts, including relevant 178 equations and temperature parameters, can be found in the Supplementary Information S2. 179 180 Here we summarise the key similarities and differences of each model as they pertain to

temperature, LTLs and other drivers in Sections 2.1, 2.2, 2.3 and Table 1 and 2. We thenexplain the experimental protocol and model outputs in Section 2.4.

183

184 **2.1.** How do models incorporate temperature impacts?

185 Across all models, individuals gain mass through anabolic processes such as food uptake and assimilation, while they lose mass through catabolic processes such as respiration. 186 Populations can also gain individuals through reproduction, and lose individuals through 187 188 mortality (Table 2). These processes are all influenced by temperature. As a result, changes 189 in ecosystem structure depend on how models resolve: (i) temperature effects on individual anabolic and catabolic processes across different functional groups, body sizes or trophic 190 levels; and (ii) how these variations drive changes in ecological interactions (Table 2). 191 192 Temperature effects on these processes are represented in all models as an exponential 193 scaling, with parameters varying widely between models (Supplementary Information S2). 194 However, within models the same temperature scaling parameters are used across all 195 functional groups and ecosystem components, excluding EcoTroph, which uses different 196 scalings depending on the ecosystem's biome.

197

The representation of anabolic and catabolic processes varies across models (Table 2). Macroecological and EcoTroph have the simplest representations, with individual mass changes resolved implicitly in each model by a single individual metabolic rate that scales with temperature and body size (for Macroecological) or trophic level (for EcoTroph). For these models, total biomass at a given body size/trophic level is determined by the metabolic carrying capacity of that size/trophic level, divided by the metabolic rate of individuals. In these two models, individual metabolic rates increase with temperature while total metabolic

205 carrying capacity at a given body size/trophic level is determined by net primary production. 206 Thus, as warming drives an increase in individual metabolism, total biomass decreases even 207 if primary production remains constant. The BOATS model uses a similar framework to 208 Macroecological and EcoTroph to determine maximum supported biomass at each body size 209 class. However, in BOATS individual mortality is resolved separately and the growth of individuals from one size class to the next is explicitly resolved. As temperatures rise, 210 211 individual growth rates in BOATS increase, increasing the speed of biomass flow from small 212 to large size classes, but also increasing mortality and reducing the maximum biomass that 213 can be supported at each body size. Taken together, these processes mean that warming causes total biomass to decrease in BOATS. 214

215

216 Within BOATS, Macroecological and EcoTroph, ecological interactions such as predator-prey 217 encounters or predator-predator competition are not explicitly resolved. Thus, temperature 218 and LTL drivers do not explicitly change interactions among individuals. However, in BOATS 219 and Macroecological, all primary producers are represented by a single body size, which is 220 inversely related to temperature; as temperature increases, the single representative body 221 size of primary producers decreases according to an empirical equation. This in turn decreases 222 the production of higher trophic level organisms, as the number of trophic steps that net 223 primary production must be transferred through to reach any given body size increases. However, since trophic transfer efficiency in these models is not temperature-dependent, an 224 225 increase in the number of trophic levels is not expected to change the ratio of small and large 226 organism biomass. In contrast, transfer efficiency decreases with warming in Ecotroph. This 227 means that warmer waters in Ecotroph will support relatively less biomass at high trophic 228 levels (large body sizes) than what they will at low trophic levels (small body sizes).

229

230 For trait-based models (APECOSM, DBPM, FEISTY and ZooMSS), individual growth is fuelled 231 by ingesting smaller organisms, with individual ingestion rates scaling with temperature and body size. For APECOSM, DBPM and FEISTY this scaling is also modulated with the density of 232 233 prey. Thus, food uptake for individuals at one size is fuelled by predation of smaller size 234 individuals, and in some cases predators can compete with each other for the same prey. These models also include other sources of mortality (destruction of population biomass). 235 236 APECOSM, DBPM and ZooMSS incorporate at least one size-dependent mortality term, and 237 FEISTY includes a single natural mortality term that is independent of body size. These 238 additional mortality sources increase with temperature (except for senescence mortality, 239 which increases with body size, in DBPM and natural mortality, which is independent of body 240 size, in FEISTY), causing population biomass to decrease with increasing temperature. In 241 FEISTY, maintenance costs increase faster with both body size and temperature than do 242 ingestion rates. APECOSM and FEISTY also explicitly resolve size and temperature-dependent 243 costs of maintaining existing biomass (metabolism). In these models, as temperature 244 increases, maintenance costs also increase, reducing the available energy for growth and 245 reproduction. If maintenance costs of existing biomass exceed energy intake from ingestion, 246 biomass decreases. As food becomes limited in APECOSM, ingestion rates scale more slowly 247 with temperature than maintenance costs, limiting the scope for new growth and potentially 248 inducing biomass to decrease as maintenance costs outpace ingestion.

249

In APECOSM, DBPM, FEISTY and ZooMSS, temperature affects anabolic and catabolic processes differently across ecosystems, which has cascading effects on how the different components of ecosystems (e.g. predators and prey) interact. In APECOSM, FEISTY and

ZooMSS for example, the scaling of maintenance costs (in APECOSM and FEISTY) and senescence mortality (in APECOSM and ZooMSS) with body size and temperature mean that large organisms are more vulnerable to warming compared to small organisms. Everything else being equal, a warming-induced decrease in large organism biomass would reduce predation mortality on smaller organisms, thus favouring small organisms in these models.

258

Unlike what happens in the size and trait-based models, anabolic and catabolic processes in 259 260 DBEM are not driven explicitly by net primary production or by the ingestion of smaller 261 organisms. Instead, individual mass increases in DBEM when anabolism exceeds catabolism, both of which are affected by temperature and other drivers (see Section 2.3). Similar to 262 APECOSM, FEISTY and BOATS, the explicit balance between anabolic and catabolic processes 263 264 drives an organism's scope for growth-if catabolism outpaces anabolism, an individual's mass 265 will decline. In DBEM, anabolism accelerates more slowly with warming compared to 266 catabolism. Thus, as waters warm, an organism's potential for growth becomes increasingly 267 limited, and their maximum size decreases.

268

Organisms do not interact in DBEM. Rather, temperature and other forcings drive the spatial distribution of species across the ocean, with species' relative abundance in a region changing with respect to temperature depending on their thermal preference, and the prevailing water temperature. Thus, as waters warm, ecosystem structure changes by individual organisms becoming smaller on average, and by different species shifting their spatial boundaries to follow their thermal preferences.

275

276 Finally, energy transfer from small to large organisms through size-based predation is not the 277 only way that different parts of the ecosystem interact; in APECOSM, BOATS, DBPM, DBEM 278 and FEISTY, energy moves from large to the smallest size classes through reproduction. In 279 these models, the flux of small organism biomass entering the population through 280 reproduction can increase or decrease, depending on the relative impacts of warming on large 281 organisms. In FEISTY for example, if large organisms are more adversely affected by warming than small organisms, the reproduction rate in larger size classes would also decline, leading 282 283 to less biomass overall.

284

285 2.2. How do models incorporate lower trophic level processes?

Net primary production sustains essentially all non-photosynthetic life in the oceans, and 286 287 limits the biomass of higher trophic levels (Ryther, 1969; Friedland et al. 2012). Solar energy 288 captured and organic matter synthesized by primary producers flow through food webs, 289 primarily by larger organisms preying on smaller organisms. FishMIP models focus on higher 290 trophic levels, so lower trophic level processes are driven by a range of Earth system model 291 forcings (Table 1). The role of lower trophic levels in setting the limits to growth for higher 292 trophic levels is represented across the eight FishMIP models in two ways. First, for BOATS, 293 DBEM, Macroecological and EcoTroph, net primary production is used to determine limits of consumer growth rates and total biomass according to trophic transfer functions. Second, in 294 295 the trait-based models (APECOSM, DBPM, FEISTY and ZooMSS), plankton biomass and export 296 production are consumed by the size classes or functional groups that feed on them. This 297 energy is then transferred to higher trophic levels through size-based predation. However, all 298 eight models considered here are one-way forced (run offline), so there is no feedback from 299 higher trophic levels to lower trophic level biomass or production. This means that for the

trait-based models, ingestion-fuelled growth of higher trophic level predators is not explicitly
 matched by predation mortality in the plankton.

302

303 The correlation of mean phytoplankton size with total primary production is an important 304 driver of ecosystem structure (Boyce et al. 2015). Phytoplankton are generally larger in more 305 productive waters (Barnes et al. 2011; Finkel et al. 2010). Given the size-structured nature of the marine ecosystem (Trebilco et al. 2013), smaller phytoplankton support longer food 306 307 chains, which are thought to support relatively less consumer biomass (Eddy et al. 2020; 308 Ryther, 1969). All models explicitly represent this phenomenon with the exception of 309 EcoTroph and DBEM. EcoTroph uses trophic level instead of body size to represent the marine ecosystem. In DBEM, changes in net primary production affect the carrying capacity of 310 311 modelled species disregarding the size of primary producers. In BOATS and Macroecological, 312 changes in food chain length are represented by a varying representative size of 313 phytoplankton, the size increasing with net primary production according to empirical equations. In DBPM and ZooMSS, the phytoplankton size-spectrum, which is the relationship 314 between primary producer abundance N and body size w, $N = aw^{b}$, is continuous, with the 315 316 intercept a and slope b set by phytoplankton biomass. In these two models, the plankton size-317 spectrum intercept is lower and the slope is steeper in less productive waters, meaning 318 relatively more small producers but less biomass overall. APECOSM and FEISTY use size-319 fractionated phytoplankton and zooplankton biomass inputs from earth system models to 320 directly set the biomass of small and large phytoplankton and zooplankton groups, with a 321 fixed size-spectrum slope assigned to each LTL group in APECOSM. APECOSM and FEISTY also 322 use export carbon to represent detrital flux across the entire water column (in APECOSM) or 323 to the seafloor to fuel the growth of benthic invertebrates (in FEISTY).

2.3. How do models incorporate other impacts?

All models in the FishMIP ensemble are driven solely by temperature and LTL drivers, with the exception of APECOSM and DBEM (Table 1). In these two models, movement of organisms between adjacent grid cells is resolved, so both models incorporate current speeds. Since APECOSM resolves the 3D density of animal biomass, the model also uses 3D photosynthetically active radiation to resolve water clarity and light penetration across the water column. Thus, in APECOSM areas with the highest consumer biomass are not necessarily regions with the highest LTL biomass, due to active and passive horizontal movements in response to temperature, light, food availability and the strength of currents. Both APECOSM and DBEM also incorporate oxygen concentration, which impacts anabolic processes; lower oxygen concentration reduces the scope for organism growth in both models, and thus reduces total biomass. DBEM also resolves the negative impacts of acidification on catabolic processes, by incorporating pH forcings. DBEM also uses salinity, sea ice and mixed layer depth forcings, alongside temperature, to establish the spatial extent of each of the >1200 fish and invertebrate species the model resolves.

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349 **Table 1** Summary of temperature, lower trophic level (LTL) and other drivers sourced from

as the ecosystem models, used by each model in the FishMIP ensemble, as well as the ecosystem

- 351 representation of each model. All drivers used by the models in this experiment had a monthly
- 352 temporal resolution.

				C
Model and key references	Temperature drivers	LTL drivers	Other drivers	Taxonomic scope
APECOSM Maury et al. 2007a; 2007b; Maury 2010; Maury and Poggiale, 2013	3D water temperature	3D small and large phytoplankton, 3D small and large zooplankton biomass*, 3D export carbon flux	3D oxygen concentration, 3D photosynthetically active radiation, 3D current velocities	All epipelagic, mesopelagic and migratory heterotrophic marine animals in the pelagic ecosystem between 15µg– 120kg.
BOATS Carozza et al. 2016; 2017	2D water temperature (averaged over top 75 m)	2D depth-integrated net primary production	NA	All commercial animal biomass from 10g–100kg.
DBEM Cheung et al. 2008; 2010; 2011; 2016	2D sea surface temperature	2D depth-integrated net primary production	2D surface and bottom oxygen concentration, salinity and pH, sea ice, mixed layer depth, 3D current velocities	>1200 fish and invertebrate species.
DBPM Blanchard et al. 2009; 2012	2D sea surface and bottom water temperature	2D depth-integrated small and large phytoplankton biomass	ΝΑ	All benthic and pelagic marine animals, weighing between 1mg and 1 tonne.
EcoTroph Gascuel and Pauly, 2009; du Pontavice et al. 2020	2D sea surface temperature	2D depth-integrated net primary production	NA	All marine animals with trophic level ≥ 2 .
FEISTY Petrik et al. 2019	2D upper pelagic (averaged over 100 m) and bottom water temperature	2D depth-integrated (top 100 m) small and large zooplankton biomass*, 2D export carbon flux to the sea floor	NA	Forage, large pelagic and demersal fish, as well as benthic invertebrates, between 1mg and 125kg.
Macroecological Jennings and Collingridge (2015)	2D sea surface temperature	2D depth-integrated net primary production	NA	All marine animals between 1mg and 1 tonne.
ZooMSS Heneghan et al. (2020)	2D sea surface temperature	2D sea surface phytoplankton biomass	NA	Nine zooplankton groups, from flagellates to jellyfish and all marine animals between 1mg and 10 tonnes.

* Where small and large zooplankton biomass are not provided by an earth system model (as is the case with CESM1-BGC, the earth system model used in this study) FishMIP splits total

355 zooplankton biomass using the fraction of total phytoplankton biomass from small and large

- 356 phytoplankton.
- 357
- 358

Table 2 Summary of temperature and lower trophic level impacts in the FishMIP model ensemble.

Model	Temperature effect	on:	Lower trophic level effect on:				
	Individual anabolic and catabolic processes	Ecosystem structure	Individual anabolic and catabolic processes	Ecosystem structure			
APECOSM	Ingestion and thus predatory mortality scale with temperature and vary with predator size as well as the density and size of prey. Assimilation, maintenance, and non-predation mortality rates also scale with temperature. Temperature effects are stronger where prey density is high. In food-limited areas, catabolic processes increase faster than anabolic processes, causing individual mass to decrease. In food-rich areas, catabolic and anabolic processes increase in the same proportion, accelerating life-cycles.	Growth and mortality rates increase with temperature. In food limited situations, this leads to less biomass, especially for large organisms. In prey-rich regions, temperature does not drive biomass down but drives a faster transfer toward large sizes causing an increase in large organisms and a decrease of small organisms due to top-down control.	Small and large plankton biomass is the primary food source of small consumer organisms. More plankton biomass increases satiation and maximizes individual growth and reproduction, thus driving increases in biomass.	More plankton biomass supports more ecosystem biomass and reduces the trophic amplification of food limitation with size. This leads to biomass increase of large organisms and the presence of larger species in the communities.			
BOATS	Warming drives higher individual growth and mortality rates, which reduces the maximum biomass that can be supported by a given level of primary production.	Phytoplankton size decreases with warming. Smaller phytoplankton means longer food chains causing biomass declines for all sizes.	Net primary production sets the limits to growth across all body size classes. Higher production means more biomass.	Phytoplankton size decreases with decreasing production. Smaller phytoplankton mean longer food chains causing biomass declines for all sizes.			
DBEM	Biomass creation occurs after catabolism is deducted from anabolism. Catabolism increases faster with warming than anabolism. Thus, biomass decreases with warming.	Catabolism increases with size faster than anabolism, so warming affects large species more and drives shifts in spatial distribution of species.	In all regions, net primary production is a key part of what sets the limits to maximum biomass across all higher trophic levels.	Lower net primary production means less consumer biomass can be supported.			
DBPM	Ingestion-driven growth, and mortality rates from predation and natural sources scale with temperature at the same rate. Thus, temperature effects largely balance, except in low food regions where natural mortality is relatively large and causes biomass to decrease.	Natural mortality costs scale with temperature but decrease with body size. Thus, warming increases mortality relatively more for small organisms compared to large, potentially causing their biomass to decrease faster.	Small and large phytoplankton biomass set the slope and intercept of the phytoplankton size- spectrum, which is the primary food source of small pelagic organisms. More phytoplankton means more biomass.	Relatively more small phytoplankton with less phytoplankton biomass, which reduces food for small organisms and increases food chain length. This should decrease overall biomass, especially for larger sizes, as senescence increases with size.			
EcoTroph	Warming drives higher individual turnover rates, and lower trophic transfer efficiency, which means fewer individuals can be supported, causing biomass to decrease.	Trophic transfer efficiency decreases with warming, causing higher trophic level biomass to decrease more than lower trophic level biomass.	Net primary production is a driver of total biomass across all trophic levels. Higher production means more biomass.	Lower net primary production means less biomass can be supported across all trophic levels.			
FEISTY	Maintenance costs, ingestion-driven growth, and mortality rates from predation scale with temperature. Maintenance costs increase faster with warming compared to ingestion, so warming reduces the scope for growth, causing biomass to decrease.	Maintenance costs increase faster than ingestion- driven growth with body size and temperature. Thus, warming will reduce the scope for large organism growth more than small organisms.	Zooplankton is food for all small consumers and medium pelagic consumers. Export production fuels benthic growth. More zooplankton biomass and export production mean more ecosystem biomass overall.	Less zooplankton biomass supports lower pelagic biomass, and more small zooplankton biomass may reduce large fish biomass due to an increase in the number of trophic steps between zooplankton and a narrower scope for growth than smaller sizes.			
Macroecological	Warming drives higher individual metabolic rates, which means fewer individuals can be supported by a given level of primary production, causing total biomass to decrease.	Phytoplankton size decreases with warming, lengthening food chains and reducing how much energy is transferred to higher trophic levels.	Net primary production is a key determinant of total biomass. Higher net primary production means more biomass.	Phytoplankton size decreases with decreasing production. Smaller phytoplankton support longer food chains, thus less biomass across all sizes.			
ZooMSS	Ingestion-driven growth and mortality rates from predation and senescence scale with temperature at the same rate. Thus, temperature effects largely balance, except where senescence mortality is large, causing biomass to decrease.	Warming negatively impacts large organisms more than small by increasing senescence. If large organism biomass declines more than small, small biomass will increase from reduced predation.	The phytoplankton spectrum—set by total phytoplankton biomass—is the main food of microzooplankton. More phytoplankton means more consumer biomass.	Less phytoplankton biomass means less food for small organisms, and relatively more small phytoplankton. Drives shifts in zooplankton composition, which stabilise food chain length.			

361 2.4. Experimental protocol

362 To isolate the impact of temperature and LTL processes on the FishMIP ensemble, we 363 conducted four simulations (Table 3) following the general approach of Carozza et al. (2018). 364 In each simulation, all models were forced with different combinations of temperature, LTL 365 and other (for APECOSM and DBEM) drivers from pre-industrial, historical and high emissions 366 scenarios (RCP 8.5; IPCC, 2014) from the CESM1-BGC earth system model (Moore et al. 2013). simulation submitted for the Coupled Model Intercomparison Project 5 (CMIP5; IPCC, 2014). 367 368 For RCP 8.5 in the CMIP5 multi-model ensemble, CESM-BGC is average in temperature 369 sensitivity and less than average in global mean NPP and export production decline (Bopp et al. 2013). All forcings were provided to modellers with a monthly temporal resolution. We do 370 not use a range (from low to high) of emission scenarios for the future, or source forcings 371 372 from multiple Earth system models, as our purpose here is to isolate sources of structural 373 uncertainty within the FishMIP model ensemble itself (Payne et al. 2016). Under the RCP 8.5 374 scenario, the CESM1-BGC model projects a global sea surface temperature increase, which is 375 particularly marked at high latitudes (Figure 1b); net primary production declines across most 376 of the tropics and mid-latitudes, but increases at high latitudes and in the eastern South 377 Pacific (Figure 1d); phytoplankton and zooplankton biomass declines across most of the 378 world's oceans, except in polar regions (Figure 1f, h). The mean change in sea surface temperature across the global ocean from 1950 to 2100 under historical (averaged over 1950-379 1960) and RCP 8.5 (averaged over 2090-2100) scenarios is +3.2°C, and for net primary 380 381 production, phytoplankton and zooplankton carbon the mean change was -14%, -8% and -21%, respectively. 382

383

384 To enable the model comparison, two standardized outputs - total consumer biomass (which 385 broadly includes all consumer with trophic level >1, see Table 1) and the biomass of large 386 consumers (≥30 cm; see Tittensor et al. 2018 for details) - were calculated from each ecosystem model. All models supplied both outputs, except DBEM which did not provide the 387 388 biomass of large consumers. Outputs were reported as depth integrated carbon biomass (g 389 m^{-2}) and aggregated to a spatial grid with a resolution of 1° on a monthly or annual time step, 390 depending on model capability. Owing to differences in model formulation total consumer 391 biomass varies widely amongst models, all else being equal (Tittensor et al. 2018). Since our 392 focus was not on explaining these differences in total biomass, but rather the differences in 393 the responses of the models to temperature and LTL changes, we compared model outputs using biomass change relative to biomass levels under the preindustrial control. Further, as 394 395 our focus was isolating impacts of temperature and LTL processes, simulations were run in 396 the absence of fishing.

397

Table 3 Summary of the experimental simulations and corresponding environmental driver
 combinations. Temperature: all temperature-related drivers (e.g., sea surface temperature);
 LTL: all lower trophic level drivers (e.g., phytoplankton biomass); Other: any drivers that are
 not related to temperature or lower trophic levels (e.g., pH). The abbreviations for forcings
 are: PI (blue) = pre-industrial control, H (yellow) = historical, RCP 8.5 (purple) = RCP 8.5.

	Simulation								
	Cor	itrol	Temperati	ure Change	LTL C	hange	All (Climate) Change		
<u>Drivers</u>	<u>1950-2005</u>	2006-2100	<u>1950-2005</u>	2006-2100	<u>1950-2005</u>	2006-2100	<u>1950-2005</u>	2006-2100	
Temperature	PI	PI	н	RCP 8.5	PI	PI	н	RCP 8.5	

LTL	PI	PI	PI	PI	н	RCP 8.5	н	RCP 8.5
Other	PI	PI	PI	PI	PI	PI	Н	RCP 8.5



Figure 1 Control (historical averaged over 1950-1960) forcing variables and the change in
those variables from climate change (RCP 8.5) from the CESM1-BGC earth system model; a,b)
Sea surface temperature, c,d) Net primary production, e,f) Phytoplankton carbon, g,h)
Zooplankton carbon. The change in each variable is measured as the mean over 2090-2100
under the RCP 8.5 scenario minus the mean over 1950-1960 (for sea surface temperature), or

the percentage change between the mean in 1950-1960 and 2090-2100 (for net primary

413 production, phytoplankton carbon and zooplankton carbon).

414

415

416 **3. Results**

417 **3.1. Global changes in total consumer biomass**

All models projected a decline of globally averaged consumer biomass in the Temperature 418 419 Change simulation, with the exception of APECOSM (Figure 2a). The spread of total global 420 consumer biomass change in response to warming ranged from around -35% for Macroecological and BOATS, to +3% for APECOSM by the end of the 21st century. EcoTroph 421 produced the third largest change after BOATS and Macroecological of around -13%. The 422 423 remaining four models (DBEM, DBPM, FEISTY, ZooMSS) simulated modest changes in global 424 consumer biomass of between -2% (FEISTY) to -7% (DBPM) in response to changes in 425 temperature alone.

426

427 The LTL Change simulation also showed globally averaged biomass decreases for most 428 models, except BOATS and Macroecological, which projected global biomass increases (Figure 429 2b). For these two models, the trajectory of global biomass change was switched in the LTL Change simulation from negative change to positive in comparison with the warming only 430 simulation. In contrast, APECOSM projected global consumer biomass to increase slightly with 431 432 warming, but decrease with LTL changes. APECOSM projected a 7% decrease in total 433 consumer biomass globally, while BOATS and Macroecological projected increases of 10-15% 434 in response to LTL changes in isolation. Maximum decreases of biomass in LTL simulations are 435 half the magnitude (up to 15%) of the decreases in warming simulations. The smallest

response to LTL changes was from EcoTroph, which projected a total consumer biomass
change of <-1%. Trends in total consumer biomass from the other five models (DBEM, DBPM,
FEISTY and ZooMSS) were grouped within a range between -5% (DBEM) and -15% (ZooMSS).

439

440 The combined temperature and LTL changes led to a decline in globally averaged consumer 441 biomass across all models except DBEM (Figure 2c). By the end of the 21st century, changes 442 in global consumer biomass in the All (Climate) Change simulation ranged from around -30% 443 for BOATS and Macroecological, to +3% for DBEM. The other five models (APECOSM, DBPM, 444 EcoTroph, FEISTY and ZooMSS) had changes in total global consumer biomass of between -445 5% (for APECOSM) and -17% (for ZooMSS). For all models except BOATS, DBEM and 446 Macroecological, climate change impacts at the global scale were largely the sum of the 447 separate global impacts of warming and LTL change, with almost no non-additive impact 448 (Figure 2d). For BOATS and Macroecological, climate change impacts caused total consumer 449 biomass to decline by about 4% more than the sum of separate warming and LTL impacts. In 450 DBEM, total consumer biomass under climate change was ~15% higher than under the 451 combined, separate impacts of warming and LTL impacts, indicating some non-additive 452 impact of cumulative temperature and LTL changes. Non-additive impacts in DBEM may also 453 be caused by additional impacts from changes in pH and oxygen levels. APECOSM, the only 454 other model to incorporate non-temperature or LTL drivers, had negligible non-additive 455 impacts, indicating these other drivers had little effect compared to warming and LTL shifts.

456



Figure 2 Model projections of percentage change in global consumer biomass, relative to the Control, from 1950-2100 for the: a) Temperature Change simulation, b) Lower Trophic Level (LTL) Change simulation, c) All (Climate) Change simulation and d) the non-additive impacts of temperature and LTL changes, calculated by taking the difference between the All Change and the sum of the Temperature and LTL Change simulations.

464

458

465 3.2. Spatial changes in total consumer biomass

Globally averaged time-series of total consumer biomass change conceal considerable spatial
variation across regions within each model, and between models in each experiment.
Temperature-induced shifts in the spatial distribution of total consumer biomass (Figure 3,

469 left column) varied from increases in many regions for APECOSM, to decreases across the 470 global ocean in DBPM, ZooMSS, EcoTroph, BOATS and Macroecological. The magnitude of the 471 total consumer biomass changes generally followed the magnitude of change in temperature 472 (Figure 1b); temperate regions that experienced the strongest warming (Figure 1b) exhibited 473 the largest decreases in biomass for these five models. FEISTY and ZooMSS consumer biomass 474 also decreased with increased temperature in many of the regions with the greatest warming. 475 However, in warm regions (Figure 1a) with relatively small temperature increases such as the 476 eastern Pacific or northern Indian Ocean, FEISTY consumer biomass increased, and small 477 increases in ZooMSS consumer biomass occurred almost entirely in very high latitude polar regions where temperature change was relatively small (Figure 1b). In contrast, APECOSM 478 consumer biomass increased across most of the global ocean in response to warming. The 479 480 exception to this pattern was in patches where phytoplankton biomass was highest (Figure 481 1c) such as the North Atlantic, the Bering Strait or the South Pacific around New Zealand. In 482 DBEM, temperature-induced changes in consumer biomass were greatest in the warmest 483 waters around the equator, where DBEM consumer biomass decreased by 60-100%. In cold high latitude waters, DBEM consumer biomass increased by \geq 60% in response to warming. 484

485

For all models, lower trophic level (LTL) induced shifts in the distribution of consumer biomass (Figure 3, centre column) show more agreement in their patterns of change; most models show biomass decreases in equatorial regions, and increases towards the poles. The exceptions here are APECOSM, FEISTY and ZooMSS which show a mix of positive and negative consumer biomass toward the north pole. Consumer biomass shifts generally followed changes in the distribution of the main LTL forcings used by each model (Figure 1d, f, h). APECOSM, DBPM, FEISTY and ZooMSS use plankton biomass inputs (Table 2), and for these

models, consumer biomass generally decreased with decreasing phytoplankton carbon
(Figure 1f) and increases were isolated to polar regions. DBEM, EcoTroph, BOATS and
Macroecological use net primary production as their LTL forcing and the spatial distribution
of changes in consumer biomass followed spatial shifts in net primary production (Figure 1d),
with increases in biomass not only in polar regions, but also in the North Pacific and in the
South East Pacific.

499

When both temperature and LTL drivers changed simultaneously in the All (Climate) Change 500 501 simulation, shifts in the distribution of consumer biomass for each model were a combination 502 of the shifts driven by separate temperature and LTL effects (Figure 3, right column; 503 Supplementary Figure S1). Across all models, temperature-induced declines in consumer 504 biomass were generally exacerbated in regions where LTL changes negatively impacted 505 consumer biomass. Overall, consumer biomass generally increased in polar waters, where all 506 LTL variables increased but temperature changed relatively little. Increases in consumer 507 biomass in DBEM were greater in polar regions under climate change, compared to the sum 508 of the separate impacts of warming and LTL shifts (Supplementary Figure S1e). Outside of 509 polar regions, the magnitude and direction of change in consumer biomass varied among 510 models, depending on their individual responses to temperature and LTL changes. For BOATS and Macroecological, the magnitude of positive and negative changes in consumer biomass 511 512 from LTL shifts in isolation were attenuated when combined with the impacts of warming in 513 the Climate Change simulation (Supplementary Figure S1g, h), however these non-additive 514 effects largely cancelled at the global scale (Figure 2d).

515



Figure 3 Maps of relative total consumer biomass averaged over 2090-2100, compared to the
Control (over 2090-2100), for the Temperature (left column), Lower Trophic Level (LTL) and

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10	uman		ρισ	\mathbf{O}

519	All Change simulations for a-c) APECOSM, d-f) FEISTY, g-i) ZooMSS, j-l) DBPM, m-o) DBEM, p-
520	r) EcoTroph, s-u) BOATS, v-x) Macroecological. Maps are ordered by the magnitude (from
521	smallest to greatest) of the negative warming impact on consumer biomass.
522	
523	3.3. Disentangling temperature and lower trophic level impacts on total consumer
524	biomass
525	Figure 4 compares the forced changes in sea surface temperature (SST) with the co-located
526	simulated changes in biomass for all grid cells in the global ocean. Regressions give negative
527	exponential slopes for all models, but with substantial variation (Supplementary Table S1).
528	Globally, consumer biomass changed between -0.5% and -2.0% for every 1°C of sea surface
529	warming for APECOSM, FEISTY, DBPM and ZooMSS, and between -4.8% and -15.4% per 1°C
530	across EcoTroph, BOATS and Macroecological (Supplementary Table S4). The models vary in
531	their degree of linearity, with DBEM projecting the greatest nonlinearity in the impacts of
532	warming between cold and warm waters (Figure 4e; Supplementary Table S4). DBEM
533	consumer biomass increased by ~50% in cold waters (<15°C SST) in response to warming
534	(Figure 4e), and decreased on average by >27% for each 1°C warming in warm (\geq 15°C SST)
535	waters.

536

Figure 5 shows the corresponding plots for LTL forcing. For all models, changes in total consumer biomass were positively correlated with changes in their respective aggregated lower trophic level (LTL) forcing (Figure 5). A 1% change in LTL forcings caused a change in total consumer biomass of between 0.6% in DBPM to 1.7% in BOATS (Supplementary Table S4). Positive correlations between consumer biomass and LTL changes ranged from r = 0.39for DBPM, to r = 0.98 for EcoTroph. For all models except DBPM, the greatest correlation

543 was between change in total consumer biomass and change in total LTL production, or 544 biomass, of the model's chosen LTL forcing (Supplementary Table S3). In models that used 545 size-fractionated LTL inputs, or additional secondary LTL inputs, changes in consumer biomass 546 were less correlated with changes in their main aggregated LTL forcing (APECOSM, DBPM, 547 FEISTY) compared to models that did not use size-fractionated or multiple LTL forcings 548 (BOATS, DBEM, EcoTroph, Macroecological, ZooMSS).



Figure 4 Change in total consumer biomass (%) against the mean change in sea surface 551 552 temperature (SST) over 2090-2100, for individual 1° grid squares, under the Temperature

553 Change simulation, compared to the Control simulation, for a) APECOSM, b) FEISTY, c) 554 ZooMSS, d) DBPM, e) DBEM, f) EcoTroph, g) BOATS, h) Macroecological. Each point is coloured according to the mean 1950-1960 historical SST in its grid cell. Dotted horizontal and vertical 555 556 black lines indicate where % change in total consumer biomass and change in temperature 557 are zero, respectively. The green line is the fitted regression (Δ Total Consumer Biomass = $\exp(\beta_0 + \beta_1 \Delta SST) + \epsilon)$ for the change in consumer biomass with warming. We use 558 559 exponential regression to calculate the line of best fit here since all models incorporate temperature effects using an exponential function (see Supplementary Information). 560 Information about the fitted regression is in Supplementary Table S1 and S4. 561



Figure 5 Change in total consumer biomass (%) against change in aggregated lower trophic 564 level forcings (LTL), from 2090-2100 under the LTL Change simulation, against the Control, for 565 566 individual 1° grid squares, compared to the Control simulation for a) APECOSM, b) FEISTY, c) ZooMSS, d) DBPM, e) DBEM, f) EcoTroph, g) BOATS, h) Macroecological, with Pearson's 567 correlation coefficient (r) reported for each. Each point is coloured according to the average 568 569 1950-1960 historical sea surface temperature (SST) in its corresponding grid cell. The black 570 solid line is the 1:1 line, and the dotted horizontal and vertical black lines indicate where % 571 change in total consumer biomass and % change in LTL are zero, respectively. The green line 572 is the fitted regression (Δ Total Consumer Biomass = $\beta_0 + \beta_1 \Delta$ LTL + ϵ) for the change in

573 consumer biomass with warming. For models that use more than one LTL variable 574 (APECOSM), or size-fractionated LTL (FEISTY and DBPM), Δ LTL is calculated from the sum of 575 all LTL forcings. Information about the fitted regression is in Supplementary Table S2 and S4. 576

577 **3.4.** Impacts of warming and lower trophic level change on ecosystem structure

In response to warming, there was little consensus in the relative change of small (<30 cm) 578 and large (\geq 30 cm) mean global consumer biomass (Figure 6a), with four models (BOATS, 579 580 EcoTroph, Macroecological, ZooMSS) showing a decrease of both and the other three models 581 (APECOSM, DBPM, FEISTY) showing a mixture of responses. Small consumer biomass increased by ~2% in both APECOSM and FEISTY in response to warming, but large consumer 582 biomass increased in APECOSM by 5% while decreasing in FEISTY by >10%. Similarly, although 583 584 small consumer biomass in DBPM and ZooMSS decreased by 3% and 6% respectively, these 585 models disagreed on the direction of change for large consumer biomass. In response to 586 warming total large consumer biomass in DBPM increased by 15%, and over 60% in some 587 regions (Supplementary Figure S2n), but in ZooMSS total large consumer biomass declined by ~2% overall. Finally, small and large consumer biomass declined in EcoTroph, BOATS and 588 589 Macroecological, and the spatial pattern of decline across was similar both across models and 590 across small and large consumer biomass (Supplementary Figure S2). There was no difference in the magnitude of the decline of small and large consumer biomass in Macroecological, 591 592 however in EcoTroph and BOATS the magnitude of the decline in large consumer biomass was 593 greater than the decline in small consumer biomass.

594

595 Changes in total small and large consumer biomass in response to lower trophic level (LTL) 596 changes show more agreement (Figure 6b). The change in total small and large consumer

597 biomass was similar in APECOSM, FEISTY, ZooMSS, EcoTroph, BOATS and Macroecological, 598 and again the spatial pattern of change in small and large consumer biomass generally 599 followed each model's respective LTL forcings (Supplementary Figure S2; Figure 1f-h). 600 However, in DBPM, total large consumer biomass declined by 40%, while small consumer 601 biomass declined by only 10%. This was reflected in the spatial patterns of biomass change 602 for DBPM, with large consumer biomass varying by over $\pm 60\%$ and small consumer biomass varying by less than $\pm 30\%$ across non-polar regions of the global ocean (Supplementary 603 604 Figure S2 o,p).

605

606 Small and large consumer biomass declined for all models (except DBEM, which was excluded 607 from this part of the analysis since it did not provide size-fractionated biomass) in response 608 to climate change (simultaneous temperature and LTL changes) impacts (Figure 6c). Large 609 consumer biomass declined more than small consumer biomass in BOATS, DBPM, EcoTroph 610 and FEISTY. In contrast, small consumer biomass declined more than large consumer biomass 611 in APECOSM and ZooMSS, and there was no difference between small and large consumer 612 biomass change in Macroecological. For all models, the impacts of climate change on small and large consumers were largely the sum of temperature and LTL impacts, with relatively 613 614 small non-additive impacts (Figure 6d).



Figure 6 Change in total small (<30 cm) consumer biomass versus change in global large 617 (>30cm) consumer biomass averaged between 2090-2100 relative to Control simulation for 618 each model (excluding DBEM, which did not provide small and large consumer biomass) in 619 620 the a) Temperature Change simulation b) Lower Trophic Level (LTL) Change simulation, c) All Change simulation and d) the non-additive impacts of temperature and LTL changes, 621 622 calculated by taking the difference between the All Change and the sum of the Temperature 623 and LTL Change simulations. The red solid line is the 1:1 line, and the dotted horizontal and vertical black lines indicate where the percentage change in global large and small consumer 624 625 biomass are zero, respectively.

626

628 **4. Discussion**

629 The results of our experimental protocol reveal commonalities, as well as contrasts among 630 the FishMIP models. All models agreed that the combination of warming and lower trophic level (LTL) shifts will cause substantial regional changes in consumer biomass. Furthermore, 631 632 no model projected a significant increase in global biomass in response to climate change. However, the impacts of warming varied markedly between models, leading to large inter-633 model disagreements. Changes in LTL drivers were more directly correlated with the 634 635 outcomes on consumer biomass, but with substantial variation among models, and strong 636 dependence on each model's choice of LTL driver. For almost all models, the combined impacts of warming and LTL changes were largely additive at the global scale, showing little 637 638 nonlinear interaction, and additional climate change drivers (e.g., oxygen, acidification, 639 current speeds) were not significant global drivers in the models that included them 640 (APECOSM and DBEM). By separating the marine ecosystem model responses to climate-641 driven warming versus LTL shifts, our results point toward the processes that need to be 642 clarified to reduce the uncertainty of how these two dominant drivers impact marine ecosystems. 643

644

645 **4.1. Warming impacts are complex**

One straightforward expectation might be that the different responses to warming reflect differences in the temperature scalings used in each model. However, the differences in temperature scalings do not readily explain the variation in the results. For instance, DBPM and Macroecological use identical temperature scalings (see Supplementary Information S2.4 and S2.7), yet DBPM's projections of warming-induced biomass decline are almost an order of magnitude smaller than those of Macroecological. This does not mean the temperature

scalings are irrelevant, but rather that the emergent results depend on the interactions of
 multiple temperature-dependent processes, operating within the structural context of each
 model.

655

656 For example, DBEM attempts to resolve preferred temperature ranges for different species, 657 while the other models consider the effect of temperature on generalized physiological processes, implicitly assuming that species moving out of a region are replaced by species 658 659 moving in with no change in ecosystem function. Although food web processes such as 660 predator-prey interactions are not explicitly included in the DBEM species-distribution model, 661 it projects an emptying of tropical waters and a corresponding build-up of biomass in polar 662 waters, as species move poleward to follow their thermal preferences. This redistribution of the biomass of >1200 recorded commercial species included in the model reflects the absence 663 of very warm water fish that can repopulate the tropics, and the small number of cold water 664 fish in the initial state (Cheung et al. 2010). It also largely explains the model's combined 665 666 impacts of warming, LTL shifts and other drivers being nonlinear: relative to extant species in 667 polar waters, a larger number of species follow their thermal niche poleward and are able to 668 take advantage of increased primary production in high latitude regions, compared to the regions they left behind. 669

670

DBEM aside, four of the models included here (APECOSM, DBPM, FEISTY and ZooMSS) project much smaller warming impacts on consumer biomass than the remaining three models (BOATS, EcoTroph and Macroecological). Although there are many differences between these models, one particularly salient feature is that the low-sensitivity models all use LTL biomass as Earth-system model (ESM) drivers for the projections, together with

temperature-dependent feeding rates. In contrast, the high-sensitivity models use ESM net primary production to directly limit the growth rates of upper trophic levels. We suggest that the discrepancy in temperature sensitivity between the model groups can be attributed, at least in part, to an inconsistency that arises from the 1-way forcing of marine models with LTL biomass. The relationship between LTL biomass (B_{LTL}), LTL production (P_{LTL}) and higher trophic level predation (Pred_{HTL}) through time *t* can be represented as:

$$\frac{dB_{LTL}}{dt} = P_{LTL}(t) - Pred_{HTL}(t)$$

In reality, if warming accelerates predation rates, but lower trophic level production remains 683 constant or does not increase as much, such that $P_{LTL}(t) < Pred_{HTL}(t)$, LTL biomass would 684 685 decrease. However in the 1-way forcing used here, LTL biomass is determined externally by the Earth system model and is not affected by predation from higher trophic levels. Thus, 686 increased predation rates from warming on fixed LTL biomass causes an increase in the flux 687 of biomass energy into higher trophic levels that is decoupled from lower trophic level 688 689 production. This increased energy input counters the increased metabolic rates and 690 associated respiratory losses, dampening biomass declines from warming. This inconsistency 691 in coupling between LTLs and higher trophic level consumers would tend towards an underestimate of warming impacts on consumer biomass. In contrast, in the production-692 693 driven models there is no spurious energy input under warming, so that warming-driven 694 increases in consumer respiration costs and decreases in representative phytoplankton size 695 act to drive biomass down strongly.

696

697 Our results also explored the warming impacts on ecosystem structure, defined as the relative698 biomass of small versus large organisms. Here, there was little consensus between models.

699 DBPM and FEISTY provide a striking example of divergent projections of ecosystem structure 700 with warming. In DBPM, ingestion-fuelled anabolism outpaces senescence-induced mortality 701 in large organisms as waters warm (Blanchard et al. 2012), causing their biomass to increase. 702 This raises predation pressure on smaller organisms, which when coupled with warming-703 induced increases in natural mortality, causes their biomass to decline. By contrast, in FEISTY, 704 biomass respiration increases faster with both body size and temperature compared to ingestion-fuelled anabolism (Petrik et al. 2019) reducing the scope for growth and causing 705 706 large organism biomass to decline with warming. Declines in large consumer biomass in 707 FEISTY with increasing temperature relieve predation pressure on small consumers, resulting 708 in an increase in their biomass, especially in tropical waters. The divergent impacts of warming 709 on individual processes and ecosystem structure reflects the lack of consensus among 710 modellers of how temperature impacts on individuals translate into ecosystem impacts.

711

712 **4.2.** Lower trophic level impacts are influenced by choice of forcing

713 The choice of LTL forcings differed between models, with each model using either biomass or 714 production variables at the phytoplankton or zooplankton level, with significant impacts on the results. Generally, spatial changes in consumer biomass were most correlated with 715 716 changes in the distribution of the LTL forcing used. The sensitivity of models to the choice of LTL forcing again indicates a lack of common understanding of how to link lower trophic levels 717 718 production with higher trophic levels, with no consensus on whether production rates or 719 standing-stock biomass should be used. We believe this problem fundamentally arises out of 720 practical necessity because of each model's one-way, offline coupling with the Earth system 721 model—were higher trophic levels and LTLs to be fully coupled, and predation feedbacks on 722 LTLs resolved, there should theoretically be no disagreement between models that use

production or biomass variables, everything else being equal. However, in the absence of twoway coupled models in the FishMIP ensemble, the development of which is a tremendous technical challenge (see Aumont et al., 2018), this problem remains to be addressed. As mentioned above, this problem also leads to inconsistency in the temperature response when plankton biomass versus net primary production rates are used.

728

Ecosystem structure did not change substantially in response to LTL changes, except in DBPM. 729 730 Large organism biomass in DBPM declined by 40% and small organisms declined by <10% in 731 response to decreases in phytoplankton biomass and resultant shifts in the size structure of 732 the phytoplankton abundance spectrum. DBPM's relatively large decrease in large consumer 733 biomass in response to phytoplankton biomass declines is a result of biomass destruction 734 through senescence mortality, which increases with body size but does not depend on food 735 density, outpacing ingestion-fuelled biomass creation. The other predation-explicit modelsincluding ZooMSS and FEISTY, which also include biomass destruction processes independent 736 737 of food density that increase with body size—did not exhibit similar declines in large organism 738 biomass. This is because in these models, ingestion-fuelled growth outpaces biomass destruction from these processes, highlighting the sensitivity of model outputs to the 739 740 parameterisation of these rates. In fact, across all models except DBPM, the change in large organism biomass with LTL change was equal to or slightly less than the change in small 741 organism biomass. 742

743

744 **4.3.** Cumulative warming and lower trophic level impacts are largely additive

Across the model ensemble, climate change impacts on total consumer biomass and
ecosystem structure were generally well-approximated by the sum of separate warming and

747 LTL impacts. This lack of non-linearity is perhaps less surprising for the majority of models that 748 only use temperature and LTL drivers to force their models (Tittensor et al. 2018), but 749 remarkably it also holds for APECOSM, which incorporates other drivers such as oxygen, pH 750 and current velocity. The fact that the overall climate change impact on consumer biomass in 751 APECOSM was close to the sum of temperature and LTL impacts indicates that the additional 752 forcings have a comparatively small effect. DBEM, which also includes additional environmental drivers, did show a much stronger non-additive impact of climate change on 753 754 overall consumer biomass, but this appeared to be driven primarily by the relocation of 755 species niches in DBEM in response to warming, rather than the other drivers. DBEM aside, only BOATS and Macroecological show significant non-linear interactions between 756 757 temperature and LTL drivers. This can be attributed to the fact that, in BOATS and 758 Macroecological, the representative size of phytoplankton used to force the models scales 759 with both net primary production and temperature, increasing in cooler waters or regions 760 with high net primary production (Dunne et al. 2005). For these two models, the spatial 761 pattern of attenuation follows shifts in net primary production, indicating that warming 762 attenuates the increases and decreases in biomass from shifts in net primary production.

763

It may be tempting to assume that the lack of nonlinear interactions in the models means that such nonlinearities are unlikely to exist in the ocean. However, an increasing number of experimental and observational studies indicate that cumulative impacts from climate change stressors such as warming, deoxygenation and acidification are likely to be nonlinear and amplifying (Sampaio and Rosa, 2020). Rather, given the rudimentary representation of many ecosystem processes in the models (e.g., no phenological or diversity-related mechanisms, simplistic or absent predation relationships), we suggest that it is more appropriate to ascribe

the lack of nonlinear interactions in marine climate change projections to our present lack ofability to resolve them in the models.

773

4.4. Improving marine ecosystem models with observational constraints

775 In this study, we have identified key sources of structural uncertainty that drive disparate 776 projections of climate change impacts on the global marine ecosystem. As a first step, the marine modelling community can work to reduce this structural uncertainty and increase the 777 778 credibility of ecosystem projections by constraining models with independent observations. 779 An increasingly popular approach to confront model projections with observations is to use emergent constraints, which relate the long-term climate sensitivity of an observable 780 ecosystem feature - such as total biomass change (Free et al. 2019) or size-spectrum slope 781 782 (Blanchard et al. 2017; Heneghan et al. 2019) - to its short-term, observed variability (Allen & 783 Ingram, 2002; Eyring et al. 2019). Models that give a closer fit to short-term observed 784 variability of an ecosystem feature are hypothesised to provide more reliable projections of 785 its long-term variability from climate change (Kwiatkowski et al. 2017; Veytia et al. 2020). 786 Moreover, within a model ensemble, each model's weighting can be linked to its ability to 787 capture the emergent constraint (Eyring et al. 2019). This provides a more sophisticated and 788 credible way to weight model projections within an ensemble, over the standard approach 789 where all models are given equal weighting (known as model democracy), irrespective of 790 performance (Knutti, 2010). Emergent constraints do not require or necessarily reward any 791 particular ecosystem representation. This is important as differing representations of the 792 marine ecosystem across the FishMIP ensemble not only represent our present uncertainty of the most important drivers structuring marine ecosystems, but also the diversity of 793 794 purpose and scope for which models have been built.

796 Finally, it is possible for models to perform well against whole-ecosystem emergent 797 constraints, while neglecting fundamental physiological or ecosystem processes (Knutti, 798 2010). Therefore, if we are to improve marine models, it is also necessary to consider 799 observational constraints on physiological processes such as the balance between growth and 800 respiration with temperature, or ecosystem processes such as the coupling of lower and higher trophic levels. Improving our understanding of how physiological processes such as 801 802 ingestion and metabolism respond to warming, and how changes in LTL processes propagate 803 through marine ecosystems, are critical steps towards model improvement and more robust 804 climate impact projections.

805

806 **5. Concluding remarks**

Projecting the global impact of climate change on marine ecosystems and fisheries is an important and challenging task. Marine ecosystem models represent the current understanding of how climate change could impact the food web and fisheries globally in the future. Yet, although these models have made great strides in recent years, our results show that the current understanding falls short in many respects.

812

Our harmonized experimental protocol clearly showed that the responses to the two most important drivers of change – warming and LTL shifts – differ widely among models. Uncertainty in the temperature sensitivities of competing processes, including both physiology and ecological interactions, undermine confidence in the emergent sensitivities, and can only be improved with better observational constraints. Meanwhile, the outcome of changes in both water temperature and LTL production depends strongly on the feedback of

consumers on the LTL biomass itself, a process which is not captured by any of the one-way
forcings available at present, and can only be rectified with fully two-way coupling, which is
itself sure to raise many new questions.

822

823 What are the implications of our results for single ecosystem model studies? The eight models 824 used here differ significantly in their design and ecosystem representation, having been built for different purposes (Tittensor et al. 2018). Although using common outputs across models 825 has been useful here to identify shared weaknesses, this approach conceals the strengths of 826 827 individual models to resolve certain processes and ecosystem components that other models do not. Thus, studies that explore the unique strengths and weaknesses of individual models 828 829 remain important, in order to explore questions that each model has been designed to 830 address. However, results of these single model studies should be interpreted within the 831 greater context of sources of structural uncertainty shared across models identified here.

832

Attempting to summarise the vast complexity of the global marine ecosystem in a handful of equations is enormously difficult. The fact that independently constructed models with contrasting architectures have arrived at many similar conclusions is encouraging, while their diversity is useful to identify common weaknesses. These initial results from the FishMIP ensemble provide a glimpse into the great promise of multi-model comparisons to improve our understanding of the global marine ecosystem and its future under change.

839

840 Code and data availability

- The experimental protocol in this paper has no code associated with it. Forcing data from CMIP5 used for the protocol, and the FishMIP model outputs presented in this paper are available on the ISIMIP servers (https://www.isimip.org/).
- 844

845 Author contributions

JLB, TDE, EDG and DPT led the conceptualisation and development of the protocol for this study, with contributions from the other authors. CH and RFH obtained and processed forcings for the modellers to complete the protocol, with assistance from JV. Model simulations were conducted by RFH, NB, CB, WC, MC, TDE, ME, JAG, DF, JG, OM, JP, CMP, HdP, JS, TCT, PAW. RFH conducted the analysis, with assistance from EDG and JLB. RFH led the writing of the text, with feedback and contributions from all authors.

852

853 Competing interests

The authors declare that they have no conflict of interest.

855

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871 References

- Allen, M. R., & Ingram, W. J. (2002). Constraints on future changes in climate and the
- 873 hydrologic cycle. *Nature*, *419*(6903), 228–232. <u>https://doi.org/10.1038/nature01092</u>
- Aumont, O., Maury, O., Lefort, S., & Bopp, L. (2018). Evaluating the potential impacts of the
- diurnal vertical migration by marine organisms on marine biogeochemistry. Global *Biogeochemical Cycles, 32*, 1622-1643. https://doi.org/10.1029/2018GB005886
- Barange, M., Merino, G., Blanchard, J. L., Scholtens, J., Harle, J., Allison, E. H., ... Jennings, S.
 (2014). Impacts of climate change on marine ecosystem production in societies dependent
 on fisheries. *Nature Climate Change*, *4*, 211. Retrieved from
- 880 https://doi.org/10.1038/nclimate2119
- Barnes, C., Irigoien, X., De Oliveira, J. A. A., Maxwell, D., & Jennings, S. (2011). Predicting
 marine phytoplankton community size structure from empirical relationships with remotely
 sensed variables. *Journal of Plankton Research*, 33(1), 13–24.
 https://doi.org/10.1093/plankt/fbq088
- Blanchard, J. L., Heneghan, R. F., Everett, J. D., Trebilco, R., & Richardson, A. J. (2017b). From
 Bacteria to Whales: Using Functional Size Spectra to Model Marine Ecosystems. *Trends in Ecology and Evolution*, 32(3), 174–186. https://doi.org/10.1016/j.tree.2016.12.003
- Blanchard, J. L., Jennings, S., Holmes, R., Harle, J., Merino, G., Allen, J. I., ... Barange, M. (2012).
 Potential consequences of climate change for primary production and fish production in large
- 890 marine ecosystems. Philosophical Transactions of the Royal Society B: Biological Sciences,
- 891 *367*(1605), 2979–2989. https://doi.org/10.1098/rstb.2012.0231

- Blanchard, J. L., Jennings, S., Law, R., Castle, M. D., McCloghrie, P., Rochet, M. J., & Benoît, E.
 (2009). How does abundance scale with body size in coupled size-structured food webs? *Journal of Animal Ecology*, 78(1), 270–280. https://doi.org/10.1111/j.13652656.2008.01466.x
- 896 Blanchard, J. L., Watson, R. A., Fulton, E. A., Cottrell, R. S., Nash, K. L., Bryndum-Buchholz, A., 897 ... Müller, C. (2017a). Linked sustainability challenges and trade-offs among fisheries, 898 aquaculture and agriculture. Nature Ecology & Evolution, 1240-1249. 1, 899 https://doi.org/10.1038/s41559-017-0258-8
- Bopp, L., Resplandy, L., Orr, J.C., Doney, S.C., Dunne, J.P., Gehlen, M., Halloran, P., Heinze, C.,
- 901 Ilyina, T., Séférian, R., Tjiputra, J., Vichi, M., (2013). Multiple stressors of ocean ecosystems in
- 902 the 21st century: projections with CMIP5 models. *Biogeosciences 10,* 6225–6245.
- 903 https://doi.org/10.5194/bg-10-6225-2013
- Boyce, D. G., Frank, K. T., & Leggett, W. C. (2015). From mice to elephants: overturning the

'one size fits all' paradigm in marine plankton food chains. *Ecology Letters*, 18(6), 504–515.

906 https://doi.org/10.1111/ele.12434

- 907 Boyce, D. G., Lotze, H. K., Tittensor, D. P., Carozza, D. A., & Worm, B. (2020). Future ocean 908 biomass losses may widen socioeconomic equity gaps. *Nature Communications*, 1–11.
- 909 Retrieved from https://doi.org/10.1038/s41467-020-15708-9
- 910 Brander, K., Neuheimer, A., Andersen, K. H., & Hartvig, M. (2013). Overconfidence in model
- 911 projections. *ICES Journal of Marine Science*, *70*, 1065–1068.

912	Brito-Morales, I., Schoeman, D. S., Molinos, J. G., Burrows, M. T., Klein, C. J., Arafeh-dalmau,
913	N., Garilao, C. (2020). Climate velocity reveals increasing exposure of deep-ocean
914	biodiversity to future warming. <i>Nature Climate Change</i> , 10(June), 576–581.
915	https://doi.org/10.1038/s41558-020-0773-5
916	Brown, J. H., Gillooly, J. F., Allen, A. P., Savage, V. M., & West, G. B. (2004). Toward a metabolic
917	theory of ecology. <i>Ecology</i> , <i>85</i> (7), 1771–1789. https://doi.org/10.1890/03-9000
918	Bryndum-Buchholz, A., Prentice, F., Tittensor, D. P., Blanchard, J. L., Cheung, W. W. L.,
919	Christensen, V., Lotze, H. K. (2020). Differing marine animal biomass shifts under 21st
920	century climate change between Canada's three oceans. Facets, (5), 105–122.
921	https://doi.org/10.1139/facets-2019-0035
922	Carozza, D. A., Bianchi, D., & Galbraith, E. D. (2016). The ecological module of BOATS-1.0: a

bioenergetically constrained model of marine upper trophic levels suitable for studies of
fisheries and ocean biogeochemistry. *Geosci. Model Dev.*, 9(4), 1545–1565.
https://doi.org/10.5194/gmd-9-1545-2016

Carozza, D. A., Bianchi, D., & Galbraith, E. D. (2017). Formulation, General Features and Global
Calibration of a Bioenergetically-Constrained Fishery Model. *Plos One*, *12*(1), e0169763.
https://doi.org/10.1371/journal.pone.0169763

929 Carozza, D. A., Bianchi, D., & Galbraith, E. D. (2018). Metabolic impacts of climate change on
930 marine ecosystems: Implications for fish communities and fisheries. *Global Ecology and*931 *Biogeography*, 28, 158–169. https://doi.org/10.1111/geb.12832

932	Chavez,	F. P., Messi,	M., &	Penningto	on, J. T. (2	011). Ma	rine	Primary I	Production	in F	Relation to
933	Climate	Variability	and	Change.	Annual	Review	of	Marine	Science,	3,	227–260.
934	https://d	doi.org/10.12	146/a	nnurev.ma	rine.0109	908.16392	17				

Cheung, W. W. L., Dunne, J., Sarmiento, J. L., & Pauly, D. (2011). Integrating ecophysiology
and plankton dynamics into projected maximum fisheries catch potential under climate
change in the Northeast Atlantic. *ICES Journal of Marine Science*, *68*(6), 1008–1018.
https://doi.org/10.1093/icesjms/fsr012

939 Cheung, W. W. L., Frolicher, T. L., Asch, R. G., Jones, M. C., Pinsky, M. L., Reygondeau, G., ...

940 Watson, J. R. (2016). Building confidence in projections of the responses of living marine 941 resources to climate change. *ICES Journal of Marine Science*, *73*, 1283–1296.

942 Cheung, W. W. L., Lam, V. W. Y., & Pauly, D. (2008). Modelling Present and Climate-Shifted

943 Distribution of Marine Fishes and Invertebrates. *Fisheries Centre Research Reports*, *16*(3), 76.

944 https://doi.org/papers3://publication/uuid/4F260971-0169-4716-AFAA-5F970130A95C

Cheung, W. W. L., Lam, V. W. Y., Sarmiento, J. L., Kearney, K., Watson, R., Zeller, D., & Pauly,
D. (2010). Large-scale redistribution of maximum fisheries catch potential in the global ocean
under climate change. *Global Change Biology*, *16*(1), 24–35. https://doi.org/10.1111/j.13652486.2009.01995.x

Coll, M., Steenbeek, J., Pennino, M. G., Buszowski, J., Kaschner, K., Lotze, H. K., ... Christensen,
V. (2020). Advancing Global Ecological Modeling Capabilities to Simulate Future Trajectories
of Change in Marine Ecosystems. *Frontiers in Marine Science*, 7, 1–23.
https://doi.org/10.3389/fmars.2020.567877

953	du Pontavice, H., Gascuel, D., Reygondeau, G., Maureaud, A., & Cheung, W. W. L. (2020).
954	Climate change undermines the global functioning of marine food webs. Global Change
955	<i>Biology</i> , <i>26</i> (3), 1306–1318. https://doi.org/10.1111/gcb.14944

- Dunne, J. P., Armstrong, R. A., Gnanadesikan, A., & Sarmiento, J. L. (2005). Empirical and
 mechanistic models for the particle export ratio. *Global Biogeochemical Cycles*, *19*(December), 1–16. https://doi.org/10.1029/2004GB002390
- 959 Eddy, T. D., Bernhardt, J. R., Blanchard, J. L., Cheung, W. W. L., Colléter, M., Pontavice, H., ...
- 960 Watson, R. A. (2020). Trends in Ecology & Evolution Energy Flow Through Marine Ecosystems :
- 961 Confronting Transfer Efficiency. *Trends in Ecology & Evolution, xx*(xx), 1–11. Retrieved from
- 962 https://doi.org/10.1016/j.tree.2020.09.006
- 963 Eyring, V., Cox, P. M., Flato, G. M., Gleckler, P. J., Abramowitz, G., Caldwell, P., ... Williamson,
- 964 M. S. (2019). Taking climate model evaluation to the next level. *Nature Climate Change*, 9(2),
- 965 102–110. https://doi.org/10.1038/s41558-018-0355-y
- 966 Finkel, Z. V., Beardall, J., Flynn, K. J., Quigg, A., Rees, T. A. V., & Raven, J. A. (2010).
- 967 Phytoplankton in a changing world: Cell size and elemental stoichiometry. *Journal of Plankton*
- 968 *Research*, *32*(1), 119–137. https://doi.org/10.1093/plankt/fbp098
- 969 Free, C. M., Thorson, J. T., Pinsky, M. L., Oken, K. L., Wiedenmann, J., & Jensen, O. P. (2019).
- 970 Impacts of historical warming on marine fisheries production. *Science*, *363*(March), 979–983.
- 971 <u>https://doi.org/10.1126/science.aau1758</u>
- 972 Friedland, K.D., Stock, C., Drinkwater, K.F., Link, J.S., Leaf, R.T., Shank, B.V., Rose, J.M., Pilskaln,
- 973 C.H., Fogarty, M.J., (2012). Pathways between Primary Production and Fisheries Yields of

	Journal Pre-proofs									
974	Large	Marine	Ecosystems.	PLOS	ONE	7,	e28945.			
975	75 https://doi.org/10.1371/journal.pone.0028945									

Gascuel, D., & Pauly, D. (2009). EcoTroph : Modelling marine ecosystem functioning and
impact of fishing. *Ecological Modelling*, 220, 2885–2898.
https://doi.org/10.1016/j.ecolmodel.2009.07.031

- 979 Heneghan, R. F., Everett, J. D., Blanchard, J. L., & Richardson, A. J. (2016). Zooplankton Are Not Fish: Improving Zooplankton Realism in Size-Spectrum Models Mediates Energy Transfer 980 981 in Food Webs. Frontiers in Marine Science, 3(October), 1–15. 982 https://doi.org/10.3389/fmars.2016.00201
- Heneghan, R. F., Everett, J. D., Sykes, P., Batten, S. D., Edwards, M., Takahashi, K., ...
 Richardson, A. J. (2020). A functional size-spectrum model of the global marine ecosystem
 that resolves zooplankton composition. *Ecological Modelling*, *435*(August), 109265.
 <u>https://doi.org/10.1016/j.ecolmodel.2020.109265</u>
- 987 Heneghan, R. F., Hatton, I. A., & Galbraith, E. D. (2019). Climate change impacts on marine

ecosystems through the lens of the size spectrum. Emerging Topics in Life Sciences, 3(2), 233-

- 989 243. <u>https://doi.org/10.0142/etls20190042</u>
- 990 IPCC, 2014: Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and
- 991 III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Core
- 992 Writing Team, R.K. Pachauri and L.A. Meyer (eds.)]. IPCC, Geneva, Switzerland, 151 pp.
- Jennings, S., & Collingridge, K. (2015). Predicting consumer biomass, size-structure,
 production, catch potential, responses to fishing and associated uncertainties in the world's

	Journal Pre-proofs							
995	marine	ecosystems.	PLoS	ONE,	10(7),	e0133794.		
996	6 https://doi.org/10.1371/journal.pone.0133794							

- Jones, M. C., Dye, S. R., Pinnegar, J. K., Warren, R., & Cheung, W. W. L. (2012). Modelling
- 998 commercial fish distributions: Prediction and assessment using different approaches.

999 Ecological Modelling, 225, 133–145. https://doi.org/10.1016/j.ecolmodel.2011.11.003

1000 Knutti, R. (2010). The end of model democracy ?: An editorial comment. *Climatic Change*, *102*,

1001 395–404. https://doi.org/10.1007/s10584-010-9800-2

1002 Kooijman, S. (2010). Dynamic Energy Budget theory for metabolic organisation (3rd ed.).
1003 Cambridge: Cambridge University Press.

1004 Kwiatkowski, L., Bopp, L., Aumont, O., Ciais, P., Cox, P. M., Laufkötter, C., ... Séférian, R. (2017).

1005 Emergent constraints on projections of declining primary production in the tropical oceans.

1006 Nature Climate Change, 7(April), 355–359. https://doi.org/10.1038/NCLIMATE3265

1007 Lefevre, S., McKenzie, D. J., & Nilsson, G. E. (2017). Models projecting the fate of fish
1008 populations under climate change need to be based on valid physiological mechanisms.

1009 Lotze, H. K., Tittensor, D. P., Bryndum-Buchholz, A., Eddy, T. D., Cheung, W. W. L., Galbraith,

1010 E. D., ... Worm, B. (2019). Global ensemble projections reveal trophic amplification of ocean

1011 biomass declines with climate change. Proceedings of the National Academy of Sciences of

 1012
 the
 United
 States
 of
 America,
 116(26),
 12907–12912.

 1013
 https://doi.org/10.1073/pnas.1900194116

 </td

Maury, O. (2010). An overview of APECOSM, a spatialized mass balanced "Apex Predators
ECOSystem Model" to study physiologically structured tuna population dynamics in their

 1016
 ecosystem.
 Progress
 in
 Oceanography,
 84(1-2),
 113-117.

 1017
 https://doi.org/10.1016/j.pocean.2009.09.013

1018 Maury, O., Faugeras, B., Shin, Y. J., Poggiale, J. C., Ari, T. Ben, & Marsac, F. (2007a). Modeling

1019 environmental effects on the size-structured energy flow through marine ecosystems. Part 1:

1020 The model. *Progress in Oceanography*, 74(4), 479–499. 1021 https://doi.org/10.1016/j.pocean.2007.05.002

1022 Maury O., J.-C. Poggiale, 2013. From individuals to populations to communities: a Dynamic

1023 Energy Budget model of marine ecosystem size-spectrum including life history diversity.

1024 Journal of Theoretical Biology. 324, 52–71. https://doi.org/10.1016/j.jtbi.2013.01.018

1025 Maury, O., Shin, Y. J., Faugeras, B., Ari, T. Ben, & Marsac, F. (2007b). Modeling environmental

1026 effects on the size-structured energy flow through marine ecosystems. Part 2: Simulations.

1027 *Progress* in *Oceanography*, 74(4), 500–514.

1028 https://doi.org/doi:10.1016/j.pocean.2007.05.001

Moore, J.K., Lindsay, K., Doney, S.C., Long, M.C., Misumi, K., (2013). Marine Ecosystem Dynamics and Biogeochemical Cycling in the Community Earth System Model CESM1(BGC): Comparison of the 1990s with the 2090s under the RCP4.5 and RCP 8.5 Scenarios. *J. Clim. 26*,

1032 9291–9312. https://doi.org/10.1175/JCLI-D-12-00566.1

Payne, M. R., Barange, M., Cheung, W. W. L., Mackenzie, B. R., Batchelder, H. P., Cormon, X.,
... Queiro, A. M. (2016). Uncertainties in projecting climate-change impacts in marine
ecosystems. *ICES Journal of Marine Science*, *73*(5), 1272–1282.
https://doi.org/doi:10.1093/icesjms/fsv231

- Petrik, C. M., Stock, C. A., Andersen, K. H., van Denderen, P. D., & Watson, J. R. (2019). Bottomup drivers of global patterns of demersal, forage, and pelagic fishes. *Progress in Oceanography*, *176*, 102124. https://doi.org/https://doi.org/10.1016/j.pocean.2019.102124
- 1040 Pinsky, M. L., Selden, R. L., & Kitchel, Z. J. (2020). Climate-Driven Shifts in Marine Species
- 1041 Ranges: Scaling from Organisms to Communities. Annual Review of Marine Science, 12, 153–
- 1042 179. https://doi.org/10.1146/annurev-marine-010419-010916
- 1043 Poloczanska, E. S., Burrows, M. T., Brown, C. J., Molinos, J. G., Halpern, B. S., Hoegh-Guldberg,
- 1044 O., ... Sydeman, W. J. (2016). Responses of marine organisms to climate change across oceans.
- 1045 Frontiers in Marine Science, 3(MAY), 1–21. https://doi.org/10.3389/fmars.2016.00062
- 1046 Pörtner, H.-O., D.M. Karl, P.W. Boyd, W.W.L. Cheung, S.E. Lluch-Cota, Y. Nojiri, D.N. Schmidt,
- 1047 and P.O. Zavialov,2014: Ocean systems. In: *Climate Change 2014: Impacts, Adaptation, and*
- 1048 Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth
- 1049 Assessment Report of the Intergovernmental Panel on Climate Change [Field, C.B., V.R. Barros,
- 1050 D.J. Dokken, K.J. Mach, M.D. Mastrandrea, T.E. Bilir, M. Chatterjee, K.L. Ebi, Y.O. Estrada, R.C.
- 1051 Genova, B. Girma, E.S. Kissel, A.N. Levy, S. MacCracken, P.R. Mastrandrea, and L.L. White
- 1052 (eds.)]. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, pp.
- 1053 411-484.
- 1054 Ryther, J. H. (1969). Photosynthesis and Fish Production in the Sea. *Science*, *166*(3901), 72 LP
 1055 76. <u>https://doi.org/10.1126/science.166.3901.72</u>
- Sampaio E., Rosa R. (2020) Climate Change, Multiple Stressors, and Responses of Marine
 Biota. In: Leal Filho W., Azul A.M., Brandli L., Özuyar P.G., Wall T. (eds) Climate Action.

1058 Encyclopedia of the UN Sustainable Development Goals. Springer, Cham.1059 https://doi.org/10.1007/978-3-319-95885-9_90

- 1060 Stock, Charles A., John, J. G., Rykaczewski, R. R., Asch, R. G., Cheung, W. W. L., Dunne, J. P., ...
- 1061 Watson, R. A. (2017). Reconciling fisheries catch and ocean productivity. *Proceedings of the*
- 1062 National Academy of Sciences, 114(8), E1441–E1449.
- 1063 https://doi.org/10.1073/pnas.1610238114
- 1064 Tittensor, D. P., Eddy, T. D., Lotze, H. K., Galbraith, E. D., Cheung, W. W. L., Barange, M., ...
- 1065 Bryndum-Buchholz, A. (2018). A protocol for the intercomparison of marine fishery and
- 1066 ecosystem models: Fish-MIP v1.0. Geoscientific Model Development, 11, 1421–1442.
- 1067 https://doi.org/https://www.geosci-model-dev-discuss.net/gmd-2017-209/
- Trebilco, R., Baum, J. K., Salomon, A. K., & Dulvy, N. K. (2013). Ecosystem ecology: Size-based
 constraints on the pyramids of life. *Trends in Ecology and Evolution*, *28*(7), 423–431.
 https://doi.org/10.1016/j.tree.2013.03.008
- 1071 van Denderen, D., Gislason, H., van den Heuvel, J., & Andersen, K. H. (2020). Global analysis
- 1072 of fish growth rates shows weaker responses to temperature than metabolic predictions.
- 1073 Global Ecology and Biogeography. https://doi.org/10.1111/geb.13189
- 1074 Veytia, D., Corney, S., Meiners, K. M., Kawaguchi, S., Murphy, E. J., & Bestley, S. (2020).
- 1075 Circumpolar projections of Antarctic krill growth potential. Nature Climate Change, 10(June),
- 1076 568–575. https://doi.org/10.1038/s41558-020-0758-4

1077	Waldron, A., Adams, V, Allan, J, Arnell, A., Asner, G, Atkinson, S., & Baccini, A. (2020).
1078	Protecting 30 % of the planet for nature : costs , benefits and economic implications areal
1079	protection in the draft post-2020 Global Biodiversity Framework. Ottawa.
1080	Woodworth-Jefcoats, P. A., Polovina, J. J., Howell, E. A., & Blanchard, J. L. (2015). Two takes
1081	on the ecosystem impacts of climate change and fishing: Comparing a size-based and a
1082	species-based ecosystem model in the central North Pacific. Progress in Oceanography, 138,
1083	533–545. https://doi.org/10.1016/j.pocean.2015.04.004
108/	
1004	
1082	
1086	
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1088	
1089	
1090	Title
1091	Disentangling diverse responses to climate change among global marine ecosystem models
1092	
1093	Highlights
1094	 Experimental study identifying uncertainty sources in FishMIP global model
1095	ensemble
1096	• Warming and lower trophic level (LTL) impacts on model predictions isolated
1097	• Coupling of lower and higher trophic levels a key driver of model warming response
1098	• LTL impacts driven primarily by each model's choice of LTL driver

• Overall climate projections mostly a linear combination of warming and LTL impacts

Declaration of interests

 \square The authors declare that they have no known competing financial interests or personal

- relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be

1107 considered as potential competing interests:

